Stand structure and future development of a managed multi-layered forest in southern Sweden:

Eriksköp - A case study

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# Stand structure and future development of a managed multi-layered forest in southern Sweden: <br> Eriksköp - A case study 

Beståndsstruktur och framtida beståndsutveckling i ett flerskiktat bestånd i södra Sverige:<br>Eriksköp - En fallstudie

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## Abbreviations

BA - Basal area, grundyta
DBH - Stem diameter at breast height (1.3 m), diameter i brösthöjd (BHD)
ha - Hectare, hektar
MAI - Mean annual increment, årligt medeltillväxt
Volume - Total stem volume over bark, beståndsvolym inkl. topp och bark
(if not indicated differently as in the economic discussion)
SE - Standard error, standardfel
NPV - Net present value, nuvärd

## Summary

A heterogeneously structured forest stand with pine overstory and naturally regenerated spruce and oak trees in different size classes was documented. The effect of target diameter cutting on stand structure and growth was analyzed as a case-study. Both, systematic sample plots and forest gaps were used to describe the stand structure after cutting. Target diameter cutting in different treatments reduced the standing volume from ca. 320 to 180 $\mathrm{m}^{3} / \mathrm{ha}$. Forest canopy gaps were created on more than $15 \%$ of the stand area. The seedling number of advanced natural regeneration was low (less than 500-1000 individuals per ha).

Based on the advanced regeneration in gaps, three different scenarios for future ingrowth into the tree layer were defined. The extreme minimum ingrowth scenario assumed about 10 cm annual height growth and rather high mortality reported in literature from other experiments (resulting in one tree annually reaching 5 cm DBH during the next 50 years). Two other scenarios assumed 20 and 30 cm annual height growth. While no mortality was presumed within the latter scenario, moderate mortality rates (reported in literature) were chosen for the intermediate scenario. The maximum scenario postulates ten trees per year and ha reaching 5 cm DBH (equal to ingrowth reported from boreal single-tree selection forests). The moderate scenario assumes four new trees per year and he. Additional scenarios after soil preparation in gaps were used, defined on the base of shelterwood experiments.

To describe the future basal area growth and continued target diameter cutting in the next 50 years, a single-tree growth model was applied, using stand age-independent estimations of the age of single trees. Thereby, a mean annual increment of $0.53-0.64 \mathrm{~m}^{2} / \mathrm{ha}$ was projected, similar to $5.6-6.8 \mathrm{~m}^{3} /$ ha volume. Some errors to estimate the standing volume in multi-layered stands were detected during the simulation process. Compared to an evenaged spruce stand planted on the same site, the expected growth of the study stand during the next 50 years was one third lower. In average, about $120 \mathrm{~m}^{3} / \mathrm{ha}$ standing volume was removed in 20-25 years-cutting cycles. To continue without longer harvest intervals after the 50 years-simulation period, soil preparation seems necessary to achieve a sustainable number of small trees. Beside timber production, profitability was also lower by selective cutting. But, the important advantage of target diameter cutting can be more equally distributed income over time, with investments costs that can be covered by profit from timber harvest at the same time. A regular income of 17000-28000 SEK per ha every 20-25 years seems possible from today's perspective.

An additional treatment with alternative target diameters to promote particular tree species did not affect the amount of removals and the length of cutting intervals substantially. But simulations with 5 cm reduced target diameters caused very heavy removals and 35-40 years to reach $300 \mathrm{~m}^{3} /$ ha standing volume again. The study includes discussions of tree species composition and development as well as a sensitivity analysis of the applied growth model.

## Samfattning

Studien beskriver ett skogsbestånd med heterogen struktur (tall i det dominerande skiktet,, gran i alla skikt, och ek) och hur måldiameterhuggning påverkar beståndsstruktur och tillväxt. Inventering av beståndsstruktur efter huggning utfördes dels inom systematiskt utlagda provytor och dels som luckinventering. De testade behandlingar med måldiameterhuggning minskade beståndsvolymen från ca. 320 till $180 \mathrm{~m}^{3} \mathrm{sk} / \mathrm{ha}$ i medeltal. Luckorna i krontaket utgjorde $15 \%$ av beståndets totala yta. Antalet småträd i självföryngringen var lågt (färre än 500-1000 individer per ha med BHD mindre än 5 cm ).

Scenarier med tre olika nivåer förinväxning av självföryngring i luckor definierades: 1) låg nivå med ca. 10 cm årlig höjdtillväxt och hög mortalitet som resulterade i en inväxning av ett träd per år och ha i trädklass över 5 cm BHD, 2) medelnivåmed ca. 20 cm årlig höjdtillväxt och en inväxning av fyra träd per år och ha, 3) hög nivå utan mortalitet och en årlig höjdtillväxt på ca. 30 cm som resulterade i 10 nya träd per år och ha. Scenariot med hög nivå motsvarar observerad inväxning i boreal blädningsskog. Baserat på erfarenheter från flera skärmställningsförsök gjordes även olika antaganden angående markberedningens inverkan på inväxningen.

Tillväxten under 50 år med fortsatt måldiameterhuggning simulerades. . Under den simulerade perioden var den årliga grundytetillväxten $0,53-0,64 \mathrm{~m}^{2} / \mathrm{ha}$ och volymtillväxten $5,6-6,8 \mathrm{~m}^{3} \mathrm{sk} / \mathrm{ha}$ i medeltal för den. Beräknad produktion i det studerade beståndet var en tredje del lägre än förväntad produktion i ett planterat likåldrigt granbestånd på samma ståndort. I medeltal var det möjligt att avverka omkring $120 \mathrm{~m}^{3} \mathrm{sk} / \mathrm{ha}$ varje $20-25$ år. För att på längre sikt kunna bibehålla ett avverkningsintervall på 20-25 år och samtidigt trygga föryngringen, rekommenderas markberedning för att initiera föryngring som kan ersätter mogna träd på lång sikt. Även intäkterna var lägre vid måldiameterhuggning jämfört med trakthyggesbruk. Vid måldiameterhuggning var emellertid intäkterna mer jämt fördelade över tiden och det fodrades heller inga dyra inverteringar i föryngring. Netto intäkterna var ungefär 17000-28000 kronor per ha varje 20-25 år (förutsatt bibehållen prisbild).

En kompletterande behandling med alternativa måldiametrar för att gynnar specifika trädslag visade inga stora effekter på uttag och huggningsintervall. Simuleringar med 5 cm lägre måldiameter resulterade i ett stort uttag och det krävdes ett avverkningsintervall på 35-40 år för att åter nå upp till beståndsvolum på $300 \mathrm{~m}^{3} \mathrm{sk} / \mathrm{ha}$ innan nästa avverkning. I studien diskuteras även hur den framtida trädslagsfördelning i beståndet kommer att utvecklas. Dessutom diskuteras eventuella felkällor vid modelleringen av skogens utveckling. Skattningen av tillväxten i flerskiktad skog innebär i flera fall en extrapolering av tillämpade modeller med ökad osäkerhet som följd.

## Introduction

## Continuous cover forestry and other silvicultural systems in Sweden and northern Europe

The currently dominant method to harvest trees in mature forest is clear-cutting. In Sweden, clear-cutting and the seed-tree method are applied in ca. $95 \%$ for harvesting mature stands (Anon 2002). Other methods are shelterwood cutting (Holgen, 1996) and single-tree selection cutting (Lundqvist, 1989). However, the latter is accounted as thinning method in Sweden, not as harvest method (Anon 2002)!
Historically, dimension cutting was used in many parts of northern Sweden in the beginning of the $20^{\text {th }}$ century. After the removal of the largest trees in pristine forest areas, this type of cutting was repeated. Finally, target diameters were too small and caused forest depletion (Holmgren, 1959). The growth of the remaining trees and regeneration processes developed much slower than expected.
Crucial for sustainable target diameter cutting is the initial stand structure and the target diameter limit which determines harvest volume and cutting cycles. If the diameter threshold is too low, forest can be depleted. If the number of small trees is too low (and large trees are very regularly distributed in the stand), the cutting could open up forest canopy similar to shelterwood conditions (with more evenly distributed regeneration in the stand). Lundqvist (1989) provided a good reference to boreal single-tree selection stands under quasi-equilibrial conditions. A reference to natural stand structures in fully stocked boreal forests can be found in Shorohova et al. (2009).

Continuous cover forestry is currently under strong debate in Sweden (i.e. NSF, 2010). For some decades, silvicultural guidance for forest owners considering selective final felling as potential option was limited compared to treatments like clear-cutting, planting and thinning. More recently, the ecological value and recreational aspects after clear-cut were discussed more controversially (also due to the decreasing area of forest with continuous tree cover, changes of the perception in the society, and new insights into natural processes of the biocenosis; see i.e. Bengtsson \& Rosell, 2010). Today, a certain potential for continuous cover forestry is recognized, especially close to urban areas or to protect species depending on continuous forest cover.
General overviews regarding continuous cover forestry in Sweden were given by Axelsson (2008), highlighting history and policy aspects, and Erefur (2010) with focus on regeneration issues. Saksa \& Valkonen (2011), Laiho et al. (2011) and Kuulavainen et al. (2012) gave an overview on Finnish research about continuous cover forestry. Kuulavainen et al. (2012) and Pommerening \& Murphy (2004) discussed the terminology to describe forms of continuous cover forestry. However, the term is to some extent still differently used in Finland, UK, or Sweden for instance. Another interesting approach in practice to continuous cover forestry is described by the Forestry Commission in UK (FC, 2004).

## Study goals and hypotheses

This case study aims at exploring diameter cutting as an option to manage heterogeneously structured forests in southern Sweden. The focus was set on the effects of target diameter cutting on stand structure and stand development per se. Additional silvicultural measures, like creating larger gaps or enrichment planting were not considered. The future stand
development with continued target diameter cuttings was simulated with the empirical growth models that exist for Sweden. However, caution is warranted when interpreting the results, as the single-tree model was applied outside of its validation range (see discussion). The study stand represents a common forest type in Sweden, with pine-dominated overstory and shade-tolerant tree species below (Figure 1). The mixture of pine and spruce for instance, covers about 300.000 ha in Götaland, comprising a high proportion of mature stands (Drössler, 2010). Such a forest type can be considered as rather suitable for alternative silvicultural methods, because the risk of wind damage is lower than in a pure, single-layered spruce forest (Valinger \& Fridman, 2011). The selected stand is more heterogeneously structured than the majority of Swedish forests.

The goal of the study was to estimate future stand growth and timber harvest over a 50 year period, and to test and discuss the effect of different ingrowth scenarios. Furthermore, the initial stand structure was documented for comparisons with other stands and for future remeasurements to validate the growth functions commonly used in forestry in Sweden. The hypotheses of this study are: (1) Target diameter cutting can create an irregular pattern of gaps, different to uniform shelterwood. (2) For the next 50 years, forest production in the study stand, managed by target diameter cutting, is predicted to be similar to an even-aged spruce stand planted on the same site. (3) The projected growth will decrease significantly, if the target diameters applied are reduced by 5 cm . (4) Ingrowth will have a small impact on projected increment the first 50 years. (5) The tree species composition will change towards more shade tolerant species during the simulated development. (6) Replanting spruce after clearcut would result in a similar net present value compared to target diameter cutting (based on 2\% interest rate).

## Experiences from continuous cover forestry in Central Europe

In other parts of Europe, the largest proportion of uneven-aged managed forests can be found in Switzerland and Slovenia (ca. 10\%), where single stands have been described and inventory methods were developed more than 100 years ago (Biolley, 1887). In France, Germany, Czech Republic, Austria and Slovakia, less than 2\% of the forests are uneven-aged (Schütz, 1994). The typical uneven-aged forest are montane fir-spruce-beech stands, but on special places a management tradition can be found in subalpine pure spruce stands and submontane pure beech stands too (Indermühle, 1978; Erteld, Gerold, Mund, Schulze \& Weller, 2005). Since 60 years, extensive research in fir-spruce-beech forests on stand structure and regeneration led to a better understanding of stand developmental pattern (Leibundgut, 1945; Mitscherlich, 1952; Kern, 1966).
Research in other heterogeneously structured forests with more light-demanding tree species is very limited. Schütz (2001) concludes that these stands often undergo a transition stage rather than approaching an equilibrium (regarding balanced inverse j-shaped diameter distributions). Under long shelter periods, the successful regeneration of light-demanding tree species is very difficult (Lüpke \& Hauskeller-Bullerjahn, 1999).

Due to an increasing public interest in nature in the 1970's, the dominance of public forest owners in Germany, high population densities, and low contribution of the forestry sector to gross national products, the paradigm prior to timber production changed to a strong emphasis of other forest functions. After a general ban of the clearfelling system in Germany
(which was applied in spruce and pine stands), shelterwood and target diameter cutting became the most common methods. In beech stands, shelterwood was already the dominant harvest method since 200 years (Hartig, 1791). Oak and many pine stands were also harvested by shelterwood cutting, but in short regeneration periods (ca. 10 years).

Today, most stands are managed by target diameter cutting which is applied in single-tree selection and shelterwood systems (Schütz, 1997; Spellmann, 1997). While target diameter cutting aims for equilibrium conditions in uneven-aged forests (Lundqvist, 1989; Schütz, 1997), it can be used to transform even-aged forest into more heterogeneously structured stands (Richter, 1995). In most cases, this means an extended regeneration period (ca. 40 years in even-aged beech stands) resulting in more heterogeneously structured natural regeneration (but still even-aged, e.g. Petritan et al., 2007). In the very long run, a transformation from even- to uneven-aged forest can be expected (Sterba \& Zingg, 2001; Schütz, 2001, 2002; O’Hara, 2001; Kerr, Morgan, Blyth \& Stokes, 2010). But, most authors pointed out that transformation of very homogeneously structured stands is difficult. In even-aged spruce forests for instance, the target diameter cutting caused considerable storm damage (Redde, 2002). Therefore, strip cutting is often recommended on labile sites.

## Material \& methods

## Stand description and history

The study stand is located in southwest Sweden ( $56^{\circ} 42^{\prime} 02^{\prime \prime} \mathrm{N}, 13^{\circ} 7^{\prime} 56^{\prime \prime} \mathrm{E}, 115-140 \mathrm{~m}$ a.s.l.) at the Tönnersjöheden experimental forest, within the transition zone between nemoral and boreal forests (Ahti, Hämet-Ahti \& Jalas, 1968; Bohn \& Weber, 2000). Mean annual precipitation in the area is 1050 mm , mean annual air temperature is $6.7^{\circ} \mathrm{C}$ and the vegetation period lasts for 215 days (when mean daily air temp. $>5^{\circ} \mathrm{C}$ ). The soil type is a podzol, developed over sandy-fine sandy moraine. The ground floor vegetation is mainly characterized by Vaccinium myrtillus and thin-leaved grasses on this mesic site. Regarding the site index according to Hägglund (1973), 32 m top height was estimated for a spruce stand at this site at the age of 100 years (based on observation in adjacent spruce plantations). The stand area was 19 ha (Figure 2).


Figure 1. Section of the stand Eriksköp.
Del av beståndet i Eriksköp.
Before stand establishment in 1912, the area was a Calluna heath land with some single trees (incl. Picea abies L. Karst) sparsely scattered across the landscape. The stand was established by seeding of pine (Pinus silvestris L.), supplemented by an abundant natural regeneration of birch (Betula pendula Roth.). The first cutting took place in 1947, releasing single pine trees in combination with a thinning from below on residual areas. In 1953 and 1958, thinnings from below were carried out. Thinnings were carried out in 1974 and 1991. During this development, additional trees like Picea abies, Quercus petraea Liebl., Sorbus aucuparia L. and Fagus sylvatica L. entered the stand. Single individuals of Salix caprea L., Alnus glutinosa L., Populus tremula L., Juniperus communis L. and Malus silvestris L. also established.


Figure 2. Montage of aerial photographs of the stand in summer 2008 immediately after the first target diameter cutting.
Flygfoto över beståndet tagna under sommar 2008, strax efter första måldiameterhuggningen.

## Stand treatments and target diameters

The stand was divided into three blocks, with four different experimental treatments repeated in each block (Figure 3):

1. Treatment $\mathbf{T}$ was defined by cutting according to target diameters. A tree was removed when the DBH was equal or larger than indicated in Table 1. These target diameters were chosen according to economic criteria and were similar to final tree diameters often used as production goal in even-aged stands.
2. Treatment TS was a target diameter cutting with soil preparation and tending of small trees as additional silvicultural measures. The soil scarification was applied in gaps. The cleaning removed some saplings with abnormal stem form, damages or spike-knots.
3. Treatment TN was a cutting with larger target diameters for broadleaves and pine. The target was to enhanced natural values and to promote other species on the expense of spruce (see Table 1).
4. The control treatment $\mathbf{C}$ was not managed.

In each of the three blocks, the single treatments were carried out on 1 ha. This area was sampled by four systematically distributed 10 m radius-plots (Figure 3). Each treatment was sampled by twelve plots on total stand area. Total plot number of the experiment was 48.

Some trees were left in the stand to increase the natural value, but none of these retention trees was selected on the plots. While 20 retention trees per ha were selected on the stand section managed according to treatment TN , ten trees per ha were chosen in other treatments.

The cutting was carried out in winter 2008/09. All removals were planned in order to be feasible by a harvester. Another target was to keep the harvesting costs low. Hence, no fixed skid road system was used for all treatments, given the responsibility to the harvester driver. However, mature trees were marked before harvest to maintain the experiment. In autumn 2010, soil scarification in treatment TS was carried out. No cleaning of dense regeneration was necessary.

Table 1. Target diameter (in cm ) of tree species according to treatment ( $\mathrm{T}, \mathrm{TS}$ and TN) and quality class 1 and 2 (class 2 describes trees with branches thicker than 6 cm , spike-knots or forks). Måldiameter (cm) för olika trädslag inom tre olika behandlingar och kvalitetsklass 1 och 2.

|  | T/TS |  | TN |
| :---: | :--- | ---: | :--- |
|  | 1 | 2 |  |
| Pine | 40 | 30 | 40 |
| Spruce | 36 | 26 | 26 |
| Birch | 30 | 20 | 30 |
| Oak | 60 | 30 | 60 |
| Beech | 50 | 30 | 50 |



Figure 3. Stand map with the distribution of treatments and sample plots. $\mathrm{C}=$ control, $\mathrm{T}=$ target diameter cutting, TS = target diameter cutting with silvicultural measures, TN = target diameter cutting for natural values. Number of the three experimental blocks in brackets. Beståndskarta med fördelning av behandlingar och provytor. $C=$ kontroll, $T=$ måldiameterhuggning, $T S=$ måldiameterhuggning med skötselåtärder i kvarvarande bestånd, $T N=$ måldiameterhuggning med naturhänsyn.

## Sampling methods

Trees On the systematically distributed sample plots with 10 m radius, trees with DBH $\geq 4.5$ cm were cross-callipered and the tree species was recorded. For every second tree, tree height was measured.
Smaller trees with DBH < 4.5 cm and height $\geq 10 \mathrm{~cm}$ (hereafter referred to regeneration) were counted on 192 circular $5 \mathrm{~m}^{2}$-subplots, located on the 10 m radius-plots ( 5 m distant from its centre in the four cardinal directions). For each individual, tree species and height class were determined. Height classes were $10-19 \mathrm{~cm}, 20-49 \mathrm{~cm}$, and 50 cm -height classes up to 299 cm . For trees with a height $\geq 3 \mathrm{~m}$ and DBH $<4.5 \mathrm{~cm}$, height and DBH were recorded. For the largest individual of each species on a subplot, annual height shoot lengths of the last three years and damages by browsing or harvest were recorded. Dead shoots or shoots with unknown scars were neglected. (Rhamnus frangula L. was ignored in the regeneration survey.)

Forest gaps The gap survey was a complete stand inventory of forest canopy openings. Openings were defined as areas where the forest floor was not covered by the crown of
trees with DBH $\geq 4.5 \mathrm{~cm}$. Minimum gap size was defined by 5 m gap length (app. $20 \mathrm{~m}^{2}$ assuming a circular gap).
The length and width of a gap was measured from the edge of adjacent tree crowns at the four points indicated in Figure 4 using a crown projection prism. Gap area was calculated using gap length and width to calculate an ellipse according to Runkle (1992). If the shape of a gap was not similar to an ellipse, the gap area was measured by polygons. If single live trees with DBH $\geq 4.5 \mathrm{~cm}$ occurred in a gap, their crown cover was measured and subtracted from the gap area. GPS coordinates of the gap centre were also recorded.
In the gaps, stumps of harvested were counted if the diameter at the height of the cut was larger than 25 cm (assuming that the tree had reached the upper stand layer). Remnants of other dead canopy trees were counted if DBH was larger than 20 cm or the stump diameter on the ground was estimated to be larger than 25 cm . Four cases of mortality were identified: 1) Fresh stump, recently harvested by target diameter cutting. 2) Old stump (by former cuttings or breakage). 3) Recent wind-throw after target diameter cutting. 4) Old wind-throw before target diameter cutting.

Regeneration in gaps Two inventory concepts were applied to study forest regeneration: The systematic survey described above, and a complete inventory in gaps. Reasons for the second inventory was the low number of seedlings found on systematic plots and the hypothetical importance of gaps for future stand development. The minimum height to record seedlings in gaps was 20 cm . Height classes above 20 cm and tree measurements were identical with the survey on permanent plots. Only the highest individual of each species per gap was selected to record height shoots of the last three years and damages.


Figure 4. Fitting length and width to the canopy opening.
Mätning av luckornas längd och bredd i krontaket.

Light measurements To describe approximate light conditions in the stand and in gaps, the crown coverage was measured by a forest canopy densiometer (Lemmon, 1956). The mirror of this tool reflects a cone with an opening angle of 24 degree and allows a distinction of light and dark spots in a stand.
The measurements were taken in August, 1.5 m above the centre of the permanent regeneration sub-plots and in the centre of selected gaps, representing the range of gap sizes in the stand. The crown coverage was calibrated by the diffuse site factor (Wagner, 1994) which represented the whole range of light conditions in the stand. The diffuse site factor was determined by fish-eye photographs and analyzed according to Wagner (1994).

Ground vegetation was classified on the sample plots according to Hägglund \& Lundmark (1994).

## Analytical methods

For the structural analysis, the stand was divided into three height strata according to Leibundgut (1952). Stand height was determined as the regression height of the tree with a DBH equal to the arithmetic mean of the diameter distribution plus three times the standard deviation (Näslund, 1936), regardless tree species. In addition, logarithmic height curves were derived for pine, spruce, birch, oak and beech.
Söderberg's (1992) functions of height, form height and bark thickness were used to calculate initial stand volume. However, future volume was estimated by BA with volume functions according to Ekö (1985).

## Simulation of future stand development

The collected tree data on the $10-\mathrm{m}$ radius plots before cutting was used for simulations. Hence, all forecasted stand developments are based on the same initial stand conditions. Besides the four experimental treatments described above, two additional simulated treatments with 5 cm reduced/increased target diameters were also tested ( $\mathrm{T}+5, \mathrm{~T}-5$ ).

The growth of trees larger than 5 cm DBH was simulated by a set of empirical models:

1. A basal area growth function for individual trees according to Elfving (2004) which is independent from stand age.
2. Functions to estimate the age of single trees at breast height according to Elfving (2003). The functions can adopt stand age optionally as independent variable. But, the results presented refer to the stand age-independent estimation of single tree ages. (However, the stand age-dependent estimation was used for validation. In that case, stand age was defined as the age of the tree with mean basal area diameter. The age of the tree with mean basal area diameter was estimated from the number of internodes counted in the field at trees with that size.)
3. Tree mortality was estimated according to Fridman \& Ståhl (2001).

Fridman \& Ståhl (2001), Elfving (2003), and Elfving (2004) used the Swedish National Forest Inventory (Ranneby et al., 1987) to parameterize their functions.

Elfving's $(2003,2004)$ models were based on permanent sample plots of the Swedish National Forest inventory (RIS, 2008). The age function of single trees in uneven-aged stands was derived from sample plots classified as uneven-aged forest (Elfving, 2003). "Unevenaged" plots were defined by minimum two age classes (with 20 years intervals) and minimum $20 \%$ standing volume in each class (RIS, 2008).
For the purpose of validation of Elfving's single tree models, a more robust stand growth model for even-aged stands by Ekö (1985) was applied. The age at breast height of pine and birch trees was set 90 years. For spruce, 60 years was assumed, and for oak and beech 40 years.
For comparisons with the development of a planted spruce stand on the same site, the forest simulator DT was applied, which is based on single tree growth functions by Elfving (2004), but calibrated with Ekö's (1985) stand growth functions (cf. Nilsson \& Fahlvik, 2006). For the simulation with DT, 1800 established spruce trees per ha with 7 m mean height after 20 years were assumed.

For this study, a simulator was developed which integrates the models by Elfving (2003, 2004) and Fridman \& Ståhl (2001), as well as ingrowth scenarios described below to forecast and evaluate the stand development for the next 50 years. Input variables refer to site conditions (latitude, altitude, distance to coast, site indices for spruce and pine, vegetation characters, soil moisture), stand characteristics (age, indicator variable for uneven-aged, number of years since thinning) and individual trees features (species, DBH and height). Data from measurements in spring 2007 was used as initial values for the simulations. The calculation procedure was repeated 25 times per plot, and summarized as average value. Repeated simulations were made due to the inclusion of stochastic elements in the mortality functions. To estimate stand volume from BA, stand volume functions according to Ekö (1985) were applied. These functions are less accurate but also less biased, especially when it comes to repeated simulations. Therefore, volume predicted by stand volume functions are considered more robust within simulations of this study.
For the simulated target diameter cuttings, single-tree removals were specified according to their diameter derived from single-tree BA projections. When total standing volume on all plots was larger than $300 \mathrm{~m}^{3} / \mathrm{ha}$ in average, a target diameter cutting was simulated. If the remaining standing volume would be less than $100 \mathrm{~m}^{3} / \mathrm{ha}$, the cutting would be delayed too (simulations with $150 \mathrm{~m}^{3} /$ ha minimum volume after cutting were tested additionally).

Ingrowth scenarios In addition to the empirical tree growth models, hypothetical scenarios of future ingrowth into the tree layer ( $\geq 5 \mathrm{~cm}$ DBH) were formulated. Three different scenarios were defined to cover a wide range of possible, reasonable ingrowth.
The minimum scenario was derived from the current regeneration, before the first growing season after first cutting. The average of the last 3 -year height shots was calculated to estimate future minimum height growth per species and height class. In addition, mortality rates described in Table 2 were assumed, based on lowest seedling survival of spruce reported in literature (Nilson \& Lundquist, 2001; Holgén \& Hånell, 2000; Wikberg, 2004).

Table 2. Annual mortality of small trees assumed in the minimum ingrowth scenario. Antagande om årlig mortalitet hos småträd enligt scenariot med låg inväxningstakt scenariot.

|  | $<1 \mathrm{~m}$ | $1-2 \mathrm{~m}$ | $>2 \mathrm{~m}$ |
| :---: | :---: | :---: | :---: |
| Spruce | $10 \%$ | $5 \%$ | $1 \%$ |
| Birch | $20 \%$ | $15 \%$ | $5 \%$ |
| Oak | $10 \%$ | $5 \%$ | $1 \%$ |
| Beech | $10 \%$ | $5 \%$ | $1 \%$ |

The maximum ingrowth scenario assumed no mortality and height growth rates of 30 cm per year. Such rates represent spruce sapling growth on similar site under open land-conditions (Hägglund, 1981). The third, moderate scenario was based on average height growth and average mortality rates used wthin the other two scenarios.

Table 3. Summary of assumptions made for the ingrowth scenarios.
Sammanfattning av antaganden för olika inväxningsscenarier.

| Scenario | without scarification |  |  | after scarification |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | regeneration reference | annual height growth | annual ingrowth with 5 cm dbh | regeneration reference | annual height growth | final ingrowth with 5 cm dbh |
| Minimum | advanced regeneration in gaps | $\begin{array}{\|c\|} \hline \text { last } \\ 3 \text {-years } \\ \text { of the } \\ \text { current } \\ \text { regen. } \\ \text { in gaps } \\ \text { (tab. 13) } \\ \hline \end{array}$ | 1 tree/ha | scarification under shelterwood (2000 spruce trees/ha or 1000 birch and 1000 pine trees) | last <br> 3-years <br> of the <br> current <br> regen. <br> in gaps | 1000 trees/ha in gaps (equal to 2\% annual mortality) |
| Medium | advanced regeneration in gaps | 20 cm | 4 trees/ha | scarification under shelterwood (5000 spruce trees/ha or 2500 birch and 2500 pine trees) | 20 cm | 2000 trees/ha in gaps (stem number reduced by cleaning at the next harvest) |
| Maximum | advanced regeneration in gaps | 30 cm | 10 trees/ha | scarification under shelterwood (20000 spruce trees/ha or 10000 birch and 10000 pine trees) | 30 cm | 2000 trees/ha in gaps (stem number reduced by cleaning at the next harvest) |

Box 1. Technical ingrowth implementation:
New trees were implemented on sample plots with the lowest BA. The proportion of selected plots was equal to the gap proportion of the stand found after cutting. On these plots intending to represent gaps, new trees with 5 cm DBH were adopted in 5 year-steps.
Additionally, different scenarios were constructed for the treatment with soil preparation. According to literature, 1000-50000 seedlings per ha can be expected five years after scarification under shelterwood (Nilsson, Gemmel, Johansson, Karlsson \& Welander, 2002; Nilsson, Örlander \& Karlsson, 2006). But, no description of the development of young trees older than ten years under shelter was found. Based on 2$5 \%$ annual mortality found for younger regeneration after scarification under shelter (Holgén \& Hånell, 2000; Nilsson et al., 2002; Nilsson et al., 2006), 0-4\% mortality for 0.5-2 m high saplings in uneven-aged forest (Nilson \& Lundqvist, 2001), and very low mortality rates in young stands (Petterson, 1992), generally $2 \%$ mortality was assumed for individuals established by scarification until reaching 5 cm DBH. The minimum scenario was derived from 2000 seedlings/ha established in scarified gaps after five years, and height growth rates according to the current advanced regeneration. In that case, 1000 trees/ha with 5 cm DBH were assumed in gaps 40 years after scarification. In the medium and maximum scenario, seedlings with 20 and 30 cm annual height growth were assumed to grow into the tree layer after 25 respective 20 years. Then, a fixed number of 2000 trees/ha in gaps was expected in both scenarios as an effect of cleaning.
On plots with extremely low $B A\left(<10 \mathrm{~m}^{2} / \mathrm{ha}\right.$ ), pine and birch was assumed as ingrowth instead of spruce.

## Results

## Documentation of stand structure

## Stand structure before cutting

Before harvest, the standing volume across all treatments was $322 \mathrm{~m}^{3} / \mathrm{ha}\left(\mathrm{SE}=12.1 \mathrm{~m}^{3}\right)$, plus $8.6 \mathrm{~m}^{3}$ deadwood. The BA of living trees amounted to $35.2 \mathrm{~m}^{2} / \mathrm{ha}\left(\mathrm{SE}=1.6 \mathrm{~m}^{2}\right.$ ) with $44 \%$ pine ( $\mathrm{SE}=3.9 \%$ ), $39 \%$ spruce ( $\mathrm{SE}=3.5 \%$ ) and $16 \%$ broadleaves ( $\mathrm{SE}=1.7 \%$ ). On the blocks, the standing volume was 270,353 and $368 \mathrm{~m}^{3} / \mathrm{ha}$ (Table 4). It ranged from 104 to 510 $\mathrm{m}^{3} / \mathrm{ha}$ on plots. Basal area was $31.8,38.1$ and $38.5 \mathrm{~m}^{2} / \mathrm{ha}$ on blocks, and ranged from 16.7 to $53.2 \mathrm{~m}^{2} / \mathrm{ha}$ on plots.
In terms of total tree number, the proportion of pine was 16\% (SE = 1.6\%) and spruce 49\% ( $\mathrm{SE}=4.0 \%$ ). The relative number of oak trees was $27 \%$, which differed considerably from its $9 \%$ basal area proportion (Table 4). Total tree number per ha was 999 (plus 32 dead trees).

Table 4. Tree species composition before cutting in the stand and the blocks Trädslagsfördelning innan huggning inom hela beståndet och inom respektive block.

| Species | Total stand |  |  | Block 1 |  |  | Block 2 |  |  | Block 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | N/ha | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{gathered}$ | Vol [m $\mathrm{m}^{3} / \mathrm{ha}$ ] | N/ha | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{gathered}$ | Vol [ $\mathrm{m}^{3} / \mathrm{ha}$ ] | N/ha | BA [ $\mathrm{m}^{2} / \mathrm{ha}$ ] | Vol [ $\mathrm{m}^{3} / \mathrm{ha}$ ] | N/ha | BA [ $\mathrm{m}^{2} / \mathrm{ha}$ ] | Vol [ $\mathrm{m}^{3} / \mathrm{ha}$ ] |
| Pine | 164 | 15.4 | 162 | 137 | 12.4 | 126 | 217 | 21.5 | 227 | 151 | 13.6 | 146 |
| Spruce | 489 | 13.8 | 118 | 392 | 9.8 | 78 | 418 | 11.2 | 90 | 716 | 21.6 | 195 |
| Birch | 51 | 2.6 | 22 | 78 | 4.2 | 35 | 44 | 1.8 | 14 | 40 | 2.0 | 17 |
| Beech | 21 | 0.4 | 2 | 26 | 0.5 | 3 | 28 | 0.6 | 4 | 10 | 0.1 | 1 |
| Oak | 271 | 3.0 | 18 | 458 | 4.8 | 29 | 257 | 3 | 18 | 119 | 1.2 | 8 |
| Others | 3 |  |  |  |  |  | 2 |  |  | 4 |  |  |
| Total | 999 | 35.2 | 322 | 1090 | 31.8 | 270 | 965 | 38.1 | 353 | 1040 | 38.5 | 368 |

The initial diameter distribution of the total stand is shown in Figure 5. Large tree classes (DBH > 30 cm ) were dominated by pine (70\%). In lower size classes, spruce (55\%) and oak (33\%) were most frequent. In the DBH class from 5 to 15 cm , the oak proportion increased to 45\%. A few very large spruce trees were also found.
An exponential function used to correlate tree size and total tree number per DBH class resulted in $R^{2}=0.81$, while the coefficient of determination was 0.70 for a linear function. The average quotient of tree numbers from one DBH class to the next 1 cm class was 1.14. The quotient ("Q-factor") between 4 cm classes was 1.33 .
The general pattern of size class distributions and tree species composition were similar in all blocks. But, the oak proportion was higher in Block 1. The highest proportion of spruce was found in Block 3 (Figure 6).


Figure 5. Diameter and tree species distribution in the stand before cutting.
BHD- och trädslagsfördelning inom hela beståndet innan huggning (asp, ek, bok, björk, gran, tall).


Figure 6. Diameter and species distributions in each block before cutting.
BHD- och trädslagsfördelning inom de tre blocken.


Figure 7. Tree heights (all species, 886 measured heights and 636 estimations based on logarithmic heights curves of tree species). Trädhöjder och höjdkurvor för tall (b/å), gran(röd) och ek.

The stand height was 27 m . (The maximum tree height in Figure 7 displays a spruce likely established before pine was seeded.) Height curves for all single tree species are given in Appendix A.
$31 \%$ of trees were found in the upper stand layer, while $50 \%$ resp. $19 \%$ occurred in the medium and lower layer (Table 5). The proportion of pine trees in the upper layer was $57 \%$, while spruce comprised $39 \%$.

Table 5. Percentage of stem number, basal area and standing volume in the three stand layers before harvest (stand- and blockwise).
Fördelningen av stamantal, grundyta och beståndsvolym över tre olika höjdklasser innan den första
huggningen.

|  | Total stand |  |  | Block 1 |  |  | Block 2 |  |  | Block 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| stand layer | N | BA | Vol | N | BA | Vol | N | BA | Vol | N | BA | Vol |
| $<8.9 \mathrm{~m}$ | 19\% | 2\% | 1\% | 21\% | 4\% | 2\% | 18\% | 2\% | 1\% | 16\% | 2\% | 1\% |
| 8.9-17.8 m | 50\% | 25\% | 18\% | 56\% | 32\% | 24\% | 50\% | 24\% | 17\% | 48\% | 23\% | 17\% |
| $>17.8$ m | 31\% | 73\% | 81\% | 23\% | 64\% | 74\% | 32\% | 74\% | 82\% | 36\% | 75\% | 82\% |
| total stand | $\begin{array}{\|c\|} \hline 999 \\ \text { trees/ha } \end{array}$ | $\begin{gathered} 35.2 \\ \mathrm{~m}^{2} / \mathrm{ha} \end{gathered}$ | $\begin{gathered} 322 \\ \mathrm{~m}^{3} / \mathrm{ha} \end{gathered}$ | $\begin{array}{\|c\|} \hline 1090 \\ \text { trees/ha } \end{array}$ | $\begin{gathered} 31.8 \\ \mathrm{~m}^{2} / \mathrm{ha} \end{gathered}$ | $\begin{gathered} 270 \\ \mathrm{~m}^{3} / \mathrm{ha} \end{gathered}$ | $\begin{gathered} 965 \\ \text { trees } / \mathrm{ha} \end{gathered}$ | $\begin{gathered} 38.1 \\ \mathrm{~m}^{2} / \mathrm{ha} \end{gathered}$ | $\begin{gathered} 353 \\ \mathrm{~m}^{3} / \mathrm{ha} \end{gathered}$ | $\begin{array}{\|c\|} \hline 1040 \\ \text { trees/ha } \end{array}$ | $\begin{gathered} 38.5 \\ \mathrm{~m}^{2} / \mathrm{ha} \end{gathered}$ | $\begin{gathered} 368 \\ \mathrm{~m}^{3} / \mathrm{ha} \end{gathered}$ |

Regarding the social hierarchy of trees, a relatively even distribution of dominant and suppressed trees was indicated in Table $6.64 \%$ of dominant trees were pine, but $80 \%$ of codominant trees were spruce. $42 \%$ of suppressed trees were oak, $52 \%$ were spruce.

Table 6. Percentage of trees in social hierarchy classes according to Schotte (1912).
Trädens fördelning över olika sociala klasser definierade enligt Schotte (1912).

| Individuals | Social class |
| :---: | :---: |
| $34.2 \%$ | Suppressed |
| $27.4 \%$ | Sub-dominant |
| $13.9 \%$ | Co-dominant |
| $24.5 \%$ | Dominant |

## Stand structure after cutting

Summarizing the three management treatments, the standing volume was reduced from approximately $320 \mathrm{~m}^{3}$ down to $180 \mathrm{~m}^{3}$ per ha (Table 7). In the two treatments T and TS the standing volume decreased from 314 and $357 \mathrm{~m}^{3}$ down to 181 and $196 \mathrm{~m}^{3}$ per ha. In treatment TN, the volume decreased from 301 to $173 \mathrm{~m}^{3} / \mathrm{ha}$. Standard errors on plot level were 21-26 $\mathrm{m}^{3}$ for removals and $16-26 \mathrm{~m}^{3}$ for the remaining volume (Table 7).

Table 7. Standing volume and standard error on plot level [ $\mathrm{m}^{3} / \mathrm{ha}$ ] before and after cutting on differently treated plots. Beståndsvolym och standardfel innan och efter den först huggningen, baserat på 12 provytor per behandling.

|  | T | TS | TN | C |
| :--- | :---: | :---: | :---: | :---: |
| Before cut | $314(23)$ | $357(25)$ | $301(17)$ | $314(26)$ |
| Removal | $133(22)$ | $161(26)$ | $128(21)$ | 0 |
| After cut | $181(18)$ | $196(16)$ | $173(21)$ | $314(26)$ |
| Plot number | 12 | 12 | 12 | 12 |

The diameter distributions in Figure 8 were relatively similar between different treatments. However, treatment TN was characterized by a smaller proportion of spruce trees for instance. Treatment T had the lowest proportion of pine.


Figure 8. Diameter distributions after cutting on differently treated plots ( 12 plots per treatment). BHD-fördelning efter fyra olika behandlingar (12 provytor per behandling).

Figure 8. Diameter distributions after cutting on differently treated plots (12 plots per treatment). BHD-fördelning efter fyra olika behandlingar (12 provytor per behandling).

The stand height decreased to 24 m ( $T$ and TS treatment) and 25 m (TN) after cutting. The percentage of trees in the medium layer increased, while the lower layer remained rather constant (Table 5 and 8). In the upper layer (> 16 m ), about 45\% of trees were pine and 49\% spruce.

Table 8. Percentage of stem number, basal area and standing volume in three stand layers after first harvest. Fördelningen av stamantal, grundyta och beståndsvolym över tre olika höjdklasser efter den första huggningen.

|  |  | T |  |  | TS |  |  | TN |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | N | BA | Vol | N | BA | Vol | N | BA | Vol |
| Lower layer | $14 \%$ | $2 \%$ | $1 \%$ | $16 \%$ | $3 \%$ | $1 \%$ | $16 \%$ | $3 \%$ | $1 \%$ |
| Medium layer | $58 \%$ | $32 \%$ | $24 \%$ | $54 \%$ | $30 \%$ | $22 \%$ | $56 \%$ | $31 \%$ | $23 \%$ |
| Upper layer | $28 \%$ | $66 \%$ | $75 \%$ | $30 \%$ | $67 \%$ | $77 \%$ | $28 \%$ | $66 \%$ | $76 \%$ |
| Total stand | 873 | 21.4 | 181 | 902 | 22.7 | 196 | 700 | 19.8 | 173 |
|  | trees/ha | $\mathrm{m}^{2} / \mathrm{ha}$ | $\mathrm{m}^{3} / \mathrm{ha}$ | trees $/ \mathrm{ha}$ | $\mathrm{m}^{2} / \mathrm{ha}$ | $\mathrm{m}^{3} / \mathrm{ha}$ | trees $/ \mathrm{ha}$ | $\mathrm{m}^{2} / \mathrm{ha}$ | $\mathrm{m}^{3} / \mathrm{ha}$ |

Table 9 shows theoretical removals, if a particular treatment would have been applied completely on all 48 plots. Then, the conventional target diameter cutting would have been reduced the standing volume from 322 to $180 \mathrm{~m}^{3}$ per ha ( 12 and $9 \mathrm{~m}^{3} \mathrm{SE}$ ). Treatment TN would have reduced the volume to $197 \mathrm{~m}^{3}$ per ha ( $11 \mathrm{~m}^{3} \mathrm{SE}$ ).

Table 9. Initial stand characteristics and theoretical removals if a particular treatment would had been applied on all 48 plots. Initial beståndskarakteristik och teoretiska uttag om de olika
behandlingarna hade tillämpats på samtliga av 48 provytorna.

| Species | before cut |  |  |  | removals according to T or TS |  |  |  | removals according to TN |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} \mathrm{N} \\ {\left[\mathrm{ha}^{-1}\right]} \end{gathered}$ | $\left\lvert\, \begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{gathered}\right.$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\left\|\begin{array}{c} \mathrm{Vol} \\ {\left[\mathrm{~m}^{3} / \mathrm{ha}\right]} \end{array}\right\|$ | $\begin{gathered} \mathrm{N} \\ {\left[\mathrm{ha}^{-1}\right]} \end{gathered}$ | $\left\|\begin{array}{c} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{array}\right\|$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\left.\left\lvert\, \begin{array}{c} \mathrm{Vol} \\ {\left[\mathrm{~m}^{3} / \mathrm{ha}\right]} \end{array}\right.\right]$ | $\begin{gathered} \mathrm{N} \\ {\left[\mathrm{ha}^{-1}\right]} \end{gathered}$ | $\left\|\begin{array}{c} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{array}\right\|$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\left.\begin{array}{\|c\|} \mathrm{Vol} \\ {\left[\mathrm{~m}^{3} / \mathrm{ha}\right]} \end{array} \right\rvert\,$ |
| Spruce | 489 | 13.8 | 18.9 | 131 | 440 | 9.0 | 16.1 | 78 | 422 | 7.8 | 15.3 | 65 |
| Pine | 164 | 15.4 | 34.6 | 145 | 109 | 8.6 | 31.7 | 76 | 136 | 11.1 | 32.3 | 100 |
| Birch | 54 | 2.6 | 24.6 | 21 | 26 | 0.4 | 13.4 | 3 | 38 | 1.0 | 18.2 | 8 |
| Beech | 21 | 0.4 | 15.1 | 3 | 21 | 0.4 | 15.1 | 3 | 21 | 0.4 | 15.1 | 3 |
| Oak | 271 | 3.0 | 11.8 | 21 | 271 | 3.0 | 11.8 | 21 | 271 | 3.0 | 11.8 | 21 |
| Total | 999 | 35.2 | 21.2 | 322 | 866 | 21.3 | 17.7 | 180 | 887 | 23.3 | 18.3 | 197 |

## Canopy gaps

In the gap survey, 174 gaps were recorded. The total gap area was $28.620 \mathrm{~m}^{2}$, equal to $15 \%$ of total stand area. But the logging area was smaller: Excluding control treatment and some wet and steep sections, approximately 15 ha were logged. Furthermore, 49 gaps did not contain recently cut trees and covered $3.960 \mathrm{~m}^{2}$ in total.


Figure 9. Floor of a medium-sized forest gap with about $500 \mathrm{~m}^{2}$ canopy opening (after soil preparation). Medelstor beståndslucka (ca. $500 \mathrm{~m}^{2}$ ) efter markberedning.

Different sizes of canopy gaps were found. The largest gap was $1.723 \mathrm{~m}^{2}$. In this gap, 25 trees were harvested in 2007, and 26 older tree remnants were also found. 112 gaps were smaller than $100 \mathrm{~m}^{2}$, making up to $24 \%$ of the total gap area (Figure 10 A ). In this size class, 2.9 remnants of trees were found per gap (of which 1.4 trees were cut in 2007). $16 \%$ of gaps (representing $13 \%$ of gap area) were recorded in the $100-200 \mathrm{~m}^{2}$ class. Gaps larger than $1000 \mathrm{~m}^{2}$ covered $18 \%$ of total gap area. Eleven gaps between 500 and $1000 \mathrm{~m}^{2}$ covered $24 \%$.
Figure 10B points out about 15-20 cut trees or remnants for gaps with $500 \mathrm{~m}^{2}$ and about 30 such gap makers for $1.000 \mathrm{~m}^{2}$-openings.


Figure 10. Gap size distribution (A) and relationship between gap size and number of removed (or dead) trees in the gap (B). Fördelning av beståndsluckor (Bild A) och samband mellan luckstorlek och antalet avverkade träd per lucka (Bild B).

Figure 11 and 2 demonstrate how gaps were scattered in the stand. More large gaps were found in block 2 and 3. There was also a tendency of gap occurrence along the skid roads. Regarding the gap making trees, 558 fresh and 329 older stumps, plus 172 remnants identified as wind thrown, were found in gaps. In nine gaps, single living trees in the gap were found.


Figure 11. Spatial distribution of differently sized gaps (illustrated as circles which are based on the GPS coordinates in the gap centre). Geografisk läge för luckor med olika storlek.

The crown coverage recorded in the centre of gaps to characterize light conditions amounted from $75 \%$ in smallest gaps to $2 \%$ in the largest gap (Figure 12). Figure 13 illustrates that diffuse site factors below $10 \%$ compared to open land conditions occur under closed canopy, while the relationship between diffuse site factor and crown coverage is rather between 10 and $70 \%$. In conclusion, the maximum diffuse site factors in gaps ranged from $30 \%$ in small gaps to $70 \%$ in the largest gap.


Figure 12. Gap size and crown coverage (covering the 24 degree opening of an imaginary cone from a terrestrial view) recorded in the centre of gaps. Samband mellan luckstorlek och krontäckning (-bedömt i luckcentrum och vertikalt uppåt inom en öppningmed +/- $12^{\circ}$ vinkel).


Figure 13. Correlation between diffuse site factor and crown coverage, evenly distributed across the whole range of light conditions in the forest. Samband mellan Diffuse site factor och krontäckning över hela ljusgradienten

On the 192 permanent regeneration plots in the stand, the crown coverage was in average $89.8 \%$ (+/- $1.2 \%$ SE), ranging from 12 to $100 \%$.

## Regeneration

In the stand, 51 individuals were found on 196 sample plots. Only $19 \%$ of the plots were occupied with regeneration. Figure 14 shows a spot where some regeneration occurred. The regeneration density was 530 individuals/ha. On occupied plots, the density was 3.000 plants/ha. Half of these plants were smaller than 50 cm (Table 10).


Figure 14. Single, naturally regenerated seedlings of spruce and beech in the stand. Enstaka självföryngrade plantor av gran och bok inom beståndet.

Table 10. Absolute number of regenerated trees in different height classes on the regeneration plots (Total sample area $960 \mathrm{~m}^{2}$ ). Height class 20-49 cm in parenthesis. Antal plantor inom olika höjdklasser registrerade inom provtagningsområdet med en total areal av $960 \mathrm{~m}^{2}$ ( 192 ytor á $5 \mathrm{~m}^{2}$ systematisk fördelad inom beståndet).

| Height <br> class [cm] | $10-49$ | $50-99$ | $100-$ <br> 199 | $200-$ <br> 299 | $\geq 300$ |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Spruce | $13(6)$ | 4 | 3 | 5 | 5 |
| Birch | 1 |  |  |  |  |
| Oak | $1(1)$ |  |  | 1 | 1 |
| Beech | $3(3)$ | 3 |  |  |  |
| Rowan | $10(3)$ | 1 |  |  |  |
| Total | $27(13)$ | 9 | 3 | 6 | 6 |

Average top shoot length per year was about 5 cm across all tree species (Table 11). Dominant tree species in the regeneration was spruce (54\%), followed by rowan (20\%). The frequency of beech was $11 \%$ and oak was $5 \% .68 \%$ of those trees were smaller than 1 m . No pine was found.

Table 11. Average annual top shoot length [cm] considering the last three years growth of the highest individuals per species on the regeneration plots. Årlig höjdtillväxt [cm] hos de högsta plantorna per yta i föregående tabellen under de tre år som föregick huggningen.

| Height <br> class [cm] | $10-49$ | $50-99$ | $100-$ <br> 199 | $200-$ <br> 299 | $\geq 300$ |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Spruce | 2 | 2 | 6 | 3 | 8 |
| Birch | - | - | - | - | - |
| Oak | 15 | - | - | - | 5 |
| Beech | - | 6 | - | - | - |
| Rowan | 4 | 10 | - | - | - |

Regeneration in gaps 2.854 individuals were counted in gaps. Thus, regeneration density was 1.003 individuals/ha. No regeneration was found in 16 of the 174 gaps. Most frequent was spruce ( $46 \%$ ), rowan ( $22 \%$ ) and birch ( $17 \%$ ). Beech and oak represented $4 \%$ each. Salix, poplar and other species represented $2 \%$ of trees. The proportion of pine was below $0.5 \%$ (Figure 15). The distribution in height classes in Table 12 shows a dominance of the lowest class for all main species with an exponentially decreasing number of seedlings in larger size classes. The proportion of tree species was relatively stable over different size classes, although the decrease of oak and beech seedling numbers was less pronounced than for others. 1298 trees/ha were found in gaps < $200 \mathrm{~m}^{2}$, and 820 trees/ha in gaps > $200 \mathrm{~m}^{2}$. The standard error of the tree density in single gaps was 141 trees/ha.

Table 12. Absolute seedling number observed in gaps distributed over height classes (Total sample area $28.620 \mathrm{~m}^{2}$ ). Antal plantor (mindre än 5 cm BHD) inom olika höjdklasser registrarade i beståndsluckorna.

| Height <br> class $[\mathrm{cm}]$ | $20-49$ | $50-99$ | $100-149$ | $150-$ <br> 199 | $200-$ <br> 249 | $250-$ <br> 299 | $\geq 300$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Spruce | 672 | 310 | 131 | 93 | 41 | 21 | 32 |
| Birch | 284 | 129 | 48 | 16 | 7 | 2 | 5 |
| Oak | 77 | 14 | 8 | 5 | 9 | 3 | 7 |
| Beech | 48 | 33 | 21 | 5 | 3 | 6 | 3 |
| Pine | 14 | 1 | - | - | - | - | - |
| Rowan | 392 | 136 | 76 | 27 | - | - | 1 |
| Others | 71 | 84 | 15 | - | 2 | - | 2 |
| Total | 1558 | 707 | 299 | 146 | 62 | 32 | 50 |



- Spruce

Birch
Oak
Beech
Pine
Rowan
Others

Figure 15. Tree species distribution of the regeneration in gaps.
Trädslagsfördelning hos självföryngringen i luckorna omedelbar efter huggning (gran, björk, ek, bok, tall, rönn, övriga trädslag).

Average top shoot length of spruce was 11 cm , with about 5 cm for small seedlings and about 20 cm for the tallest individuals (Table 13). Top shoot length across all species was 12.5 cm .

Table 13. Annual top shoot lengths [in cm ] of the last three years of regeneration in gaps in height classes [in cm]. Number of shoot length measurements in parenthesis. Årlig höjdtillväxt [cm] hos de högsta plantorna per lucka under de tre år som föregick huggningen (Antal av mätningar i parentes).

|  | $20-49$ | $50-99$ | $100-149$ | $150-199$ | $200-249$ | $250-299$ | $\geq 300$ | All size <br> classes |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Spruce | $3.8(48)$ | $6.4(57)$ | $9.3(43)$ | 13.1 <br> $(51)$ | $9.7(39)$ | $6.4(36)$ | 19.1 <br> $(60)$ | 11.2 <br> $(334)$ |  |
|  |  |  |  |  |  |  |  |  |  |
| Birch | 20.5 | 16.4 | 18.0 | 13.6 | 26.7 | $16.8(6)$ | $28.6(7)$ | 18.9 |  |
|  | $(26)$ | $(42)$ | $(22)$ | $(12)$ | $(12)$ |  |  | $(127)$ |  |
| Oak | $6.1(63)$ | 10.6 | 19.5 | $6.4(11)$ | $5.6(20)$ | 7.7 | 12.5 | 8.7 |  |
|  |  | $(19)$ | $(14)$ |  |  | $(6)$ | $(19)$ | $(152)$ |  |
| Beech | $6.2(12)$ | $9.5(44)$ | 15.9 | 23 | 22.2 | 31.5 | $24.4(9)$ | 15.8 |  |
|  |  |  | $(12)$ | $(3)$ | $(20)$ | $(16)$ |  | $(102)$ |  |
| Pine | $6.5(29)$ | 7.3 | 0 | 0 | 0 | 0 | 0 | $6.6(32)$ |  |
|  |  | $(3)$ |  |  |  |  |  |  |  |
| Others | 9.5 | 13.6 | 21.8 | 17.9 | 15 | 0 | 17.8 | $(5)$ | 13 |
|  | $(133)$ | $(117)$ | $(37)$ | $(10)$ | $(2)$ |  |  | $(304)$ |  |

Estimated from the heights of the largest individuals in the regeneration and the height curves of trees above 5 cm DBH, the height of spruce (and pine) with 5 cm DBH was assumed as 4 m . For other tree species, the height was approximately 5.5 m (see Figures A6-A8 in Appendix A).
The last height shoot was damaged by browsing on $57 \%$ of the broadleaved tree species. This was observed for the 474 individuals measured for height. In contrast only $2 \%$ of the spruce seedlings were browsed (Table 14).

Table 14. Browsing damages of different tree species of regeneration in gaps. Betesskador hos olika trädslag I självföryngringen.

| Tree species | Total tree <br> number | Damaged by <br> browsing | Damaged by <br> harvest |
| :--- | :---: | :---: | :---: |
| Spruce | 128 | $2 \%$ | $22 \%$ |
| Oak | 74 | $51 \%$ | $22 \%$ |
| Birch | 73 | $42 \%$ | $19 \%$ |
| Beech | 43 | $37 \%$ | $28 \%$ |
| Pine | 12 | $33 \%$ | $8 \%$ |
| Rowan | 106 | $73 \%$ | $3 \%$ |
| Willow | 20 | $90 \%$ | $5 \%$ |
| Others | 18 | $50 \%$ | $11 \%$ |
| Total | 474 | $41 \%$ | $16 \%$ |

## Simulation study

## The derivation of ingrowth scenarios

As stated in the methodology section, ingrowth scenarios with one, four and ten new trees per year and ha were used. However, the minimum ingrowth scenario represents expected ingrowth based on the features of advanced regeneration today. Table 15 shows expected ingrowth in five year steps on a 19 ha base, calculated from tree numbers in Table 12, height growth in Table 13, and mortality rates in Table 2. One new tree per year and ha would be equal to 95 new trees every 5 years in Table 15 ( 95 trees/19 ha $=1$ tree $\times 19$ ha $\times 5$ years).

Table 15. Expected ingrowth based on characteristics of current advanced regeneration in gaps (black) and additional numbers (grey) to fulfill the average rates (bold) of the scenarios. Numbers are new trees on 19 ha total stand area after 5 years. Förväntad inväxning inom hela beståndsarealen på 19 ha enligt tre olika scenarier (Minimum scenariot antar 10 cm årlig höjdtilluäxt och hög mortalitet, Maximum: 30 cm per år och ingen dödlighet. Se samfattning)

|  | year | 2013 | 2018 | 2023 | 2028 | 2033 | 2038 | 2043 | 2048 | 2053 | 2058 |  |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| scenario | period | 5 | 10 | 15 | 20 | 25 | 30 | 35 | 40 | 45 | 50 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| M |  | spruce | 30 | 19 |  | 34 | 150 | 114 | 141 | 154 | 154 | 154 |
| I | $\mathbf{1}$ tree | birch | 2 |  | 5 | 15 | 36 |  |  |  |  |  |
| N | per ha | oak |  | 6 |  |  | 2 |  | 6 | 2 | 4 | 4 |
|  |  | beech |  | 3 | 5 | 7 | 15 | 14 | 16 |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| M |  |  |  |  |  |  |  |  |  |  |  |  |
| E |  | spruce | 52 | 116 | 344 | 465 | 526 | 526 | 526 | 526 | 526 | 526 |
| D | 4 trees | birch | 4 |  | 6 | 35 | 138 |  |  |  |  |  |
| I | per ha | oak | 7 |  | 13 | 11 | 60 |  |  |  |  |  |
| U |  | beech | 3 |  | 8 | 22 | 55 |  |  |  |  |  |
| M |  |  |  |  |  |  |  |  |  |  |  |  |


| M | $\mathbf{1 0}$ | spruce | 94 | 535 | 672 | 1171 | 1171 | 1171 | 1171 | 1171 | 1171 | 1171 |
| :--- | :---: | :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| A | trees | birch | 14 | 193 | 284 |  |  |  |  |  |  |  |
| X | per ha | oak | 19 | 27 | 77 |  |  |  |  |  |  |  |
|  |  | beech | 28 | 59 | 48 |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |

Assuming the height growth in Table 13 and mortality rates in Table 2, the black numbers in Table 15 indicate tree numbers that would pass the 5 cm threshold in five year periods. Most of the small trees would have reached 5 cm DBH after 15,25 or 35 years depending on the scenario. The grey figures indicate tree numbers that had to be added to achieve the average ingrowth rates applied in the scenario construction. The ingrowth during this latter period corresponds to $1.6,5.5$ and 12.3 trees per year and ha for the whole stand.

Box 2. Calculation of mortality estimates indicated by adjacent height classes
The presumed mortality in Table 2 was corroborated by the decline of tree numbers from one height class to a larger class (Table 12) and the estimated time period to grow from the medium height of one class to another (e.g. from 35 cm to 75 cm for the two smallest height classes). The mean annual height growth of both classes (Table 13) determined the time period to grow to the next height class (e.g. spruce in the smallest class: 7.8 years $=(75 \mathrm{~cm}-35 \mathrm{~cm}) / 5.1 \mathrm{~cm})$. This time period and the tree number in both height classes were used to calculate an annual decline of tree number (which is in the example
$\left.9.1 \%=1-\left(\frac{311 \text { trees }}{672 \text { trees }}\right)^{\frac{1}{7.8} \text { years }}\right)$.
All tree species showed a reasonable trend of decreasing mortality with increasing height. While tree number decline of spruce and oak was stable, the limited number of individuals caused a high variation for beech. However, more apt mortality rates used for the simulation (Table 2) resulted in the same number of beech trees for the final scenarios as the rates of 5\% (<1m), 15\% (1-2 m) and 0\% ( $>2 \mathrm{~m}$ ) derived directly from advanced regeneration. In general, mortality rates of tree species derived from advanced regeneration were similar to rates assumed in Table 2.

## Future increment and harvests

Without tree removals, the growth simulation projected a BA increase from 35 to $53 \mathrm{~m}^{2} / \mathrm{ha}$ during the 50 year simulation period (Figure 16). If losses by natural mortality were included, the MAI of BA was to $0.53 \mathrm{~m}^{2}$ per year and ha.
For silvicultural treatments, the simulation projected BA levels from 18 to $35 \mathrm{~m}^{2} / \mathrm{ha}$, with 2025 years intervals between tree removals (Figure 16). The MAI ranged between 5.6-6.4 $\mathrm{m}^{3} / \mathrm{ha}$, depending on ingrowth. Natural mortality caused annual losses of four trees per ha (equal to $0.8 \mathrm{~m}^{3} / \mathrm{ha}$ ). All management treatments did not indicate a growth regression.


Figure 16. Basal area development of different treatments according to Elfving (2004). Projections are based on stand age-independent single tree-ages (Elfving, 2003), $100 \mathrm{~m}^{3} / \mathrm{ha}$ minimum standing volume after cut, and the medium ingrowth scenario.
Grundytans utveckling enligt vid tillämpning av de fyra olika behandlingarna T, TS, TN och C.

Initial volume estimates according to Ekö (1985) differed 0-13 m³/ha from the values calculated according to Söderberg (1992). Stand age-independent and -dependent projections did not differ significantly as shown in the next two section.

Control treatment In terms of projected volume, MAI was $6.7 \mathrm{~m}^{3}$ per ha. No ingrowth was implemented due to high stand densities on all plots over time. Starting with $322 \mathrm{~m}^{3} / \mathrm{ha}$, the standing volume increased to $571 \mathrm{~m}^{3} / \mathrm{ha}$ (Table 16). The stand age-dependent estimation predicted $574 \mathrm{~m}^{3} /$ ha volume and $53.4 \mathrm{~m}^{2} / \mathrm{ha} \mathrm{BA}$. Mortality reduced the tree number from 999 to 764 trees/ha. These losses represented $84 \mathrm{~m}^{3} /$ ha dead wood during the whole simulation period (resp. $8.3 \mathrm{~m}^{2} / \mathrm{ha} \mathrm{BA}$ ), constantly increasing over the simulation period from 1.2 to $2.2 \mathrm{~m}^{3} /$ ha annually. The total volume production during the period amounted to $334 \mathrm{~m}^{3} / \mathrm{ha}$. Supplementing Table 16, Table C2 and C3 in Appendix C provide numbers to each scenario and growth model. - During the simulation period, BA growth declined from 0.57 $\mathrm{m}^{2} / \mathrm{ha}$ in the first decade to $0.48 \mathrm{~m}^{2} / \mathrm{ha}$ in the last decade.

Table 16. Control. Initial and simulated stand characteristics after 50 years. Projections are based on 48 plots and the stand age-independent estimation of single-tree ages.
Behandling C (kontroll). Initial och simulerad beståndskarakteristik efter 50 år för alla 48 provytor.

| Species | start of simulation |  |  |  | MAI |  | total | end of simulation |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} \mathrm{N} \\ \text { [ha] } \end{gathered}$ | $\left\|\begin{array}{c} B A \\ {\left[m^{2} / h a\right]} \end{array}\right\|$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\left[\begin{array}{c} \text { Vol } \\ {\left[\mathrm{m}^{3} / \mathrm{ha}\right]} \end{array}\right]$ | $\left\|\begin{array}{c} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{array}\right\|$ | $\left.\left\lvert\, \begin{array}{c} \text { Vol } \\ {\left[\mathrm{m}^{3} / \mathrm{ha}\right]} \end{array}\right.\right]$ | $\begin{gathered} \text { mortality } \\ {\left[\mathrm{m}^{3} / \mathrm{ha}\right]} \\ \hline \end{gathered}$ | $\begin{gathered} \mathrm{N} \\ {[\mathrm{ha}]} \end{gathered}$ | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\begin{gathered} \text { Vol } \\ {\left[\mathrm{m}^{3} / \mathrm{ha}\right]} \end{gathered}$ |
| Spruce | 489 | 13.8 | 18.9 | 131 | 0.25 | 3.3 | 35 | 393 | 23.1 | 27.3 | 260 |
| Pine | 164 | 15.4 | 34.6 | 145 | 0.18 | 2.4 | 31 | 139 | 21.6 | 44.5 | 233 |
| Birch | 54 | 2.6 | 24.6 | 21 | 0.02 | 0.2 | 6 | 38 | 2.8 | 31.1 | 25 |
| Beech | 21 | 0.4 | 15.1 | 3 | 0.01 | 0.2 | 1 | 15 | 0.8 | 26.2 | 9 |
| Oak | 271 | 3.0 | 11.8 | 21 | 0.06 | 0.7 | 12 | 179 | 4.8 | 18.5 | 43 |
| Total | 999 | 35.2 | 21.2 | 322 | 0.53 | 6.7 | 84 | 764 | 53.1 | 29.8 | 571 |

At the start of the simulation period, half of the plots ranged between 30 and $40 \mathrm{~m}^{2} / \mathrm{ha}$ BA (Figure 17). After 50 years, $50 \%$ of the plots ranged between 46 and $58 \mathrm{~m}^{2} / \mathrm{ha}$. The extreme values among the 48 plots were 33 and $71 \mathrm{~m}^{2} / \mathrm{ha}$.


Figure 17. Initial variation of basal area on plots and projected basal area for the control treatment after 50 years.
Grundytans variation mellan provytorna i början och slutet av simuleringsperioden enligt behandling C.

Figure 18 shows the decrease in numbers of small trees over time. For instance, trees with DBH < 19 cm decreased from 600 to 250 trees per ha. There are no trees in the smallest diameter class with 5-6 cm DBH.


Figure 18. Initial and projected diameter distributions [in 2 cm classes] projected for the control treatment at age 0,25 and 50 (stand age-independent estimation of tree ages). Initial och beräknad BHD-fördelning efter 25 och 50 år inom behandlingen $C$ utan avverkning [ 2 cm BHD klasser!].

Treatment T After 50 years, tree number decreased by 194 trees per ha. The MAI for this period was projected with $5.8-6.3 \mathrm{~m}^{3} / \mathrm{ha}$, depending on the ingrowth scenario (Table C5 in the Appendix). According to the moderate ingrowth scenario, about $120 \mathrm{~m}^{3} / \mathrm{ha}$ of the standing volume would be removed after 25 and 45 years (Table 17). Standing volume did not fall below $150 \mathrm{~m}^{3} /$ ha after the next two simulated harvests, independent from the two thresholds initially set for simulation. As long as standing volume before cut was minimum $300 \mathrm{~m}^{3} / \mathrm{ha}$, no effects of the two thresholds on harvest removal and remaining stand volume were detected. Simulation results with other ingrowth scenarios are shown in Table C5 in Appendix C.

Table 17. Treatment T-medium ingrowth scenario, initial and simulated stand parameters after 50 years (based on 48 plots, stand age-independent single-tree ages).

| Species | start of simulation |  |  |  | removals after |  |  |  | MAI |  | end of simulation |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\left\|\begin{array}{c} N \\ {\left[\mathrm{ha}^{-1}\right]} \end{array}\right\|$ | $\left\|\begin{array}{c} B A \\ {\left[m^{2} / h a\right]} \end{array}\right\|$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\left\|\begin{array}{c} \text { Vol } \\ {\left[\mathrm{m}^{3} / \mathrm{ha}\right]} \end{array}\right\|$ | $\left[\mathrm{m}^{2} / \mathrm{ha}\right]$ | $\begin{aligned} & \text { ears } \\ & {\left[\mathrm{m}^{3} / \mathrm{ha}\right]} \end{aligned}$ | $\begin{array}{r} 45 \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{array}$ | $\begin{aligned} & \text { ears } \\ & {\left[m^{3} / h a\right]} \end{aligned}$ | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{gathered}$ | $\left\lvert\, \begin{gathered} \mathrm{Vol} \\ {\left[\mathrm{~m}^{3} / \mathrm{ha}\right]} \end{gathered}\right.$ | $\left\|\begin{array}{c} N \\ {\left[h a^{-1}\right]} \end{array}\right\|$ | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\left\lvert\, \begin{gathered} \mathrm{Vol} \\ {\left[\mathrm{~m}^{3} / \mathrm{ha}\right]} \end{gathered}\right.$ |
| Spruce | 440 | 9.0 | 16.1 | 77 | 4.2 | 46 | 4.6 | 52 | 0.30 | 3.3 | 410 | 13.0 | 20.1 | 127 |
| Pine | 109 | 8.6 | 31.7 | 76 | 6.0 | 60 | 5.2 | 51 | 0.11 | 1.3 | 21 | 2.4 | 37.5 | 22 |
| Birch | 26 | 0.4 | 13.4 | 3 | 0.2 | 2 | 0.1 | 1 | 0.01 | 0.1 | 32 | 0.4 | 12.8 | 3 |
| Beech | 21 | 0.4 | 15.1 | 3 | 0.4 | 4 | 0.5 | 5 | 0.02 | 0.2 | 8 | 0.3 | 22.1 | 3 |
| Oak | 271 | 3.0 | 11.8 | 21 | 0.3 | 2 | 1.3 | 12 | 0.13 | 1.1 | 200 | 7.1 | 21.2 | 55 |
| Total | 866 | 21.3 | 17.7 | 180 | 11.0 | 114 | 11.7 | 121 | 0.56 | 5.9 | 672 | 23.2 | 21.0 | 209 |

A block-wise simulation according to the medium ingrowth scenario revealed MAI ranging from $5.8 \mathrm{~m}^{3} /$ ha in block 2 to $6.3 \mathrm{~m}^{3} /$ ha in block 3 with higher proportion of spruce (Table C7C 12 in Appendix C). Cutting intervals were about 25 years. The volume proportion of spruce increased from 58 to $81 \%$ in block 3, and from 30 to $42-44 \%$ in the two others.

The MAI of BA was $0.56 \mathrm{~m}^{2} /$ ha (or 0.53 and 0.60 with min and max ingrowth). Thus, BA growth was predicted $7 \%$ higher for treatment $T$ compared to the control, while volume growth was $11 \%$ lower than the control!
During the simulation period, BA growth declined from $0.62 \mathrm{~m}^{2} / \mathrm{ha}$ in the first decade to 0.51 $\mathrm{m}^{2} /$ ha in the last decade.

Figure 19 shows the variation of BA between sample plots (according to the moderate ingrowth scenario and stand age-independent single tree-ages). The mean values were reduced by $15 \mathrm{~m}^{2} /$ ha during the simulation period, but the variation did not decrease after cuttings.

Figure 19. Box plot with median, $10^{\text {th }}, 25^{\text {th }}, 75^{\text {th }}$, and $90^{\text {th }}$ percentiles of the basal area on plots at different times during the simulation (treatment T ). Grundytans variation mellan provytorna i början och slut av simuleringsperioden (Behandling $T$ ).


Beside reduced tree number, the diameter distributions in Figure 20 indicated a lack of small trees and a surplus of trees with $15-20 \mathrm{~cm}$ DBH after 50 years. The minimum ingrowth scenario resulted in a bell-shaped distribution with a high number of trees between 15-30 cm DBH. More evenly distributed tree numbers across size classes was projected within the moderate scenario. The maximum scenario indicated a peak in DBH class $8-10 \mathrm{~cm}$ with 150 trees after 25 years, and a decrease of tree numbers with increasing size for trees larger than 15 cm DBH after 50 years. For smaller trees, a decreasing number was projected. The proportion of spruce at the end of the simulation period was $55-66 \%$ of the total tree number, and $55-59 \%$ of total BA.


Figure 20. Projected diameter distributions according to treatment $T$ and different ingrowth scenarios at the age of 25 and 50 years (based on 48 plots and stand age-independent single-tree age). Beräknade BHD-fördelningar efter 25 och 50 år enligt behandling T med tre olika inväxtscenarierna.

Treatment TS Depending on new trees after scarification, the MAI was $5.8-6.4 \mathrm{~m}^{3} / \mathrm{ha}$ volume respective $0.55-0.64 \mathrm{~m}^{2} /$ ha BA (see Table C 13 in Appendix C for different ingrowth scenarios). At the end of the simulation period, the number of living trees was 30-100\% higher compared to treatment T. The final standing volume, five years after the last harvest, was $214 \mathrm{~m}^{3} / \mathrm{ha}$ according to the moderate ingrowth scenario (Table 18).

Table 18. Treatment TS - moderate ingrowth, initial and simulated stand parameters after 50 years (based on 48 plots, stand age-independent single-tree ages). Behandling TS med moderat inväxningsscenario - initiala och simulerade beståndsparametrar efter 50 år.

|  | start of simulation |  |  |  | removals after |  |  |  | MAI |  | end of simulation |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | $\left\|\begin{array}{c} N \\ {\left[h a^{-1}\right]} \end{array}\right\|$ | $\left\|\begin{array}{c} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{array}\right\|$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \\ \hline \end{gathered}$ | $\left.\left\lvert\, \begin{array}{c} \text { Vol } \\ {\left[\mathrm{m}^{3} / \mathrm{ha]}\right.} \end{array}\right.\right]$ | $\left\lvert\, \begin{array}{r} 25 y \\ {\left[m^{2} / h a\right]} \end{array}\right.$ | $\begin{aligned} & \text { eears } \\ & {\left[m^{3} / \mathrm{ha}\right]} \end{aligned}$ | $\left\lvert\, \begin{array}{r} 45 y \\ {\left[m^{2} / h a\right]} \end{array}\right.$ | $\begin{aligned} & \text { years } \\ & {\left[\mathrm{m}^{3} / \mathrm{ha}\right]} \end{aligned}$ | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{gathered}$ | $\left\|\begin{array}{c} \mathrm{Vol} \\ {\left[\mathrm{~m}^{3} / \mathrm{ha}\right]} \end{array}\right\|$ | $\left\|\begin{array}{c} \mathrm{N} \\ {\left[h a^{-1}\right]} \end{array}\right\|$ | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\begin{gathered} \text { Vol } \\ {\left[\mathrm{m}^{3} / \mathrm{ha}\right]} \end{gathered}$ |
| Spruce | 440 | 9.0 | 16.1 | 78 | 4.2 | 45 | 4.5 | 51 | 0.30 | 3.3 | 545 | 13.4 | 17.7 | 128 |
| Pine | 109 | 8.6 | 31.7 | 76 | 6.0 | 60 | 5.2 | 51 | 0.13 | 1.4 | 143 | 3.0 | 16.4 | 26 |
| Birch | 26 | 0.4 | 13.4 | 3 | 0.2 | 2 | 0.1 | 1 | 0.01 | 0.1 | 48 | 0.4 | 10.3 | 3 |
| Beech | 21 | 0.4 | 15.1 | 3 | 0.4 | 4 | 0.5 | 5 | 0.02 | 0.2 | 8 | 0.3 | 22.0 | 3 |
| Oak | 271 | 3.0 | 11.8 | 21 | 0.3 | 2 | 1.3 | 12 | 0.13 | 1.1 | 208 | 7.1 | 20.9 | 55 |
| Total | 866 | 21.3 | 17.7 | 180 | 11.1 | 113 | 11.58 | 120 | 0.59 | 6.0 | 952 | 25.2 | 18.0 | 214 |

According to the maximum scenario, the standing volume would be $239 \mathrm{~m}^{3} / \mathrm{ha}$. For the minimum scenario, $172 \mathrm{~m}^{3} /$ ha were estimated just after harvest (see Tables C 13 in Appendix C). Stand age-dependent single-tree ages caused lower growth predictions the more ingrowth was assumed (Table C14 in the Appendix). The variation of the BA on plots during the simulation period in Figure 21 was very similar to treatment T .

Figure 21. Box plot with median, $10^{\text {th }}, 25^{\text {th }}, 75^{\text {th }}$, and $90^{\text {th }}$ percentiles of the basal area on plots at different times during the simulation (treatment TS). Grundytans variation mellan provytorna i början och slutet av simuleringsperioden (Behandling TS).


Future diameter distributions show a large number of small trees after 25 or 50 years (Figure 22). In the medium scenario, 618 trees/ha with $4.5-12 \mathrm{~cm}$ DBH were estimated, with 1113 trees/ha in total. The minimum scenario resulted in 262 trees below 12 cm DBH and 692 trees/ha in total, while the maximum scenario projected 591 trees below 12 cm DBH and 1201 trees/ha in total. During the simulation period, BA proportion of spruce increased from $39 \%$ to $57 \%$. In terms of tree number, spruce increased from initially 50 to $60-63 \%$ (Table C13 in the Appendix).
Figure 22．Projected development of the diameter distribution according to treatment TS and different ingrowth scenarios at the age of 25 and 50 years （based on stand age－independent single－tree age）．
Beräknad BHD－fördelning efter 25 och 50 år enligt behandling TS med tre olika inväxtscenarierna．






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Treatment TN had the lowest standing volume after 50 years, and a low number of trees (Table 19). The MAI was simulated with $5.6-6.1 \mathrm{~m}^{3} /$ ha volume respective $0.53-0.61 \mathrm{~m}^{2} / \mathrm{ha}$ $B A$. Figure 23 indicates a decreased variation of BA on plots after heavier removals by the $3^{\text {rd }}$ cutting.

Table 19. Treatment TN - moderate ingrowth scenario, initial and simulated stand parameters after 50 years (based on stand age-independent single-tree ages). Behandling $\mathrm{TN}^{2}$ med moderat inväxningsscenario - initiala och simulerade beståndsparametrar efter 50 år.

|  | start of simulation |  |  |  | removals after |  |  |  | MAI |  | end of simulation |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | $\left[\begin{array}{c} \mathrm{N} \\ {\left[\mathrm{ha}^{-1}\right]} \end{array}\right]$ | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\left[\begin{array}{c} \text { Vol } \\ {\left[\mathrm{m}^{3} / \mathrm{ha]}\right]} \end{array}\right.$ | $\begin{array}{r} 20 \mathrm{y} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{array}$ | $\begin{aligned} & \text { eears } \\ & {\left[\mathrm{m}^{3} / \mathrm{ha}\right]} \end{aligned}$ | $\left\lvert\, \begin{array}{r} 45 y \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{array}\right.$ | $\begin{aligned} & \text { eears } \\ & {\left[\mathrm{m}^{3} / \mathrm{ha}\right]} \end{aligned}$ | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{gathered}$ | $\left[\begin{array}{c} \mathrm{Vol} \\ {\left[\mathrm{~m}^{3} / \mathrm{ha}\right]} \end{array}\right.$ | $\left\|\begin{array}{c} N \\ {\left[\mathrm{ha}^{-1}\right]} \end{array}\right\|$ | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\left[\begin{array}{c} \mathrm{Vol} \\ {\left[\mathrm{~m}^{3} / \mathrm{ha}\right]} \end{array}\right.$ |
| Spruce | 422 | 7.8 | 15.3 | 65 | 5.1 | 53 | 6.6 | 70 | 0.25 | 2.7 | 304 | 6.9 | 17.0 | 60 |
| Pine | 136 | 11.1 | 32.3 | 100 | 6.0 | 60 | 8.0 | 81 | 0.14 | 1.6 | 29 | 3.2 | 37.4 | 29 |
| Birch | 38 | 1.0 | 18.2 | 8 | 0.6 | 5 | 0.3 | 3 | 0.01 | 0.1 | 32 | 0.5 | 14.8 | 4 |
| Beech | 21 | 0.4 | 15.1 | 3 | 0.0 | 0 | 0.0 | 0 | 0.02 | 0.3 | 22 | 1.4 | 28.8 | 15 |
| Oak | 271 | 3.0 | 11.8 | 21 | 0.1 | 1 | 0.1 | 1 | 0.13 | 1.1 | 222 | 8.4 | 21.9 | 65 |
| Total | 887 | 23.3 | 18.3 | 197 | 11.7 | 119 | 15.0 | 155 | 0.55 | 5.7 | 609 | 20.4 | 20.7 | 174 |

Figure 23. Box plot with median, $10^{\text {th }}, 25^{\text {th }}, 75^{\text {th }}$, and $90^{\text {th }}$ percentiles of the basal area on plots at different times during the simulation (treatment TN).
Grundytans variation mellan provytorna i början och slutet av simuleringsperioden (Behandling TN).


The diameter distributions with moderate ingrowth displayed a lower number of small trees and a more pronounced surplus of medium-sized trees than at the start of simulation (Figure 24). Only the maximum scenario demonstrated the potential after 25 years to maintain a similar stand structure as today. After 50 years, a lack of trees with DBH $<12 \mathrm{~cm}$ was indicated.
The final proportion of spruce was 43-60\% of trees, depending on assumed ingrowth. For trees with DBH < 20 cm , the spruce proportion was $5-6 \%$ higher.
Figure 24．Projected development of the diameter distribution in treatment TN according to ingrowth scenarios at the age of 25 and 50 years （based on 48 plots and stand age－independent single－tree age）．
Beräknad BHD－fördelningar efter 25 och 50 år enligt behandling TN med tre olika inväxtscenarierna．


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120
100
80
60
40
20
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Alternation of target diameters Cuttings with 5 cm higher target diameters (treatment T+5) resulted in higher stand densities than treatment T with original target diameters (ranging from 27 to $40 \mathrm{~m}^{2} /$ ha over the simulation period). Harvest removals were lower, but 20-25 years cutting intervals were similar to treatment T .
Cuttings according to 5 cm lower diameters reduced BA from approx. 35 to $15 \mathrm{~m}^{2} / \mathrm{ha}$, the cutting interval increased to 35 years (Figure 25). The MAI was 0.55 and $0.57 \mathrm{~m}^{2} / \mathrm{ha} \mathrm{BA} \mathrm{for}$ treatment T+5 and T-5. MAI in terms of volume resulted in $0.6 \mathrm{~m}^{3} /$ ha differences annually between Treatment T+5 and T-5 (Table 20 and 21).


Figure 25. Basal area development according to cutting with 5 cm higher (blue), 5 cm lower target diameters (red), and the original target diameters (black). Projections are based on stand age-independent single-tree ages and $100 \mathrm{~m}^{3} /$ ha minimum standing volume after cut. Grundytans utveckling enligt behandlingen $T$ (svart) i jämförelse med 5 cm högre (blå) och 5 cm lägre (röd) måldiameter.

With larger diameters, the variation of the $B A$ on plots increased somewhat during the simulation period (Figure 26), covering a large range just before cutting from 17 to $61 \mathrm{~m}^{2} / \mathrm{ha}$, but also after cutting from 7 to $53 \mathrm{~m}^{2} / \mathrm{ha}$. The diameter distributions in Figure 27 indicated a very small number of trees between $5-10 \mathrm{~cm}$ DBH after 50 years.

Table 20. Simulated initial stand parameters and after 50 years, with mean annual increment and harvest, according to treatment $\mathbf{T}$ with $\mathbf{5} \mathbf{~ c m}$ higher target diameters (stand ageindependent single-tree ages, moderate ingrowth). Behandling $T$ med 5 cm högre måldiameter (moderat inväxt) - initiala och simulerade beståndsparametrar efter 50 år.

| Species | start of simulation |  |  |  | MAI |  | Time of cutting [years] | $\begin{array}{\|l} \begin{array}{l} 1 \text { th cut } \end{array} \\ \text { removal } \end{array}$ | 2nd cut removal[m/ha] | $\left.\begin{array}{\|c\|} 3 \mathrm{rd} \text { cut } \\ \text { removal } \\ {[\mathrm{m} / \mathrm{ha}]} \end{array} \right\rvert\,$ | end of simulation |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} \mathrm{N} \\ {\left[\mathrm{ha}^{-1}\right]} \end{gathered}$ | $\left\lvert\, \begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{gathered}\right.$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\left[\begin{array}{c} \text { Vol } \\ {\left[\mathrm{m}^{3} / \mathrm{ha}\right]} \end{array}\right]$ | $\begin{gathered} \text { BA } \\ {\left[\mathrm{m}^{2} / \mathrm{ha}\right]} \end{gathered}$ | $\left\|\begin{array}{c} \mathrm{Vol} \\ {\left[\mathrm{~m}^{3} / \mathrm{ha}\right]} \end{array}\right\|$ |  |  |  |  | $\left[\begin{array}{c} \mathrm{N} \\ {\left[\mathrm{ha}^{-1}\right]} \end{array}\right]$ | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\begin{gathered} \mathrm{Vol} \\ {\left[\mathrm{~m}^{3} / \mathrm{ha}\right]} \end{gathered}$ |
| Spruce | 489 | 13.8 | 18.9 | 131 | 0.27 | 3.3 |  | 44 | 39 | 47 | 347 | 14.09 | 22.7 | 143 |
| Pine | 164 | 15.4 | 34.6 | 145 | 0.15 | 1.8 |  | 52 | 58 | 52 | 51 | 6.40 | 39.8 | 62 |
| Birch | 54 | 2.6 | 24.6 | 21 | 0.01 | 0.1 |  | 16 | 4 | 1 | 21 | 0.36 | 14.8 | 3 |
| Beech | 21 | 0.4 | 15.1 | 3 | 0.02 | 0.2 |  | 0 | 2 | 5 | 11 | 0.46 | 23.2 | 4 |
| Oak | 271 | 3.0 | 11.8 | 21 | 0.10 | 0.9 |  | 1 | 1 | 5 | 199 | 6.52 | 20.4 | 52 |
| Total | 999 | 35.2 | 21.2 | 322 | 0.55 | 6.2 | 5/30/50 | 114 | 105 | 110 | 629 | 27.83 | 23.7 | 265 |

Figure 26. Box plot with median, $10^{\text {th }}, 25^{\text {th }}, 75^{\text {th }}$, and $90^{\text {th }}$ percentiles of the basal area on plots at different times during the simulation (treatment $\mathrm{T}+5$ ). Grundytans variation mellan provytorna $i$ början och slutet av simuleringsperioden (Behandling T men med 5 cm högre måldiameter).




Figure 27. Initial and projected diameter distributions according to cuttings with 5 cm higher target diameters at the age of 0,25 and 50 years (moderate ingrowth). Beräknad BHD-
fördelningar efter 25 och 50 år enligt behandling $T$ med 5 cm högre måldiameter (moderat inväxning).

Table 21. Simulated initial stand parameters and after 50 years, with mean annual increment and harvest, according to treatment $\mathbf{T}$ with $\mathbf{5} \mathbf{~ c m}$ lower target diameters (stand ageindependent single-tree age, medium ingrowth scenario). Behandling $T$ med 5 cm lägre måldiameter (moderat inväxt) - initiala och simulerade beståndsparametrar efter 50 år.

| Species | start of simulation |  |  |  | MAI |  | Time of cutting [years] | 1th cut removal [m/ha] | 2nd cut removal [m/ha] | end of simulation |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{array}{\|c} \mathrm{N} \\ {\left[\mathrm{ha}^{-1}\right]} \\ \hline \end{array}$ | $\begin{gathered} B A \\ {\left[\mathrm{~m}^{2} / \mathrm{ha]}\right]} \\ \hline \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \\ \hline \end{gathered}$ | $\begin{gathered} \mathrm{Vol} \\ {\left[\mathrm{~m}^{3} / \mathrm{ha}\right]} \\ \hline \end{gathered}$ | $\begin{array}{\|c\|} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \end{array}$ | $\begin{array}{\|c} \mathrm{Vol} \\ {\left[\mathrm{~m}^{3} / \mathrm{ha}\right]} \end{array}$ |  |  |  | $\left\lvert\, \begin{gathered} \mathrm{N} \\ {\left[\mathrm{ha}^{-1}\right]} \end{gathered}\right.$ | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2} / \mathrm{ha}\right]} \\ \hline \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\left[\begin{array}{c} \mathrm{Vol} \\ {\left[\mathrm{~m}^{3} / \mathrm{ha}\right]} \end{array}\right.$ |
| Spruce | 401 | 7.2 | 15.1 | 59 | 0.32 | 3.4 |  | 69 | 85 | 551 | 13.45 | 17.6 | 124 |
| Pine | 71 | 4.8 | 29.4 | 40 | 0.07 | 0.8 |  | 104 | 67 | 9 | 0.97 | 36.1 | 9 |
| Birch | 22 | 0.2 | 10.6 | 1 | 0.01 | 0.1 |  | 20 | 1 | 55 | 0.42 | 9.8 | 3 |
| Beech | 19 | 0.3 | 14.0 | 2 | 0.02 | 0.2 |  | 1 | 7 | 8 | 0.31 | 22.8 | 3 |
| Oak | 271 | 2.9 | 11.8 | 21 | 0.15 | 1.2 |  | 0 | 19 | 195 | 7.31 | 21.9 | 57 |
| Total | 783 | 15.4 | 15.8 | 124 | 0.57 | 5.6 | 0/35 | 194 | 180 | 818 | 22.46 | 18.7 | 195 |

5 cm lower target diameters resulted in heavy removals of 179-212 $\mathrm{m}^{3} / \mathrm{ha}$ and low standing volumes of 124 and $122 \mathrm{~m}^{3} / \mathrm{ha}$ after cut (with moderate ingrowth). An important difference to treatment $T$ were longer harvest intervals of $35-40$ years (Table C21 in the Appendix). But, $B A$ and volume increased steadily after such heavy removals. Compared to treatment $T$, the MAI decreased from 5.9 to $5.6 \mathrm{~m}^{3} / \mathrm{ha}$ (Table C21).
The more or less bimodal diameter distribution after 50 years in Figure 29 points on a larger number of small trees than the original target diameter cutting. But, decreasing BA variation between plots can indicate more homogeneous stand structures too (Figure 28).

Figure 28. Box plot with median, $10^{\text {th }}, 25^{\text {th }}, 75^{\text {th }}$, and $90^{\text {th }}$ percentiles of the basal area on plots at different times during the simulation (treatment T-5). Grundytans variation mellanprovytorna i början och slutet av simuleringsperioden (Behandling T med 5 cm lägre måldiameter).



Figure 29. Initial and projected diameter distribution according to cuttings with 5 cm higher target diameters at the age of 0,25 and 50 years (moderate ingrowth scenario). Beräknad BHD-
fördelningar efter 25 och 50 år enligt behandling $T$ med 5 cm lägre måldiameter (moderat inväxning).

Projected tree species composition Current diameter distributions indicated already future changes from pine towards more spruce and oak in the overstory. The proportion of spruce trees increased in all treatments with moderate ingrowth. In the control, the proportion increased by $2.5 \%$ after 50 years (Table 22). Even in treatment TN, the spruce proportion increased by $2.3 \%$. Only according to the minimum scenario, its proportion decreased by $3.9 \%$. The maximum ingrowth scenario indicated an increase by $8.8 \%$. In treatment T and TS, an increase by 10.2 respective $6.4 \%$ was projected. In the treatments with alternated target diameters an increase by 5.2 and $16.1 \%$ spruce of total tree number was projected. In terms of BA, the proportion increased by $4.4 \%$ in the control treatment. An increase by 13.4 and $10.8 \%$ was simulated for treatment T and TS. In treatment TN, the BA proportion was projected relatively constant with $0.3 \%$. Alternated target diameters resulted in increases of 11.4 and $13.3 \%$.

Table 22. Projected changes of spruce proportion after 50 years depending on treatment and ingrowth scenario (no projections according to the minimum and maximum ingrowth scenario were made for treatment T+5 and T-5). Beräknade föründringar av granandel efter 50 år, beroende på behandling och inväxningsscenario ( $\mathrm{N} \%=$ relativ stamantal, BA\% = relativ andel av grundyta).

| Treat- <br> ment | $\operatorname{yyy}$ | N\% |  | BA\% |  |  |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: |
| C | +2.5 | +2.5 | +2.5 | +4.4 | +4.4 | +4.4 |
| T | +3.4 | +10.2 | +11.8 | +11.6 | +13.4 | +16.0 |
| TS | +14.1 | +6.4 | +8.3 | +13.9 | +10.8 | +15.2 |
| TN | -3.9 | +2.3 | +8.8 | -2.9 | +0.3 | +4.5 |
| $\mathrm{~T}+5$ | - | +5.2 | - | - | +11.4 | - |
| $\mathrm{T}-5$ | - | +16.1 | - | - | +13.3 | - |

## Estimated income according to simulation results

With 110-140 m3 harvest removals (equal to 88-112 m3 marketable wood) in 20-25 year intervals, a considerable positive income is possible. Constant future prices equal to the average from 2006-2011 (source: SÖDRA price list), no saw timber for broadleaves, timber quality 1 for pine stems classified in the stand as quality class 1, and pallet timber for crown compartments and pine trees with quality class 2 were assumed. Calculations were based on $125 \mathrm{SEK} / \mathrm{m}^{3}$ harvesting costs, equal to the first cutting. Under such conditions, about $30000 \pm$ 5000 SEK can be expected as net revenue at each harvest according to treatment T and TS. However, the gross income proportion of pine was $62 \%$ at the first harvest, and will decrease rapidly. $35 \%$ of the first income according to treatment $T$ came from pine logs with timber quality 1 . These high quality logs had a volume proportion of $24 \%$ of the marketable wood. No price differences between quality classes for spruce were assumed. Log lengths were calculated according to Nilsson \& Fahlvik (2006). Table 22-24 present the estimated income and costs according to these presumptions and a discount rate of $2 \%$, according to the three different treatments $\mathrm{T}, \mathrm{TS}$, and TN. Table 25 describes income and costs if the final felling would be clearcutting. (Results for 0 and $4 \%$ interest rates are shown in Appendix E.)

Table 22. Yield table for treatment T. Figures were calculated per ha. Interest rate 2\%. Produktionstabell för behandling T, värden per ha, ränta 2\%.

|  |  | Stand before thinning |  |  | Removed |  |  |  |  |  |  | Net |  | Annual |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Time |  | BA | V |  | BA | V | MAI |  | Income | Cost | income | PV | revenue | NPV |
|  | [ yrs ] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | [ $\mathrm{m}^{3}$ ] | Treatment | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] |
| Spruce | 0 | 489 | 13.8 | 129 | 50 | 4.8 | 52 |  |  |  |  |  |  |  |  |
| Pine | 0 | 164 | 15.4 | 144 | 55 | 6.9 | 69 |  |  |  |  |  |  |  |  |
| Broadleaves | 0 | 346 | 5.9 | 45 | 29 | 2.2 | 18 |  |  |  |  |  |  |  |  |
| Total | 0 | 999 | 35.2 | 318 | 134 | 13.9 | 140 |  | Harvest | 49518 | 14000 | 35518 | 35518 |  |  |
| Spruce | 25 | 459 | 15.6 | 152 | 61 | 4.2 | 45 |  |  |  |  |  |  |  |  |
| Pine | 25 | 103 | 11.8 | 113 | 44 | 6.0 | 60 |  |  |  |  |  |  |  |  |
| Broadleaves | 25 | 287 | 6.6 | 51 | 16 | 0.9 | 8 |  |  |  |  |  |  |  |  |
| Total | 25 | 849 | 34.1 | 316 | 121 | 11.0 | 113 | 6.1 | Harvest | 39845 | 11300 | 28545 | 17399 |  |  |
| Spruce | 45 | 423 | 16.2 | 163 | 62 | 4.7 | 53 |  |  |  |  |  |  |  |  |
| Pine | 45 | 57 | 7.3 | 71 | 36 | 5.1 | 51 |  |  |  |  |  |  |  |  |
| Broadleaves | 45 | 276 | 8.8 | 71 | 28 | 1.9 | 17 |  |  |  |  |  |  |  |  |
| Total | 45 | 756 | 32.3 | 305 | 126 | 11.7 | 121 | 5.9 | Harvest | 43290 | 12100 | 31190 | 12794 | 2185 | 65711 |

$\mathrm{N}=$ number of stands, $\mathrm{BA}=$ basal area, $\mathrm{V}=$ total volume over bark, $\mathrm{PV}=$ present value, $\mathrm{NPV}=$ net present value

Table 23. Yield table for treatment TS. Figures were calculated per ha. Interest rate 2\%.
Produktionstabell för behandling TS, värden per ha, ränta 2\%.

|  |  | Stand before thinning |  |  | Removed |  |  |  |  |  |  | Net |  | Annual | NPV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Time |  | BA | V |  | BA | V | MAI |  | Income | Cost | income | PV | revenue | year 0 |
|  | [ yrs ] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | [ $\mathrm{m}^{3}$ ] | Treatment | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] |
| Spruce | 0 | 489 | 13.8 | 129 | 50 | 4.8 | 53 |  |  |  |  |  |  |  |  |
| Pine | 0 | 164 | 15.4 | 144 | 55 | 6.9 | 69 |  |  |  |  |  |  |  |  |
| Broadleaves | 0 | 346 | 5.9 | 45 | 29 | 2.2 | 19 |  |  |  |  |  |  |  |  |
| Total | 0 | 999 | 35.2 | 318 | 134 | 13.9 | 140 |  | Harvest | 49518 | 14000 | 30518 | 30518 |  |  |
|  | 0 |  |  |  |  |  |  |  | Soil prep. |  | 2500 |  |  |  |  |
|  | 0 |  |  |  |  |  |  |  | Cleaning |  | 2500 |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 25 | 641 | 15.8 | 151 | 63 | 4.2 | 45 |  |  |  |  |  |  |  |  |
| Pine | 25 | 191 | 12.0 | 113 | 44 | 6.0 | 60 |  |  |  |  |  |  |  |  |
| Broadleaves | 25 | 307 | 6.7 | 51 | 17 | 0.9 | 8 |  |  |  |  |  |  |  |  |
| Total | 25 | 1139 | 34.5 | 315 | 124 | 11.1 | 113 | 6.0 | Harvest | 39845 | 11300 | 23545 | 14351 |  |  |
|  | 25 |  |  |  |  |  |  |  | Soil prep. |  | 2500 |  |  |  |  |
|  | 25 |  |  |  |  |  |  |  | Cleaning |  | 2500 |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 45 | 555 | 16.6 | 165 | 63 | 4.5 | 51 |  |  |  |  |  |  |  |  |
| Pine | 45 | 137 | 7.8 | 75 | 38 | 5.2 | 51 |  |  |  |  |  |  |  |  |
| Broadleaves | 45 | 283 | 8.9 | 71 | 28 | 1.9 | 18 |  |  |  |  |  |  |  |  |
| Total | 45 | 975 | 33.3 | 311 | 129 | 11.6 | 120 | 6.1 | Harvest | 43290 | 12000 | 26290 | 10784 | 1607 | 55653 |
|  | 45 |  |  |  |  |  |  |  | Soil prep. |  | 2500 |  |  |  |  |
|  | 45 |  |  |  |  |  |  |  | Cleaning |  | 2500 |  |  |  |  |

$\mathrm{N}=$ number of stands, $\mathrm{BA}=$ basal area, $\mathrm{V}=$ total volume over bark, $\mathrm{PV}=$ present value, $\mathrm{NPV}=$ net present value

Table 24. Yield table for treatment TN. Figures were calculated per ha. Interest rate 2\%.
Produktionstabell för behandling TS, värden per ha, ränta 2\%.

|  |  | Stand before thinning |  |  | Removed |  | V | MAI |  | Income | Cost | Net income | PV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Time |  | BA | V |  | BA |  |  |  |  |  |  |  |
|  | [yrs] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | [ $\mathrm{m}^{3}$ ] | Treatment | [SEK] | [SEK] | [SEK] | [SEK] |
| Spruce | 0 | 489 | 13.8 | 129 | 68 | 6.0 | 64 |  |  |  |  |  |  |
| Pine | 0 | 164 | 15.4 | 144 | 28 | 4.3 | 44 |  |  |  |  |  |  |
| Broadleaves | 0 | 346 | 5.9 | 45 | 17 | 1.6 | 13 |  |  |  |  |  |  |
| Total | 0 | 999 | 35.2 | 318 | 113 | 11.9 | 121 |  | Harvest | 30886 | 12100 | 18786 | 18786 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 20 | 436 | 12.7 | 119 | 84 | 5.1 | 53 |  |  |  |  |  |  |
| Pine | 20 | 131 | 14.4 | 136 | 42 | 6.0 | 60 |  |  |  |  |  |  |
| Broadleaves | 20 | 303 | 6.6 | 51 | 16 | 0.7 | 6 |  |  |  |  |  |  |
| Total | 20 | 870 | 33.7 | 306 | 142 | 11.7 | 119 | 6.0 | Harvest | 46843 | 11900 | 34943 | 23516 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 45 | 379 | 12.7 | 123 | 103 | 6.6 | 70 |  |  |  |  |  |  |
| Pine | 45 | 84 | 11.0 | 108 | 55 | 8.0 | 81 |  |  |  |  |  |  |
| Broadleaves | 45 | 280 | 9.7 | 80 | 11 | 0.4 | 4 |  |  |  |  |  |  |
| Total | 45 | 743 | 33.4 | 311 | 169 | 15.0 | 155 | 5.7 | Harvest | 59851 | 15500 | 44351 | 18193 |

$\mathrm{N}=$ number of stands, $\mathrm{BA}=$ basal area, $\mathrm{V}=$ total volume over bark, $\mathrm{PV}=$ present value, $\mathrm{NPV}=$ net present value

Table 25. Yield and cost projections for clearfelling, planting of spruce on this site, and typical even-aged forest management with 60 years rotation. Figures calculated per ha with the forest simulator DT (Nilsson \& Fahlvik, 2006). Interest rate 2\%.
Produktionstabell för kalavverkning och granplantering, värden per ha, ränta $2 \%$.

|  |  | Stand before thinning |  |  | Removed |  |  | MAI |  | Income | Cost | Net income | PV | Annual revenue | $\begin{gathered} \text { NPVV } \\ \text { (year 0) } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Time |  | BA | V |  | BA | V |  |  |  |  |  |  |  |  |
|  | [yrs.] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | [ $\mathrm{m}^{3}$ ] | Treatment | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] |
| Spruce | 0 | 489 | 13.8 | 129 | 489 | 13.8 | 129 |  |  | 32904 |  |  |  |  |  |
| Pine | 0 | 164 | 15.4 | 144 | 164 | 15.4 | 144 |  |  | 72079 |  |  |  |  |  |
| Broadleaves | 0 | 346 | 5.9 | 45 | 346 | 5.9 | 45 |  |  | 9223 |  |  |  |  |  |
| Total | 0 | 999 | 35.2 | 318 | 999 | 35.2 | 318 |  | Harvest | 114206 | 31800 | 64406 | 64406 |  |  |
|  | 0 |  |  |  |  |  |  |  | Soil prep. |  | 3000 |  |  |  |  |
|  | 0 |  |  |  |  |  |  |  | Planting |  | 15000 |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 7 |  |  |  |  |  |  |  | Cleaning |  | 1000 | -1000 | -871 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 15 |  |  |  |  |  |  |  | Cleaning |  | 1500 | -1500 | -1115 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 30 | 1783 | 27.5 | 165 | 796 | 10.7 | 64 | 5.5 | Thinning | 11954 | 8240 | 3714 | 2050 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 40 | 637 | 21.3 | 177 | 318 | 9.3 | 76 | 7.9 | Thinning | 19660 | 6251 | 13409 | 6073 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 60 | 573 | 40.4 | 454 | 573 | 40.4 | 454 | 9.9 | Harvest | 160584 | 15325 | 145259 | 44272 | 2665 | 114816 |

$\mathrm{N}=$ number of stands, $\mathrm{BA}=$ basal area, $\mathrm{V}=$ total volume over bark, $\mathrm{PV}=$ present value, $\mathrm{NPV}=$ net present value
The net income by partial felling was about 30000 SEK every 20-25 years compared to more than 100000 SEK every 60 years by clearfelling. Subsequent costs by site preparation, planting and pre-commercial thinning, as well as moderate income by thinning have to be taken into account too (Table 25).
Considering only today, the net income by target diameter cutting at year 0 is roughly $50 \%$ compared to clear-cutting.

The net present value (NPV) can consider longer time horizons, assuming the current costs and prices. Discounting the costs of the next 20 years to year 0 by an interest rate of $2 \%$, the NPV plus the initial income would amount to 62421 SEK for the clearfelling strategy and 30418 SEK for treatment TS. For treatment T, the NPV would amount to 35618 SEK. Choosing another time horizon of 40 years, the NPV plus the initial income would amount to 70544 SEK for clearfelling and 44769 SEK for treatment TS ( 53017 SEK for treatment T). For 60 years, the NPV plus the initial income by the second final harvest would amount to 114816 SEK for clearfelling, and 55553 SEK for treatment TS. Generally, these differences ranged from $48-63 \%$, depending on time period. Choosing another interest rate of $3 \%$, differences between clearfelling and treatment TS would range from 49 to 61\%. See Table E1-E8 in the appendix for 0 and $4 \%$ interest rates!

Cash flow calculations do not depend on interest rates. The income according to treatment TS in Table 23 resulted in an average net annual revenue of 1600 SEK according to (while treatment T and TN are equal to 2200 respective 2000 SEK annually). Following the clearfelling option, the net annual revenue over the 60 year period with income from two clearcuts amounted to 3738 SEK (or 2665 SEK if first clear-cut is neglected). Referring to cash flow, the profitability of target diameter cutting is about 43-59\% over the simulation period, compared to even-aged spruce management.

To even out different 50 and 60 years time periods and different years of considerable income, a net present value in perpetuity (for infinite harvest cycles) was calculated. In case of even-aged management, the value corresponds to the soil expectation value according to Faustmann's (1849) formula plus income of the first harvest. In that case, a continuous stand development with regular harvest intervals every 22 years could be assumed for treatment TS. In addition, a lower income due to the future lack of pine trees was assumed in Figure 30 (average price for pine was $520 \mathrm{SEK} / \mathrm{m} 3$ and for spruce was $375 \mathrm{SEK} / \mathrm{m} 3$ ). As a rough guess, 120-150 m3/ha total volume over bark, equal to $96-120 \mathrm{~m} 3$ marketable wood with an average price of $350-400$ SEK/m3 minus 125 SEK harvesting costs would provide 2160033000 SEK income. Subtracting costs for scarification and cleaning, the net income could be 16600-28000 SEK regularly every 22 years (Figure 30). Based on such sketches of the future, the NPV in perpetuity would vary between 68000-77000 SEK for treatment TS and 130000 SEK for continuous clearfelling and re-planting. Time lines with presumed incomes and costs are presented in Figure 30 and 31 (with light colors for assumptions outside the simulation period).


Figure 30. Hypothetical income when treatment TS would be continuously applied (sketch as basis for the calculation of the net present value for infinite harvest cycles).
förväntade teoretiska intäkter vid behandling TS med markberedning skulle resultera i kontinuerliga måldiameterhuggningar varje 22 år, beroende på hållbar utveckling av nya träd för att ersätta mogna träd (priser antas vara konstanta).


Figure 31. Hypothetical income according to even-aged spruce management (sketch as basis for a net present value calculation for infinite rotation cycles).
Förväntade teoretiska intäkter vid kalavverkning och plantering av gran, upprepat med en omloppstid på 60 år (priserna antas vara konstanta).

## Discussion

## Initial stand structure, canopy gaps and regeneration

## Gaps

Different sizes of gaps provided a wide spectrum of different environmental conditions. The range of light ratios to open-land conditions in the center of gaps was similar to the total range reported under shelterwood (Strand, Ottoson-Löfvenius, Bergsten, Lundmark \& Rosvall, 2006). In addition, lower light conditions at the edge of gaps and under canopy increased the spectrum compared to ordinary shelterwood. Hence, the first hypothesis of this study could not be falsified: The target diameter cutting did create an irregular pattern of gaps in different size classes with light conditions more diverse than in uniform shelterwoods. But, maximum radiation in the largest gaps did not exceed values obtained in shelterwoods (Strand et al., 2006).

## Stand characteristics

The major characteristic of this stand was the heterogeneous forest structure with trees in all sizes and different ages. A stem number of 1000 trees per ha would imply a high density in old even-aged stands, but this number is normal in uneven-aged forests. E.g. Lundqvist (1989) found 567-1624 trees/ha before and 333-980 trees/ha after the first cut. The BA of $35.2 \mathrm{~m}^{2} /$ ha represented the upper level reported for boreal single tree selection forests before first cut (Lundqvist, 1989). However, the variation of BA from 10 to $50 \mathrm{~m}^{2} / \mathrm{ha}$ on single plots indicated a large variability in stand density.
The variation of tree heights dependent on DBH was also larger than usually found in evenaged stands (i.e. Söderberg, 1992).

Single tree sizes The trees covered a large range of sizes from seedlings to mature trees at the time of the first target diameter cutting. The stand represented an old managed forest in an advanced stage of succession with heterogeneous tree species composition and forest structure. But, it did not represent an uneven-aged managed forest under equilibrium conditions where shade tolerant species tend to dominate (Lundqvist, 1989; Lähde, Laiho, Norokorpi \& Saksa, 2002; Lüpke, 2004). and where a large number of small trees can replace the largest trees after harvest. Too many trees were part of the upper stand layer. Although a slightly exponential decrease of tree numbers with increasing tree size was found, the decrease could be more pronounced with more trees below 10 cm DBH and less trees between $20-40 \mathrm{~cm}$ DBH. In combination with the regeneration results, the initial diameter distributions indicated a lack of trees below 7 cm DBH. Thus, a sustainable supply of new trees cannot be expected without newly established seedlings.
There are few examples of diameter distributions of boreal spruce selection forest under quasi-equilibrium conditions (where the initial stand structure is maintained for long periods). For instance Lundqvist (1989) demonstrated a relative tree number decline in 4 cm DBH classes from 20-35\% of total tree number in class $8.5-12.5 \mathrm{~cm}$, over $15-20 \%$ at 20 cm , to $5-10 \%$ at 30 cm , with largest trees about $40-50 \mathrm{~cm}$ DBH. In the study stand in Halland, $40 \%$ of trees were found in class $8-12 \mathrm{~cm}$, and $25 \%$ in class $18-22 \mathrm{~cm}$ (neglecting the 200 trees/ha below 8 cm DBH for the comparison). However it should be noted that Lundquist (1989) study is on spruce managed with selection cutting. The equilibrium conditions could be different in a mixed stand managed with target diameter cuttings (Cameron, 2007).

Standing volume Estimations of the optimum standing volume of uneven-aged boreal spruce forests range between $150-250 \mathrm{~m}^{3} /$ ha (Chrimes, 2004), indicating a steep decline of increment below $100 \mathrm{~m}^{3} / \mathrm{ha}$. This finding is in line with empirical yield studies in Scandinavia by Bøhmer (1957), Lundqvist (1989) and Andreassen (1994).
Other references from central European Plenterwald forests have to be used cautiously in Scandinavia, because tree growth and regeneration establishment are favored by climate (i.e. Kunstler, Albert, Courbaud, Lavergne, Thuiller, Vieilledent, Zimmermann \& Coomes, 2011). In addition, the tree species composition is different with fir, spruce and beech. However, these references can also help to set a frame of growth expectations for the relatively southern site of the study stand compared to the northern references mentioned above. While Mitscherlich (1952) concluded that 200-500 $\mathrm{m}^{3} / \mathrm{ha}$ standing volume would have no significant influence on growth, Schütz (1997) highlighted the importance of an optimum stand volume: Depending on defined target diameters, optimum stand volumes should amount to approx. $250 \mathrm{~m}^{3} / \mathrm{ha}$ for 60 cm , and $350 \mathrm{~m}^{3} / \mathrm{ha}$ for 110 cm target size, at least for the example region in Switzerland. Since target diameters in south Sweden are smaller, the optimum standing volume has to be lower than $250 \mathrm{~m}^{3} / \mathrm{ha}$. In-between the frame from $100-250 \mathrm{~m}^{3} /$ ha indicated in the literature, $150-200 \mathrm{~m}^{3} / \mathrm{ha}$ appear most reasonable as optimum standing volume, considering the large latitudinal differences of the reference forests.
Considering the study stand, the development can be characterized by a succession from old pine forest towards uneven-aged forest. Boreal single tree selection stands can only be considered as a long-term target for the stand development of this forest type with a continuous forest cover. Currently, $322 \mathrm{~m}^{3} / \mathrm{ha}$ standing volume appear high regarding the establishment of regeneration and optimum stand growth in the transformation stage. For instance, the volume of a 95 years old, pure pine stand on this site would amount to 340 $\mathrm{m}^{3} /$ ha (Persson, 1992).

The tree species composition highlighted the difference to uneven-aged forest under equilibrium conditions. Compared to spruce, the pine overstory with lower light interception is supposed to improve the conditions for tree growth and establishment. Even so, the relative number of smallest trees was lower in the study stand than in references to single tree selection stands.
With respect to the overstory composition, single spruce trees with DBH $>30 \mathrm{~cm}$ were found in all blocks or treatments, thus a scattered distribution can be assumed. The spruce proportion is increasing with smaller sizes, but remains relatively constant between $5-17 \mathrm{~cm}$ DBH. In these size classes, the proportion of oak is also high and could suggest to be an important tree species for the future. However, the estimated growth of oak on this site is very low (Carbonnier, 1975) and generally decreases under shelter (Noack, 2006). From this perspective, the small tree size can be seen as an effect of slow growth rates of oak trees established at similar time when spruce trees entered the stand. The different proportion of small oak and spruce trees in block 1 and 3 was found to be notable for future comparisons regarding ingrowth and growth, but the general patterns were regarded as similar for the overall stand analysis and simulations made within this study.

## Box 2. Enrichment planting

The study neglected enrichment planting. But principally, the planting of new trees in gaps can be the save option to achieve sufficient ingrowth! However, this option is more expansive (especially with subsequent protection measures against browsing). Group planting and single tree protection can cause more than the double amount of planting costs on large areas. Practical guidance (planting technique, cost estimations) are given by Hagner (2004). Again, shade-tolerant tree species are more favorable.

Regeneration results clearly demonstrated a lack of regeneration. This conclusion can be drawn from both regeneration surveys, in the stand and in gaps. A usual size of total sample area for natural regeneration uneven-aged stands was used in the systematic stand survey (Nilson and Lundqvist, 2001; Staupendahl, 1997) to deliver representative results for the whole stand. The accuracy of the regeneration density in gaps may be somewhat lower, because the area of many gaps was estimated with the shape of an ellipse (Runkle, 1992) and the survey area was very large in the largest gaps. Such large single survey areas lead to the exclusion of the lowest height class from 10 to 20 cm after the field inventory to assure adequate accuracy. Considering potential error sources, the density in gaps was still twice as high as in the whole stand ( 385 versus 1003 individuals $\geq 20 \mathrm{~cm}$ per ha). Eventually, new gaps were more frequently established around old gaps (as the 49 gaps without any recently cut tree indicated).
The dominance of spruce in the understory under the absence of large-scale disturbances was demonstrated within shelterwood experiments (Nilsson et al., 2002), single-tree selection experiments (Lundqvist, 1989, Lähde et al., 2002) and by observations in pristine forests (Leemans, 1991; Linder, 1998; Shorohova et al., 2009)). The competiveness of the species (e.g. due to small susceptibility against browsing), was also confirmed by this study.

[^1]The highest variation in individual numbers occurs during the initial regeneration stage. As mentioned in Box 3 , the occurrence of natural regeneration is hardly predictable. Pukkala (1987) described driving forces as seed crops, stockable area, proportion of full seeds, and germination. Seed production is not easy to predict, depending on weather conditions and its influence on seed years (Ek \& Monserud, 1974). The spatial distribution of parent trees is also important (Kolström \& Pukkala, 1992; Huth, 2009), especially for oak and beech in this case.
Regarding seedlings, Saksa \& Valkonen (2011) could show a negative effect of local BA (determined on a 5 m -radius plot) on the number of $11-130 \mathrm{~cm}$ high spruce plants, while no correlation was found for very young seedlings with height < 11 cm . In conclusion, BA appeared very limited to reflect local growth conditions of smallest seedlings, but its influence is expected to increase with larger tree sizes. The latter was also suggested by Granhus (2001). After establishment, there are still many factors which can determine the success of regeneration, in particular water deficiencies, light, competition with ground vegetation, or herbivores in later stages (see Jäderlund et al. 1997; Brang, 1998).

Natural regeneration under different shelter-conditions To enhance natural regeneration in three multi-layered spruce stands, Granhus, Hanssen \& Chantal (2008) manipulated the canopy cover in combination with soil preparation by creating a 0.25 ha canopy openings and three shelterwood units (with BA from 11 to $20 \mathrm{~m}^{2} / \mathrm{ha}$ ) in Norway, resulting in 15-30 seedlings per site prepared spot (two years after a rich mast). He found high mortality rates of $20-30 \%$ during the $1^{\text {st }}$ winter and $2^{\text {nd }}$ summer after germination.
Saksa (2004) found 6000-25000 seedlings per ha (11-130 cm high) in five uneven-aged forest stands in southern Finland 5-10 years after single tree selection cutting: In average, 200 small trees/ha past the 130 cm tree height threshold annually, but the numbers varied largely from 10 to 1000 between stands. Kolström (1993) observed seedling densities from 100 to 26.500 individuals/ha (0.1-4 m in height).
For spruce plenter forests in Sweden, Lundqvist (1995) estimated approx. 50-100 stems/ha passing 1.3 m height, assuming annual mortality rates about $5 \%$ for seedlings $<1.3 \mathrm{~m}$. Holgén \& Hånell (2000), Örlander \& Karlsson (2000), Glöde (2002), Nilsson et al. (2002), Karlsson \& Nilsson (2005), and Nilsson et al. (2006) described about 1000-50000 established seedlings per ha 5-10 years after shelterwood cutting, displaying highly variable tree numbers. Additionally, positive effects of scarification on seedling occurrence and of shelterwood density on survival were demonstrated. Concluding from the studies in this section, a large natural variation of seedling occurrence is possible.
Beside seedling numbers, seedling sizes and height growth are important. The minimum ingrowth scenario assumed 11 cm annual height growth for spruce as found for the advanced regeneration before cutting. This value is comparable to average height shoots reported from a study in central Sweden seven years after thinning from above ( $30 \%$ and $60 \%$ removals): Nilson \& Lundqvist (2001) referred to $0.5-2 \mathrm{~m}$ high regeneration in a multilayered spruce stand that increased annual height increment from $3-4$ to $10-12 \mathrm{~cm}$. Chrimes \& Nilson (2005) demonstrated that height increment is better correlated to canopy openness than to BA. In the average, they reported a height increment of 25 mm for seedlings ( $0.1-0.5 \mathrm{~m}$ ), of 50 mm for "saplings" ( $0.5-2 \mathrm{~m}$ ), and of 75 mm for "small trees" from 2 m height to 5 cm DBH. Both, Nilson \& Lundqvist (2001) and Chrimes \& Nilson (2005), calculated average height increment of individuals, while the height increment in this study
refers to the tallest seedlings. Dominant seedlings in gaps were considered to represent future new trees better than the average (e.g. Petritan \& Lüpke, 2009). Glöde (2002) studied the height growth of spruce regeneration ( 1.1 m mean height of seedlings, 2.1 m mean height of dominant seedlings) after shelterwood removal in Sweden. The height shoots increased from approx. $8 \mathrm{~cm} /$ a before cutting to more than $20 \mathrm{~cm} / \mathrm{a}$ six years after cut (with a depression the first two years). Dominant seedlings increased from approx. 15 cm to more than 30 cm per year. No reduction of seedling numbers could be detected.
During the first five years after shelterwood cutting, less than 5 cm mean annual height increment was found in naturally regenerated spruce in southern Sweden (Nilsson et al., 2002; Nilsson et al., 2006). Örlander \& Karlsson (2000) could show for one shelterwood experiment ( $80-160$ trees $/ \mathrm{ha}$ ) that the mean accumulated height increment increased from $5.8 \mathrm{~cm} / \mathrm{a}$ (for seedlings < 20 cm height), over $13.7 \mathrm{~cm} / \mathrm{a}$ (20-50 cm height) and $21.4 \mathrm{~cm} /$ ( $50-$ 100 cm ), to 30.8 cm ( $>100 \mathrm{~cm}$ ). But, high shelterwood densities ( 320 trees/ha) resulted in less than 5 cm annual height increment for seedlings $<1 \mathrm{~m}$.
Summarizing these studies, about 10 and 30 cm annual height growth in gaps can represent well the arbitrary thresholds chosen for the simulation.

Mortality of small trees From literature, 2-5\% annual mortality can be estimated for 10-50 cm high spruce seedlings in uneven-aged stands (Lundqvist, 1995; Nilson \& Lundqvist, 2001; Eerikäinen, 2007). For $50-200 \mathrm{~cm}$ tall spruce plants, Nilson \& Lundqvist (2001) found 0-4\% rates. Mortality under spruce shelterwood without scarification was reported by Nilsson et al. (2002) with about $20 \%$ annually for the first five years (about $50 \%$ during the first year after germination). Under pine shelter, $25 \%$ mortality were found for a 5 years period (Nilsson et al., 2006), equal to $6 \%$ annually in average. From these figures, mortality rates assumed in the minimum scenario appear two times higher. No mortality according to the maximum scenario was likely an underestimation, while the medium scenario reflected well these mortality rates.
Nevertheless, mortality rates assumed in this study stand are a rough estimation. Climate change and micro-site dependent conditions increase the uncertainty of future trees recruitment. But, mortality rates estimated from literature and medium scenario mortality were similar. Oscillating values for beech had to be smoothed. No references to the mortality of birch or trees with 2-5 cm DBH under shelter conditions were indicated in literature. But, tree mortality can generally assumed to decrease with increasing tree size (i.e. Lundqvist, 1995; Eid \& Tuhus, 2001; Juknys et al., 2006).

Stand characteristics to describe ingrowth According to Schütz (2001), higher standing volume has a negative effect on tree recruitment. Reversely, regeneration processes determine management options and stand density in the long run (Tahvonen et al., 2010). However, most experiments demonstrate weak correlations between stand BA and regeneration growth or occurrence (e.g. Lundqvist \& Fridman, 1996; Bachofen, 1999; Nilson \& Lundqvist, 2001; Chrimes, 2004). Beside the higher number of seedlings in gaps, no effect of different volume levels on regeneration was found by this study due to the low number of seedlings in the stand survey.
Chrimes (2004) found a stronger influence of canopy openness on the height growth of seedlings and Kuusipalo (1985) described a relationship between BA and canopy openness (where BA explained $63 \%$ of the variation of canopy openness; BA and the proportion of
spruce could explain 75\%). Under pine shelter, Strand et al. (2006) revealed stronger correlations with the distance to the nearest tree than with light conditions. All three factors, BA, light and distance to the nearest canopy tree (associated with root competition e.g.) are altered by creating gaps. Therefore, improved growth conditions for advanced regeneration can be expected after cutting. In addition, gaps are considered as important feature to initiate and promote regeneration according to the concept of a natural forest cycle (e.g. Leeman, 1991; Liu \& Hytteborn, 1991; Dai, 1996).
However, two main questions remain: 1 . When the new regeneration will be established in a particular stand? 2. How many individuals can be expected to grow into the tree layer within a certain time period? The final answer in this case study can be obtained by future measurements only. See also Lundqvist (2012) regarding regeneration and ingrowth estimates in heterogeneously structured forest!

Ingrowth rates Based on data from re-visited permanent sample plots of the Swedish National forest inventory, regression models to estimate the future ingrowth of spruce, pine and birch were developed by Wikberg (2004). According to this model, spruce saplings had the highest probability to grow taller in the understory (Wikberg, 2004). This is in line with the forest ecological theory that spruce is most competitive if fire and other large-scaled disturbances are excluded (e.g. Engelmark \& Hytteborn, 1999; Shorohova et al., 2009). Similar conclusions can be drawn from empirical studies (e.g. Lundqvist, 1989; Liu \& Hytteborn, 1991; Hofgaard, 1993).
Across all types of current forests in Sweden (excluding very young stands), Wikberg (2004) calculated a general probability of $11 \%$ for spruce saplings to pass 40 mm DBH after 5 -years. The ingrowth decreased with increasing stand density and age, and increased with increasing site index. The ingrowth rate ranged between 5 and $30 \%$. However, the inventory data used for the model did not refer explicitly to silvicultural measures aiming for regeneration. Contrary, recent thinning reduced the ingrowth probability due to harvest damages. Specifically for the study stand in Halland, 2-4 new trees per ha and year were projected. A reduction of BA to $15 \mathrm{~m}^{2} /$ ha was forecasted to give annually 13 new trees/ha, although this is an extrapolation from the validated model due to the small proportion of such forests.
Pukkala et al.(2009) projected approx. 2-7 new trees/ha per year for heterogeneously structured forests in Finland with $20-35 \mathrm{~m}^{2}$ BA per ha. For stands with $10 \mathrm{~m}^{2} / \mathrm{ha}$ BA, minimum 5 new trees (total tree number 250 trees/ha), and maximum 40-80 new trees ( 1800 trees/ha) were estimated.

There are some northern European case studies describing observed ingrowth in unevenaged spruce forests managed by single-tree selection (see also Lundqvist, 2012). Lundqvist (1989) recorded about 4-14 trees per ha and year that past an 8.5 cm DBH threshold in eleven spruce plenter forest stands in Northern and Central Sweden. However, $40 \%$ of trees below 8.5 cm DBH were removed at the first cutting in six of the eleven stands. Lähde et al. (2002) documented ingrowth in 23 heterogeneously structured, spruce dominated stands across Finland. He found 170 seedlings ( $<1.3 \mathrm{~m}$ ) per ha growing to saplings ( 1.3 m height to $9 \mathrm{~cm} \mathrm{DBH})$, and 80 saplings/ha passing the 9 cm threshold during 7-14 years monitoring periods after single-tree selection cutting, which is equal to 5-12 trees/ha annually. In average, about 10 trees per ha and year seem to be a reasonable ingrowth in uneven-aged boreal forests.

Another reference of ingrowth from more southern latitudes was provided by Tremer (2008) for the forest district Sellhorn in northern Germany. There, the forest sites are characterized by sandy soils with poor or medium nutrient supply (which is more similar to Swedish conditions than other sites in Germany). Two regeneration inventories in a 7 years-interval were carried out on 869 sample plots in the whole forest district. The stands were evenaged. Dominant tree species were pine ( $66 \%$ of BA of the district) and spruce ( $25 \%$ ). Beech and oak comprised $5 \%$, other broadleaves $4 \%$ of BA. (No soil preparation or underplanting was conducted.) The largest proportion of ingrowth was found for spruce, equivalent to 13 trees per year and ha (crossing a threshold of 7 cm DBH). However, on many plots no ingrowth was observed. A correlation between ingrowth and crown cover pointed out a linear increase from roughly 7 to 30 trees per ha annually, when crown cover increased from zero to $4.000 \mathrm{~m}^{2} / \mathrm{ha}$, and a linear decrease from 30 new trees/ha to zero when crown cover decreased from 4.000 to $12.500 \mathrm{~m}^{2} / \mathrm{ha}$. No direct correclation between BA and ingrowth was found (Tremer, 2008).
Considering only the plots where spruce regeneration was found at the first inventory, 26 trees per year and ha were registered in average. Comparing with our study stand, only on 20\% of regeneration plots occurred seedlings.
Pooling the studies in Table 26 to one figure, about 10 trees/ha seem to be a reasonable annual ingrowth, although natural variation can be great. Under proper conditions, such an average tree number is supposed to replace $3-4$ pole trees (with $10-20 \mathrm{~cm}$ DBH) which are needed to replace a harvested tree in a single tree selection forest (Schütz, 2001).

Regarding our study stand, with special regard to the current regeneration state, ten new trees per year and ha appear too high for the next 25 years. Before cutting, one tree per year and ha growing into a 5 cm size class seemed more likely according to regeneration characteristics. But, changing conditions after cutting are supposed to accelerate the rate.
Table 26. Regeneration and ingrowth studies in two- or multi-layered spruce-dominated forests in Northern Europe

| tree species | tree size [cm] | tree age [a] | seedling density [ $\mathrm{N} / \mathrm{ha}$ ] | annual mortality [\%] | height <br> growth <br> [cm/a] |  | threshold of ingrowth [cm] | forest type | study type | author |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| spruce |  | 0.5-1.5 |  | ca. 30 |  |  |  | shelterwood | observation | Granhus et al. (2008) |
| spruce | 10-50 |  | 1-176 | 2-7 |  | $<20$ | height 10 | uneven-aged | observation | Nilson and Lundqvist (2001) |
| spruce | 10-50 |  |  |  | 2.5 |  |  | uneven-aged | observation | Chrimes and Nilson (2005) |
| multiple | 11-130 |  | 6000-25000 |  |  | 200 | height 130 | uneven-aged | observation | Saksa (2004) |
| spruce | <130 |  |  | 5 |  | 50-100 | height 130 | uneven-aged | observation | Lundqvist (1995) |
| multiple | 97-148 |  | 41000 | 6-12 | 14-30 |  |  | shelterwood | observation | Holgén and Hånell (2000) |
| spruce | Hmean 110 |  | 24000-38000 |  | 8-20 |  |  | shelterwood | observation | Glöde (2002) |
| spruce | Hdom 210 |  |  |  | 15-30 |  |  | shelterwood | observation | Glöde (2002) |
| spruce | 50-200 |  | 6-164 | 0-4 | 3-12 |  |  | uneven-aged | observation | Nilson and Lundqvist (2001) |
| spruce | 50-200 |  |  |  | 5 |  |  | multi-storied | observation | Chrimes and Nilson (2005) |
| spruce | probably 0-400 |  | 81000 | 0-80 | 5.8-30.8 |  |  | shelterwood | observation | Örlander and Karlsson (2000) |
| multiple | 10-400 |  | 100-26500 |  |  |  |  | uneven-aged | observation | Kolström (1993) |
| spruce | 10-400 | average 5 | 2095 | ca. 5 |  |  |  | uneven-aged | modelling | Eerikäinen (2007) |
| spruce | 200-5 cm dbh |  | 67-688 |  | 7.5 |  |  | multi-storied | observation | Chrimes and Nilson (2005) |
| spruce |  |  |  |  |  | 1-6\% of saplings | dbh 4 | overall | modelling | Wikberg (2004) |
| spruce |  |  |  |  |  | 2-4\% of saplings | dbh 4 | stand-specific | modelling | Wikberg (2004) |
| spruce |  |  |  |  |  | approx. 2-7 | dbh 5 | "uneven-sized" <br> (BA $\left.20-35 \mathrm{~m}^{2} / \mathrm{ha}\right)$ | modelling | Pukkala et al. (2009) |
| spruce |  |  |  |  |  | 4-14 | dbh 8.5 | uneven-aged | observation | Lundqvist (1989) |
| multiple |  |  |  |  |  | 5-12 | dbh 9 | multi-storied | observation | Lähde et al. (2002) |
| spruce |  |  |  |  |  | 13 | dbh 7 | even-aged | observation | Tremer (2008) |

Comparison of ingrowth derived from literature and from advanced regeneration The study revealed differences between ingrowth rates estimated from advanced regeneration and rates concluded from literature. The stand-specific estimations fit well to Wikberg's (2004) ingrowth model, but represented rates below the average reported in uneven-aged forest. Ingrowth rates can be expected to increase after 15-35 years due to initiated regeneration by reduced stand density and gaps, according to the references from multi-layered managed forests (Pukkala et al., 2009) and shelterwoods (Nilsson et al., 2002).
Regarding maximum stem number of spruce regeneration after scarification, much denser stands than 2000 trees/ha can occur, when trees reach 5 cm DBH. For instance, Falck \& Rydberg (1992) and Petterson (1992) documented extreme dense pole stands with more than 10000 trees/ha under shelter and in plantations. But including cleaning as additional measure, 2000 trees/ha were assumed to achieve a sufficient single tree stability and tree growth according to findings by Rumpf \& Dittges (2008). Very dense spruce regeneration can develop considerably thinner and less stable dominant trees (Petterson, 1992).

## Future stand growth, diameter distributions and simulation constraints

Model data base The simulation used models based on permanent sample plots of the National forest inventory. The model to estimate the specific age of single trees was derived from 14.870 single tree observations on sample plots classified as uneven-aged (Elfving, 2003). Due to the definition of uneven-aged by the forest inventory, plots with advanced regeneration in old even-aged forest could be over-represented. In addition, Lundström (2008) found more plots suitable for single tree selection in northern Sweden. No specific silvicultural measures to promote single trees were recorded by the inventory. In addition, about 300.000 ha of old pine-spruce forest were determined in southern Sweden (Drössler, 2010), where a notable proportion of stands with two or more heightlayers can be assumed.

## Comparison with observations and other growth simulations

The plausibility of the BA simulation results was confirmed by two additional growth models (Ekö, 1985; Elfving, 2004), demonstrated in Figure 32 for the control. Both models are based on stand age. For a first validation, the BA development according to Elfving (2004) with 60 and 80 years stand age deviated little from the assumed 70 years ( $0.8-2.4 \%$ difference after 50 years without cutting). But, final growth validation with observed growth after five or more years is still necessary.
The simulation of stand basal area growth according to Ekö (1985) was applied to a model stand on the same site with similar tree species proportions (divided in a 95 years old overstory of pine and birch and in an understory of 60 years old spruce and 50 years old beech and oak). The model by Ekö (1985) indicated a more pronounced growth regression after some decades. The growth projection resulted for instance in $1.8 \mathrm{~m}^{2} /$ ha lower BA after 50 years within the control treatment compared to the presented simulation results. Regarding the stand age-dependent and -independent functions by Elfving (2004), slight differences were revealed. For the total simulation period, similar growth rates were reflected (Figure 32). All three models projected similar BA growth for the control with values ranging from 52.5 to $54.4 \mathrm{~m}^{2} / \mathrm{ha} \mathrm{BA}$ at the end of the simulation period.

The age-dependent growth model by Elfving (2004) was compared with 15 years observation in an uneven-aged spruce stand in Central Sweden, managed by single-tree selection (Lundqvist, Chrimes, Elfving, Mörling \& Valinger, 2007). Initial diameter distributions were characterized by a large number of smallest trees (in average 500 trees/ha with $5-10 \mathrm{~cm}$ DBH), a pronounced exponential decrease of tree numbers, and about 50 cm DBH maximum. The model overestimated the BA growth on stand level by $11 \%$ for thinning from above with BA removals of $33 \%$ and $50 \%$ (Elfving, 2009). However, several large trees died due to wind damage. For the control, predictions were $2 \%$ lower than observed. For small trees with 5-10 cm DBH, the observed values were 20-50\% higher than expected from the model (Elfving, 2009). Although the effect on total BA growth is small for short- or medium-term periods, a critical question is how correct the growth of smaller trees after release is reflected by the model. The few observations and the data base of the model do not rule out possibilities that suppressed trees react stronger after silvicultural releases than predicted.
From observation in managed, even-aged stands over several decades, an underestimation of growth of largest trees, and an overestimation of growth of smallest trees was revealed (Fahlvik et al. 2012).
Another test for validation of the age-dependent model by Elfving (2004) was made with 916 years observation in pristine forest reserves. The different forest types of the study plots were described by Linder (1998). Here, the increment was underestimated by $5 \% .2 / 3$ of the large variation could be explained by the model (Elfving, 2006a).
Furthermore, the age-dependent model was used for simulations by Elfving (2006b) to compare wood production of even-aged and uneven-aged stands. Considering all spruce sites in Sweden, Elfving (2006b) estimated for uneven-aged spruce stands $15 \%$ less growth compared to even-aged stands, which was also concluded from single-tree selection experiments in uneven-aged coastal spruce stands in Norway (Andreassen, 1994).
Another simulation study in Sweden, was carried out by Wikström (2008) who compared the growth of single-tree selection stands with even-aged forest by using one BA growth function for all tree species according to Elfving (2005). This function was calibrated for even-aged stands and projected growth rates similar to even-aged stands.

No increase of growth was indicated during the simulation period. The volume increment was higher in unmanaged forest than in managed forest (both totally and during the last decade). This is contrary to the BA projections and indicates wrong volume estimations. See Table 16 and 17 for instance, where $6.7 \mathrm{~m}^{3} / \mathrm{ha} \mathrm{MAI} \mathrm{in} \mathrm{volume} \mathrm{was} \mathrm{calculated} \mathrm{for} \mathrm{unmanaged}$ forest and $5.9 \mathrm{~m}^{3} / \mathrm{ha}$ for managed forest, while $0.53 \mathrm{~m}^{2} / \mathrm{ha} \mathrm{MAI} \mathrm{in} \mathrm{BA} \mathrm{was} \mathrm{projected} \mathrm{for}$ unmanaged forest and $0.56 \mathrm{~m}^{2} /$ ha for managed forest. Therefore, estimations of volume are less reliable in multi-layered forest than BA.


Figure 32. Simulation of BA development of the control treatment according to different growth models (stand age-dependent and -independent estimations of single tree ages, and stand-level BA growth). Simuleringar av grundytans tillväxt i kontrollbehandlingen med olika tillväxtmodeller.

Regarding the original treatments with forest management, BA levels before and after cut remained rather constant during the 50 years period. No depletion of the forest was indicated by the simulations of different types of target diameter cutting. The increment did not increase over time, as it might be expected by an increasing proportion of spruce from 40 to $55-60 \%$ of $B A$. In fact, the annual BA increment decreased from 0.6 to $0.5 \mathrm{~m}^{2} /$ ha over the simulation period, eventually pointing on the influence of single tree ages.
Comparing with a spruce plantation on the same site, the MAI projections for volume were very low. Only $60 \%$ of the $9.9 \mathrm{~m}^{3} \mathrm{MAI}$ which was projected for a 60 -years rotation. Similar ratios were found for treatment TS (61\%) and TN (58\%). If real volume growth rates after target diameter cutting would be more similar to projections in the control, wood production would be $68 \%$ compared to the DT-projection of an even-aged spruce stand. The site potential (= productive capacity of the site according to optimum management excluding risks) was estimated even higher: $10.4 \mathrm{~m}^{3}$ per ha and year for spruce according to Hägglund \& Lundmark (1982).
More reliably, $64 \%$ BA growth was estimated for treatment T, compared to $0.88 \mathrm{~m}^{2} / \mathrm{ha}$ MAI by the DT simulator. For treatment TS, $67 \%$ of the BA growth of an even-aged spruce stand was calculated.

Ingrowth estimations A sufficient numbers of new trees with sustainable tree supply can only be expected under treatment TS, after scarification. Under the minimum and intermediate scenario, treatment T and TN cannot be judged as sustainable if future tree harvests in reasonable time intervals should take place. The intermediate ingrowth scenario pointed on harvest intervals of several decades.
Such expectations are in line with $\varnothing$ yen \& Nilsen (2004) who were not satisfied with the recruitment of new trees after selective cutting in irregularly structured stands in Norway, but also with Kint et al. (2006) and Cameron \& Hand (2010). The two latter studies described stands with comparable tree species composition and provide background for forest transformation apart from modeling, but from a perspective with empirical observations and the expectations from the science of vegetation dynamics. They expect more than 50 years to develop sustainably heterogeneous forest structures.

Box 4. Discussion of technical ingrowth implementation in this study The goal within the simulation was to implement ingrowth simply, but logically from silvicultural perspective. Therefore, an approach was chosen with basal area levels that favor the establishment and growth of small trees (Wikberg, 2004). The simplification from crown cover to basal area was supported by estimates with open canopy on sample plots. Table B1 in Appendix B shows the relation between canopy openness and low BA. Kuusipalo (1985) and Chrimes (2004) found similar relationships. In addition, the proportion of plots with BA below $15 \mathrm{~m}^{2} /$ ha was similar to the proportion of gaps. Advantage of this approach was a description by silviculturally controllable stand characteristics, although many other important factors were neglected.
As follow-up consideration of our analysis, a separate implementation of ingrowth in 5-year time steps would probably result in more accurate estimations. Eventually, regeneration numbers in Table 15 might increase more in the last decades. However, another goal was to keep the ingrowth scenarios understandable by assuming 1,4 and 10 new trees per ha and year, including the state of regeneration today. Still, there is potential that future tree numbers will be lower or exceed the extreme scenarios (particularly if more ingrowth would occur under closed canopy than currently counted). However, the applied maximum scenario assumes two times more ingrowth than estimated for regeneration with 30 cm height growth for the next 15 years. Considering rates described in literature and the snapshot of current regeneration, annual ingrowth larger than five trees per ha under closed canopy and more than 40 trees per ha in gaps seems rather unlikely. Certainly, such hypothetical thinking has to be tested by remeasurements of the regeneration in 10 or 20 years.

Box 5. Possible sources of error when predicting stand basal area, volume and single tree growth Wrong measurements of empirical data, statistical bias, wrong assumptions in the model constructed, and extrapolation/misinterpretation by the user are potential sources for wrong predictions. The error due to basal area measurements is considered to be very small, because it amounts usually to roughly 1\% (Kramer \& Akca, 1995) and was carried out by the professional stuff of the experimental forest.
A considerable bias can be expected due to differences between the empirical data base used to build the growth model and the specific stand under investigation. An open question is the proportion of unevenaged single tree selection and of old pine stands with naturally regenerated spruce trees in the empirical data, both possible according to the definition of uneven-aged forest in the NFI. In addition, the response of suppressed trees after silvicultural release cannot be described properly by the model because of the lack of such detailed information in the inventory data. More generally, structural indices or spatial competition indices related to neighboring trees become more relevant with increasing heterogeneity of the stand structure. Still, distance-dependent growth simulators are also rather limited to describe the growth of single trees under more extreme silvicultural single-tree treatments for several decades (Yue et al. 2008, Albrecht et al. 2009, Mette et al. 2009, Albrecht et al. 2011, Albrecht et al. 2012).
The comparison of the model projection with observations in single-tree selection forest revealed a slight underestimation for the control treatment, and an overestimation for managed forest which is difficult to judge due to the interference with storm (Elfving, 2004). According to Elfving (2004), storm damage could explain the overestimation. If growth rates were biased by $2 \%$, then the final BA values of the control after 50 years would differ $0.8 \%$. After 25 years, BA values before cutting would be $0.7 \%$ different. Assuming $5 \%$ bias as a worst case, BA values would differ 2-3\% after 50 years undisturbed growth. After 25 years, before cutting, the values would differ by $1.7 \%$ from the predictions. Differences in total BA growth due to different stand structure on blocks amounted to $2.5 \%$ after 50 years. Mathematically, the bias vanishes when comparing to the model projections with each other. Comparing the basal area growth predictions of different treatments for the next 50 years, an error of $\pm 5 \%$ is roughly assumed by the authors. Differences between the predictions and actual future growth are expected to be larger (as predicted growth for evenaged forest already may differ 10-15\% from observed growth ion different regions of Sweden).

Form heights cause an additional error when calculating stand volume, in average $\pm 6 \%$ (Kramer \& Akca, 1995). Volume functions by Ekö (1985) are probably less accurate then Söderberg's (1992) functions for instance, but are more robust in our model because volume estimations are based on rather correct basal area estimations every five years. Both types of volume functions were not validated for multi-layered stands. In even-aged stands, the volume functions by Ekö (1985) differ by 0-4\%. 11\% higher volume growth rates in the unmanaged treatment, while basal area growth was slightly higher in managed treatments indicates even larger errors (i.e., the forecast for treatment C was only $6.5 \mathrm{~m}^{3} / \mathrm{ha} \mathrm{MAI} \mathrm{and} \mathrm{for} \mathrm{T} 5.5 \mathrm{~m}^{3} / \mathrm{ha}$, while BA was 0.53 and $0.56 \mathrm{~m}^{2} / \mathrm{ha} \mathrm{MAI)}. \mathrm{Therefore} \mathrm{roughly} 15 \$,$% errors in estimation of total volume$ production after 50 years is assumed by the authors. Effects of storm, insects or climate change were not considered within this simulation study. Finally, there is no empirical material to evaluate predicted growth of oak under such heterogeneous forest conditions.

Regarding the diameter development of a particular tree in uneven-aged forest, the few observations indicated very high variation between trees (Elfving, 2006a). Although average growth was reflected well and two thirds of the variation could be explained, single tree growth varied up to 50 or $200 \%$. Assuming a projected annual diameter growth of 0.2 cm , this tree could also grow 0.1 or 0.4 cm per year. Larger growth responses might particularly occur after silvicultural releases. In extreme cases, a particular tree with initially 20 cm DBH could grow to 25 or 40 cm instead of 30 cm after 50 years.
Considering such a large variation, diameter distributions after 25 years should be interpreted by summarizing two size classes to one 4 cm diameter class. After 50 years, 8 cm diameter classes are recommended for interpretation. However, despite the uncertainty, the simulated diameters were used to determine tree removals according to target diameters.
Finally, an underestimated growth of small trees was indicated by a study in northern Sweden which revealed 20-50\% larger growth rates of small trees with 5-10 cm DBH than predicted (Elfving, 2009). According to these rates, newly ingrown trees with 5 cm DBH would not grow to 9 cm after 25 years, but to $10-11 \mathrm{~cm}$ DBH.

Projection of future diameter distributions The diameter distributions calculated in 50 years indicated a similar percentage of large trees (> 20 cm ) as diameter distributions described by Lundqvist (1989). Regarding small trees, only treatment TS demonstrated accentuated peaks of ingrowth needed for sustainable future timber harvest. In treatment $T$, only the maximum scenario shows a potential to maintain a stand structure similar to the initial forest today. Under moderate ingrowth assumptions, scarification is necessary to promote a sufficient number of new trees. Considering a total gap area of $15 \%$ in the stand (where most of ingrowth is expected to occur after the scarification), a multi-layered, heterogeneous forest structure of the stand is expected in 50 years. Under treatment T and TN, a lack of regeneration is suspected.

More detailed interpretation of future diameter distributions is mostly speculation due to the limitations and the lack of validation of the growth model (Box 5). However, being aware of limitations and errors due to extrapolation, this simulation study could identify and analyze important growth trends relevant for target diameter cutting. Nevertheless, relatively large deviation from projected diameters can be expected for particular trees. Therefore, re-measurements of the experimental plots over several decades are of tremendous importance to increase the empirical knowledge about stand development and growth under continuous cover forestry.

## Income comparisons

Estimations of gross income and costs showed that a considerable income was obtained by the first target diameter cutting in winter 2008/09 already (about 30-50\% of net income by clear-cut). Considering 60 years or infinite time periods, the NPV of treatment TS was also about $50 \%$ compared to clear-felling. The large impact of the first withdrawal on NPV, but also lower growth, cause the lower value. Since treatment T and TN are likely to lack mature trees in 50 years, NPV calculations with indefinite harvest cycles every 20-25 years were not considered (but can be calculated from Table 23 and 24 , assuming no income during the next 40 years after simulation).
Comparing the total forest development cycle between a selection system and clear-felling, Andreassen \& Oyen (2002) calculated 15\% lower NPV for uneven-aged forests in Norway, based on $2 \%$ discount rate as in the presented study. Usually $3 \%$ and $2.5 \%$ interest rate are used in economic analyses of silviculture in Sweden (Brukas \& Weber, 2009), while $1 \%$ is used in German forestry practice for instance (Möhring, 2001). 2\% were chosen in this study, because forest owners interested in continuous cover forestry might rank profitability lower, eventually. Appendix E demonstrates the sensitivity of NPV to 0 or $4 \%$ rates. Independently from the choice of interest rate, a major conclusion can be drawn: Clear-felling is more profitable. Only when the income of the first harvest is neglected, target diameter cutting with scarification could achieve a higher NPV if discount rates would be larger than 3.2\%. But, there will be periods without income when investments by pre-commercial thinning become necessary. When target diameter cutting is applied, no investment costs will occur or cannot be covered by income at the same time.
Due to difficulties to predict monetary values in 20 years or longer, results have a clear discussion character. Storm and insect damages neglected in the simulation might also alternate the economic results. - Other calculation approaches to compare clear-felling and
partial cutting can be found in Knoke \& Plusczyk (2001) and Emmingham et al. (2002) for instance. Emmingham et al. (2002) calculated the financial value within a 10 years-period only, but including NPV, the value of residual timber, the value of bare land, and the value of planted trees together (which was lower in the treatment with clear-felling). Knoke \& Plusczyk (2001) expected considerably lower income compared to the clear-felling system for stands under transformation towards uneven-aged forest because of delayed income.

## Conclusions

Research conclusions The first hypothesis of the study could not be falsified: (1) Target diameter cutting did create an irregular pattern of gaps in different size classes with light conditions more diverse than found in uniform shelterwoods.
In mature, pine-dominated forest, low forest growth can be expected when applying selective cuttings. By clear-felling, a maximum timber production or a profit maximization can be achieved (because planted spruce stands are expected to produce roughly $50 \%$ more wood within the next 50 years). However, target diameter cutting provided considerable income already at the first time of harvest, which was about $50 \%$ of the income by clear-cut. Additionally, target diameter cutting is expected to provide more equally distributed income over time. Most likely, there will be no urgent need to cover investments costs without achieving some profit at the same time. Other aspects, like aesthetic values or mimicking natural disturbances may give reason to choose target diameter cutting as the most preferable option. It can even be argued that the future forest would be more resilient in case of single-tree fall caused by storm, because established seedlings and smaller trees could replace gaps in the forest canopy.
The projected BA growth was 63-67\% lower compared to a planted spruce stand. Concerning potential modeling errors, the authors expect that the real growth in future will be deviate less than $+/-15 \%$ from projection. However, single-tree growth forecasts in heterogeneously structured forest are extrapolations in most cases. In addition, future standing volume was less reliable to predict than BA. The proportion of old trees and other tree species than spruce, and the possible underestimation of increment of suppressed trees after release might be seen as reason for the surprisingly low productivity. The forest is currently in a transition phase developing from even-aged towards uneven-aged forest during the whole simulation period. Higher increment is expected in single-tree selection forest under equilibrium conditions. The second hypothesis of the study was rejected: (2) For the next 50 years, forest production in the study stand, managed by target diameter cutting, is lower compared to an even-aged spruce stand planted on the same site.
The different thresholds of target diameter did not result in significant growth differences. Therefore, the third hypothesis could not be verified: (3) No decrease of BA growth in treatment T-5 was indicated by the growth model.
The fourth hypothesis could not be falsified: (4) Ingrowth will have a small impact on projected increment the first 50 years. The largest difference was found between the minimum and maximum scenario according to treatment TS with $0.1 \mathrm{~m}^{2} /$ ha respective 0.6 $\mathrm{m}^{3} / \mathrm{ha}$ BA increment annually.
The future tree species composition is expected to change towards more spruce (based on current diameter distribution, regeneration characteristics, and the simulation). According to the forecast, the future proportion of spruce trees increased by $15-20 \%$ in treatment T and

TS, and 2\% in treatment TN and the control. More generally, literature indicates that the choice of tree species is limited under continuous forest cover conditions: Shade-tolerant tree species (spruce, beech) are more likely to dominate in the understory in the long run. The fifth hypothesis could not be falsified: (5) Tree species composition will change towards more shade tolerant species during the simulated development.
With the economic calculations presented before, the hypothesis regarding NPV was falsified. The clear-felling strategy would always result in a higher NPV compared to target diameter cutting, especially due to the high initial income. Considering infinite cutting cycles with today prices, the NPV was about $70 \%$ higher. Hence, the last hypothesis was rejected:
(6) Replanting spruce after clearcut would result in a similar net present value compared to target diameter cutting (based on $2 \%$ interest rate).

Silvicultural conclusions most relevant to forest practice The lack of small trees highlighted the importance to aim for more regeneration in the future to continue with target diameter harvests. Therefore, soil preparation is strongly recommended in mature forests. Another option could be enrichment planting which was not included in this study but was described by Hagner (2004). Without additional regeneration measures, a lack of mature trees after 50 years is likely to cause considerably harvest delays.
In order to provide more practical management guidelines for heterogeneously structured stands (with steadily decreasing tree number in larger size classes, see Box 6) on comparable sites in southern Sweden, about 150-200 $\mathrm{m}^{3} /$ ha standing volume after cut could provide a first practical orientation to achieve regeneration and assure stand growth to some extent. In the studied case, continuous harvest intervals of 20-25 years can be expected if a sufficient number of new trees will occur. Without advanced regeneration or scarification measures, 40-60 years between tree harvests are likely after three target diameter cuttings (beyond the time horizons considered in this simulation). However, the diameter distribution of a particular stand can crucially effect harvest levels and intervals. If advanced regeneration is already established to a large extent, tending of young spruce trees might be necessary in very dense regeneration patches to ensure proper single-tree growth and stability (cf. Rumpf \& Dittges, 2008).
Without silvicultural regeneration measures, the forest will regenerate naturally, but in longer time periods. Developing towards uneven-aged forests, Treatment T and TN could also provide a management option for stands set a-side for natural values according to green management plans (NBF, 2001).


Obviously, such heterogeneously structured forests provide feasible alternatives to clearcutting, i.e. with the seed-tree method to develop future pine forest, or with target diameter cutting to promote continuous forest cover. However, the forest owner is responsible to communicate the management goals in terms of yield, tree species composition, ecological and aesthetic values. Based on such goal settings, forest managers or consultants should give support towards different directions of forest development (not necessarily demanding high production and profitability). With this case study, we hope to provide a bit more guidance to combine the experiences from single-tree selection, shelterwood, seed-treemethod, and clear-cutting to estimate and describe future stand development, growth and costs, in order to fit the owner-specific goals better.

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## References

Ahti T, Hämet-Ahti L, Jalas J. 1968. Vegetation zones and their sections in northwestern Europe. Annales Botanici Fennici 5; 169-211.
Albrecht A, Hein S, Kohnle U, Biber P. 2009. Evaluierung des Waldwachstumssimulators Silva 2.2 anhand langfristiger ertragskundlicher Versuchsflächen in Baden-Württemberg (Evaluation of the forest growth simulator Silva 2.2 by long-term forest yield experioments in Baden-Württemberg). Allg. Forst- u. J.-Ztg. 180, 55-64. (In German, with English summary)
Mette T, Albrecht A, Ammer C, Biber P, Kohnle U, Pretzsch H. 2009. Evaluation of the forest growth simulator SILVA on mature mixed Silver fir - Norway spruce stands in South-West Germany. Ecological Modelling 220, 1670-1680.
Albrecht A, Kohnle U, Nagel J. 2011. Übertragbarkeit empirischer statistischer
Waldwachstumsmodelle: Prüf- und Anpassungsverfahren anhand des Beispiels BWinPro für BadenWürttemberg (Applicability of empirically-statistical forest growth models: Validation and modification procedures exemplified by BWinPro for Baden-Württemberg). Allg. Forst- u. J.-Ztg. 182, 11-23. (In German, with English summary)
Albrecht A, Kohnle U, Nagel J. 2012. Parametrisierung und Evaluierung von BWinPro für BadenWürttemberg anhand waldwachstumskundlicher Versuchsflächendaten (Parameterization and evaluation of BWinPro for Baden-Württemberg by forest yield experimental data. FFF-Bände 51. Freiburg, 40pp. (In German)
Andreassen K. 1994. Development and yield in selection forest. Meddelande Skogforsk 47; 1-37. Andreassen K, Øyen B-H. 2002. Economic consequences of three silvicultural methods in unevenaged mature coastal spruce forests of central Norway. Forestry 75; 483-488.
Anonymus. 2002. R-Polytax (JO0310). Statistics Sweden, Stockholm.
Anonymus. 2009. Blädningsbruk (Single tree selection). Skogsskötselserien 11. Skogsstyrelsen. Available at: http://www.skogsstyrelsen.se/Global/PUBLIKATIONER/Skogsskotselserien/PDF/11Bladningsbruk.pdf [cited 20.11.11]. (In Swedish)
Axelsson R. 2008. Forest policy, continuous tree cover forest and uneven-aged forest management in Sweden's boreal forest. Licentiate thesis. Swedish University of Agricultural Sciences. Department of forest products. ISBN 978-91-85911-56-1.
Bachofen H. 1999. Gleichgewicht, Struktur und Wachstum in Plenterbeständen (Equlibrium, structure and growth in plenterforests). Schweizer Zeitschrift für Forstwesen 150; 157-170. (In German)
Bengtsson L, Rosell S. 2010. Hyggesfritt Skosbruk (Clear-cut free forest management).
Skogsstyrelsen. Rapport. Available at: http://www.skogsstyrelsen.se/Global/aga-och-
bruka/Skogsbruk/Skogsbruks\%c3\%a5tg\%c3\%a4rder/Avverkning/Broschyr\%20hyggesfritt\%20skogsbr uk.pdf [cited 21.11.11]. (In Swedish)
Bilke G. Mortzfeld'sche Lochbestände - ein Weg für den Waldumbau in Nordostdeutschland (Mortzfeld's stands - a way of forest conversion in northeast Germany). Ökologischer Waldumbau im nordostdeutschen Tiefland. Eberswalder Forstliche Schriftenreihe Band XXIII, 111-120 (In German) Biolley H. 1887. Quelques réflexions sur le jardinage à propos des publications de M.
Gournaud.Schweizer Zeitschrift für Forstwesen 38: 189-192. (In French)
Bohn U, Weber H. 2000. Karte der natürlichen Vegetation Europas (Map of natural vegetation in Europe). Bundesamt für Naturschutz. Bonn. (In German)
Brang P. 1998. Early seedling establishment of Picea abies in small forest gaps in the Swiss Alps. Canadian Journal of Forest Research 28; 626-639.
Brukas V, Thorsen BJ, Helles F, Tarp P. 2001. Discount rate and harvest policy. Forest policy and economics 2; 143-156.
Brukas V, Weber, N. 2009. Forest management after the economic transition-at the crossroads between German and Scandinavian traditions. Forest Policy and Economics 11; 586-592.
Bøhmer JG. 1957. Bledningsskog II (Single tree selection forest). Tidskrift for skogsbruk 65; 203-247. (In Norwegian with French summary)
Cameron AD. 2007. Determining the sustainable normal irregular condition: A provisional study on a transformed, irregular mixed species stand in Scotland. Scandinavian Journal of Forest Research 22; 13-21.
Cameron AD, Hands MOR. 2010. Developing a sustainable irregular structure: an evaluation of three inventories at 6-year intervals in an irregular mixed-species stand in Scotland. Forestry 83; 469-475. Carbonnier C. 1975. Produktionen i kulturbestånd av ek i södra Sverige. Production of managed oak stands in southern Sweden. Studia Forestalia Suecica 125;1-89 (In Swedish with English summary). Chrimes D, Nilson K. 2005. Overstorey density influence on the height of Picea abies regeneration in northern Sweden. Forestry 78; 433-442.

Chrimes D. 2004. Stand development and regeneration dynamics of managed uneven-aged Picea abies forests in boreal Sweden. Acta Universitatis Agriculturae Sueciae, Sylvestria 304. ISBN 91-576-6538-9.
Dai X. 1996. Influence of light conditions in canopy gaps on forest regeneration: a new gap light index and its application in a boreal forest in east-central Sweden. Forest Ecology and Management 84; 187-197.
Drössler L. 2010. Tree species mixtures - a common feature of southern Swedish forests. Forestry 83; 433-441.
Drössler L, Attocchi G., Monrad Jensen A. 2012. Occurrence and management of oak in southern Swedish forests. Forstarchiv 83; 163-169.
Eerikäinen K, Miina J, Valkonen S. 2007. Models for the regeneration establishment and the development of established seedlings in uneven-aged, Norway spruce dominated forest stands of southern Finland. Forest Ecology and Management 242; 444-461.
Ek AR, Monserud RA. 1974. Trials with program FOREST: Growth and reproduction simulation for mixed species even- or uneven-aged forest stands. In J. Fries (ed.) Growth models for tree and stand simulation. Proc. IUFRO Working party S.4.01-4 meetings, 1973. Dep. For. Yield Res., Royal College of Forestry, Stockholm. Research Note 30.
Ekö PM. 1985. A growth simulator for Swedish forests, based on data from the national forest survey. Swedish University of Agricultural Sciences. Department of Silviculture, Report 16. ISSN 0348-8969. (In Swedish with English summary)
Elfving B. 2003. Ålderstilldelning till enskilda träd i skogliga tillväxtprognoser (Individual-tree basal area growth functions for forest growth predictions). Swedish University of Agricultural Sciences, Department of Silviculture, Working paper 182. 44 pp. ISSN 0281-7292. (In Swedish with English summary)
Elfving B. 2004. Grundytetillväxtfunktioner för enskilda träd, baserade på data från
riksskogstaxeringens permanenta provytor (Basal area growth functions for single trees, based on data of the Swedish national forest inventory). Swedish University of Agricultural Sciences, Department of Silviculture. Umeå. Unpublished manuscript. (In Swedish)
Elfving B. 2005. En grundytetillväxtfunktion för alla trädslag i hela landet (A basal area growth function for all tree species in the country). Swedish University of Agricultural Sciences, Department of Silviculture. Umeå. Unpublished manuscript. (In Swedish)
Elfving B. 2006a. Urskogens dynamik (Pristine forest dynamics). Swedish University of Agricultural Sciences, Department of Silviculture. Umeå. Unpublished manuscript. (In Swedish)
Elfving B. 2006b. Produktion vid byte från trakthyggen till blädning (Forest productivity when changing from clearfelling to single tree selection). In: (eds) Karlsson B. 2006. Trakthyggesbruk och kontinuitetsskogsbruk med gran, en jämförande studie. Even-aged stand system for Norway spruce versus Continuous-cover forestry. Skogforsk, Redogörelse 5; 12-23. (In Swedish)
Elfving B. 2009. Analys av beståndsutvecklingen pà yta 2280 Fagerland/Hammerdal. Swedish University of Agricultural Sciences, Department of Silviculture. Umeå. Unpublished manuscript. (In Swedish)
Emmingham WL, Oster P, Bennett M, Kukulka F, Conrad K, Michel A. 2002. Comparing short-term financial aspects of fore management options in Oregon: implications for uneven-aged management. Forestry 75; 489-494.
Erefur C. 2010. Regeneration in Continuous Cover Forestry Systems. Doctoral thesis. Swedish University of Agricultural Sciences. Acta Universitatis Agriculturae Sueciae 2010:42.
Erteld T, Gerold D, Mund M, Schulze E-D, Weller E. 2005. Vorrat, Zuwachs und Nutzung im plenterwaldartigen Buchenwald (Standing volume, increment and harvest in single-tree selection beech forest). AFZ 13/2005: 702-706. (In German)
Fahlvik N, Elfving B, Wikström P. 2012. Evaluation of growth functions used in the Heureka forest planning system. (Submitted manuscript).
Falck J, Rydberg D. 1992. Skötsel av tvåskiktade bestånd (Management of two-layered stands). Slutrapport till SJFR. Report to the Swedish Council for Forestry and Agricultural Research. Unpublished manuscript. (In Swedish)
Faustmann M. 1849. Berechnung des Werthes, welchen Waldboden, sowie noch nicht haubare Waldbestände für die Waldwirtschaft besitzen (Calculation of the value which forest land and immature stands possess for forestry). Allgemeine Forst- und Jagdzeitung 15; 441-455. (In German) FC 2004. Managing continuous cover forests. Operational guidance booklet No. 7. Pages 38-39.
Forestry Commission. Available at: http://www.forestry.gov.uk/fr/INFD-7APJ5K
Fridman J, Ståhl G. 2001. A three-step approach for modelling tree mortality in Swedish forests.
Scandinavian Journal of Forest Research 16; 455-466.

Frost I. 1997. Dispersal and establishment of Quercus robur. Ph.D. thesis, dissertation 305, Faculty of Science and Technology, Uppsala University, Sweden.
Glöde D. 2002. Survival and growth of Picea abies regeneration after shelterwood removal with singleand double-grip harvester systems. Scandinavian Journal of Forest Research 17; 417-426.
Granhus A. 2001. Partial cutting in Norway spruce: impacts on advance regeneration and residual stand. Doctoral thesis. Agricultural University of Norway, Department of Forest Sciences, Ås. 30 pp. ISBN 82-575-0461-0.
Granhus A, Hanssen KH, Chantal M. 2008. Emergence and seasonal mortality of naturally regenerated Picea abies seedlings: impact of overstory density and two site preparation methods. New forests 35; 75-87.
Grundmann BM, Bolte A, Bonn S, Roloff A. 2011. Impact of climatic variation on growth of Fagus sylvatica and Picea abies in Southern Sweden. Scandinavian Journal of Forest Research 26; 64-71. Hägglund B. 1973. Om övre höjdens utveckling för gran i södra Sverige (Height development of dominant trees for spruce in southern Sweden). Skogshögskolan, Institution för skogsproduktion, Rapport 24. (In Swedish)
Hägglund B. 1981. Forecasting growth and yield in established forests. Swedish University of Agricultural Sciences, Department of Forest Survey, Report 31. ISBN 91-576-0797-5.
Hägglund B, Lundmark J-E. 1982. Handledning i bonitering med Skogshögskolans
Boniteringssystem - Del 2 Diagram och tabeller (Guide for site productivity assessments in Swedish forests. Part 2 Diagrams and tabels). Skogsstyrelsen. Jönköping. ISBN 91-85748-7. (In Swedish) Hägglund B, Lundmark J-E. 1994. Handledning i bonitering med skogshögskolans boniteringssystem - Del 3 Markvegetationstyper - skogsmarksflora (Guide for site productivity assessments in Swedish forests. Part 3 Forest floor vegetation sites). Skogsstyrelsen. Jönköping. ISBN 91-85748-14-5. (In Swedish)
Hagner M. 2004. Naturkultur. Ekonomiskt skogsbruk kännetecknat av befriande gallring och berikande Plantering (Naturkultur - forest management with single tree release and enrichment planting). ISBN 91-631-5010-7 (In Swedish)
Hartig GL. 1791. Anweisung zur Holzzucht (Instructions for foresters to produce timber). Marburg. Available at: http://books.google.com/books?id=Tuc6AAAAcAAJ\&pg=PR5\#v=onepage\&q\&f=false [cited 24.11.2011] (In German, old script)
Hickler T, Fronzek S, Araújo MB, Schweiger O, Thuiller W, Sykes MT. 2009. An ecosystem modelbased estimate of changes in water availability differs from water proxies that are commonly used in species distribution models. Global Ecology and Biogeography 18; 304-313.
Hofgaard A. 1993. Structure and regeneration patterns in a virgin Picea abies forest in northern Sweden. Journal of Vegetation sciences 4; 601-608.
Holgen P. 1996. Shelterwood systems definitions, functions and conditions for application in Swedish coniferous forests : a review. Swedish University of Agricultural Sciences, Department of Silviculture, Working report 122, Umeå.
Holgén P, Hånell B. 2000. Performance of planted and naturally regenerated seedlings in Picea abiesdominated shelterwood stands and clearcuts in Sweden. Forest Ecology and Management 127; 129138.

Holmgren A. 1959. Skogarna och deras vård i övre Norrland intill år 1930 (The forests and their management in upper Norrland until year 1930). Domänstyrelsen, Stockholm. (In Swedish) Indermühle MP. 1978. Struktur-, Alters- und Zuwachsuntersuchungen in einem Fichten-Plenterwald der subalpinen Stufe (Structure, age, and increment of a subalpine spruce single-tree selection forest).
Beiheft der Zeitung des Schweizer Forstvereins 60. (In German)
Jäderlund A, Zackrisson O, Dahlberg A, Nilsson M-C. 1997. Interference of Vaccinium myrtillus on establishment, growth, and nutrition of Picea abies seedlings in a northern boreal site. Canadian Journal of Forest Research 27: 2017-2025
Juknys R, Vencloviene J, Jurkonis, N, Bartkevicius E, Sepetiene J. 2006. Releation between individual tree mortality and tree characteristics in a polluted and non-polluted environment. Environmental Monitoring and Assessment 121; 519-542.
Karlsson M, Nilsson U. 2005. The effects of scarification and shelterwood treatments on naturally regenerated seedlings in southern Sweden. Forest Ecology and Management 205;183-197. Kätzel R., Löffler S., Winter S., Kallweit R. 2005. Zum Einfluss von Überschirmung und Begründungsverfahren auf den Entwicklungserfolg von Eichen- und Buchen-Voranbauten in der Initialphase (Influence of canopy density and regeneration method on development of oak and beech under pine shelter in the initial stage). Ökologischer Waldumbau im nordostdeutschen Tiefland. Eberswalder Forstliche Schriftenreihe Band XXIII, 79-101 (In German)

Kern KG. 1966. Wachstum und Umweltfaktoren im Schlag- und Plenterwald (Growth and environmental factors in even-aged forest and single tree selection forest). Schriftenreihe der forstlichen Abteilung der Universität Freiburg 5. (In German)
Kerr G, Morgan G, Blyth J, Stokes V. 2010. Transformation from even-aged plantations to an irregular forest: the world's longest running trial area at Glentress, Scotland. Forestry 83; 329-344.
Kint V, Geudens G, Mohren GMJ, Lust N. 2006. Silvicultural interpretation of natural vegetation dynamics in ageing Scots pine stands for their conversion into mixed broadleaved stands. Forest Ecology and Management 223; 363-370.
Knoke T, Plusczyk N. 2001. On economic consequences of transformation of a spruce (Picea abies (L.) Karst.) dominated stand from regular into irregular age structure. Forest Ecology and Management 151; 163-179.
Kolström T. 1993. Modelling the development of an uneven-aged stand of Picea abies. Scandinavian Journal of Forest Research 8; 373-383.
Korpel' S. 1995. Die Urwälder der Westkarpaten (The pristine forest reserves in the western Carpathians). Gustav Fischer, Stuttgart, Jena, New York. (In German)
Kuuluvainen T, Tahvonen O, Aakala T. 2012. Even-aged and uneven-aged forest management in boreal Fennoscandia: A review. Ambio DOI 10.1007/s13280-012-0289-y.
Kunstler G, Albert CA,Courbaud B, Lavergne S, Thuiller W, Vieilledent G, Zimmermann NE, Coomes DA. 2011. Effects of competition on tree radial-growth vary in importance but not in intensity along climatic gradients. Journal of Ecology 99; 300-312.
Kuusipalo J. 1985. On the use of tree stand parameters in estimating light conditions below the canopy. Silva Fennica 19; 185-196.
Lähde E, Laiho O, Norokorpi Y, Saksa T. 2002. Development of Norway spruce dominated stands after single-tree selection and low thinning. Canadian Journal of Forestry Research 32; 1577-1584. Laiho O, Lähde E, Pukkala T 2011. Uneven- vs even-aged management in Finnish boreal forests. Forestry 84; 547-556.
Leemans R. 1991. Canopy gaps and establishment patterns of spruce (Picea abies (L.) Karst.) in two old-growth coniferous forests in central Sweden. Vegetatio 93; 157-165.
Lemmon PE. 1956. A spherical densiometer for estimating forest overstory density. Forest Science 2; 314-320.
Leibundgut H. 1945. Waldbauliche Untersuchungen über den Aufbau von Plenterwäldern (Silvicultural studies on the stand structure of single tree selection forests). Mitteilungen der Schweizer Anstalt für forstliches Versuchswesen 24: 219-296. (In German)
Leibundgut H. 1956: Empfehlungen für die Baumklassenbildung und Methodik bei Versuchen über die Wirkung von Waldpflegemaßnahmen (Recommendations for tree classification and methodology to study the effect of forest tending measures). IUFRO Sekt. 23. 10. report. (In German with French summary)
Linder P. 1998. Stand structure and successional trends in forest reserves in boreal Sweden. Doctor's Dissertation. Acta Universitatis agriculturae Sueciae. Silvestria 72. ISBN 91-576-5606-1.
Lüpke B. 2004. Risikominderung durch Mischwälder und naturnaher Waldbau: ein Spannungsfeld. (Risk aversion by mixed forest and close-to-nature management: An area of conflicts.) Forstarchiv 75; 43-50. (In German with English summary).
Lüpke B, Hauskeller-Bullerjahn M. 1999. Kahlschlagfreier Waldbau: Wird die Eiche an den Rand gedrängt? (Silviculture without clearcuts: Will the proportion of oak decrease?) Forst und Holz 54: 363368.

Lundqvist L. 1989. Blädning i granskog—strukturförändringar, volymtillväxt, inväxning och föryngring på försöksytor skötta med stamvis blädning. Doctoral thesis. Swedish University of Agricultural Sciences, Department of Silviculture, Umeå. (English chapters)
Lundqvist L. 1995. Simulation of sapling population dynamics in uneven-aged Pecia abies forests. Annals of Botany 76; 371-380.
Lundqvist L. 2012. Virkesproduktion och inväxning I skiktad skog efter höggallring (Timber production and ingrowth i multi-layered forest after thinning from above). Skogsstyrelsen rapport 11, 2012. ISSN 1100-0295. (In Swedish)
Lundqvist L, Fridman E. 1996. Influence of local stand BA on density and growth of regeneration in uneven-aged Picea abies stands. Scandinavian Journal of Forest Research 11; 364-369.
Lundqvist L, Chrimes D, Elfving B, Mörling T, Valinger E. 2007. Stand development after different thinnings in two uneven-aged Picea abies forests in Sweden. Forest Ecology and Management 238; 141-146.

Lundström A. 2008. Regionala analyser om kontinuitetsskogar och hyggesfritt skogsbruk (Regional analyses of old-growth forests and continuous cover forestry). Rapport 7. Skogsstyrelsens förlag. Jönköping. (In Swedish)
Liu Q, Hytteborn H. 1991. Gap structure, disturbance and regeneration in a primeval Picea abies forest. Journal of Vegetation Science 2; 391-402.
Mitscherlich J. 1952. Der Tannen-Fichten-(Buchen)-Plenterwald (The fir-spruce-(beech) plenterwald). Schriftenreihe der Badischen Forstlichen Versuchanstalt., Freiburg im Breisgau. (In German)
Möhring B. 2001. The German struggle between the 'Bodenreinertragslehre' (land rent theory) and 'Waldreinertragslehre' (theory of the highest revenue) belongs to the past - but what is left? Forest Policy and Economics 2;195-201.
Mosandl R., Kleinert A. 1998. Development of oaks (Quercus petraea (Matt.) Liebl.) emerged from bird-dispersed seeds under old-growth pine (Pinus silvestris L.) stands. Forest Ecology and Management 106, 35-44.
NBF. 2001. Instruktion för datainsamling vid grön skogsbruksplanering
(Instructions for collection of data during green forest management planning). The National Board of Forestry. Jönköping. Sweden.
Nilson K, Lundqvist L. 2001. Effect of Stand structure and density on development of natural regeneration in two picea abies stands in Sweden. Scandinavian Journal of forest Research 16; 253259.

Nilsson U, Fahlvik N. 2006. Ekonomisk analys av praktisk produktionsoptimering i gran-planteringar (Economical analysis of forestry production optimization in spruce plantations). In Slutrapport för Fiberskogsprogrammet. Bergh J, Oleskog G. Eds. Swedish University of Agricultural Sciences, Southern Swedish Forest Research Centre. Alnarp. pp. 106-129.
Nilsson U, Gemmel P, Johansson U, Karlsson M, Welander T. 2002. Natural regeneration of Norway spruce, Scots pine and birch under Norway spruce shelterwoods of varying densities on a mesic-dry site in southern Sweden. Forest Ecology and Management 161; 133-145.
Nilsson U, Örlander G, Karlsson M. 2006. Establishing mixed forests in Sweden by combining planting and natural regeneration - Effects of shelterwoods and scarification. Forest Ecology and Management 237; 301-311.
Noack M. 2006. Wachstumsgesetzmäßigkeiten der Trauben-Eiche unter Kiefernschirm. (Growth of Sessile oak under pine shelter). Schriftenreihe agrarwissenschaftliche Forschungsergebnisse, Band 28. Hamburg 2006, ISBN 978-3-8300-2331-9. (In German)

NSF 2010. Skogsbruk utan hyggen (Forest management without clearcuts). Naturskyddsförening. Rapport. Available at:
http://www.naturskyddsforeningen.se/upload/skog/rapport_skogsbruk_utan_hyggen.pdf [cited 20.11.11]. (In Swedish)

O'Hara K. 2001. The silviculture of transformation - a commentary. Forest Ecology and Management 151; 81-86.
Örlander G, Karlsson C. 2000. Influence of shelterwood density on survival and height increment of Picea abies. Scandinavian Journal of Forest Research 15; 20-29.
Øyen B-H, Nilsen P. 2004. Growth and recruitment after mountain forest selective cutting in irregular spruce forest. A case study in Northern Norway. Silva Fennica 38; 383-392.
Persson O A. 1992. En produktionsmodell för tallskog i Sverige. (A growth simulator for Scots pine in Sweden.) Department of Forest Yield Research, Report 31. 206 pp. ISSN 0348-7636. (In Swedish with English summary)
Petritan AM, Lüpke B, Petritan IC. 2007. Effects of shade on growth and mortality of maple (Acer pseudoplatanus), ash (Fraxinus excelsior) and beech (Fagus sylvatica) saplings. Forestry 80: 397412.

Petritan AM, Lüpke B. 2009. Struktur und Entwicklung von überschirmten Buchen-Eschen-BergahornDickungen aus Naturverjüngung. (Structure and development of naturally regenerated beech-ashsycamore thickets under canopy.) Forstarchiv 80: 119-128. (In German with English summary) Petterson N. 1992. The effect of stand development on different spacing after planting and precommercial thinning in Norway spruce (Picea abies (L.) Karst.) and Scots pine (Pinus sylvestris L.) stands. Swedish University of Agricultural Sciences. Department of Forest Yield Research. Report 34. ISSN 0348-7636.
Pommerening A, Murphy ST. 2004. A review of the history, definitions and methods of continuous cover forestry with special attention to afforestation and restocking. Forestry 77; 27-44.
Pukkala T. 1987. Simulation model for natural regeneration of Pinus sylvestris, Picea abies, Betula pendula and Betula pubescens. Silva Fennica 21; 37-53.

Pukkala T, Lähde E, Laiho O. 2009. Growth and yield for uneven-sized forest stands in Finland. Forest Ecology and Management 258; 207-216.
Ranneby B, Cruse T. Hägglund B. Jonasson H. Swärd J. 1987. Designing a new national forest survey for Sweden. Studia Forestalia Suecica 177.
Redde N. 2002. Risiko von Sturm- und Folgeschäden in Abhängigkeit vom Standort und von waldbaulichen Eingriffen bei der Umwandlung von Fichtenreinbeständen (Risk of storm- and secondary damages dependent on site and silvicultural interventions to transform pure spruce stands). Dissertation. Fakultät für Forstwissenschaften und Waldökologie der Universität Göttingen. (In German with English Summary)
Reif A, Gärtner S. 2007. Natural regeneration of the deciduous oak species Pedunculate Oak (Quercus robur L.) and Sessile Oak (Quercrus petraea Liebl.) - a literature review with focus on wood pasture. Waldökologie online 5: 79-116. (In German with English summary)
Richter J. 1995. Der Übergang zur Zielstärkennutzung in gleichaltrigen Fichtenbeständen. (The transition to target diameter cutting in even-aged spruce stands.) Forst und Holz 50; 414-415. (In German)
RIS 2008, Fältinstruktion 2008 - Riksinventeringen av skog (Field instructions - National inventory of forest). Swedish University of Agricultural Sciences, Department of Forest Resource Management. Umeå. (In Swedish)
Rumpf H, Ditges J. 2008. Jugendwachstum von Fichtennaturverjüngung in Abhängigkeit von Überschirmungsdichte und Pflegestrategie (Early growth of spruce regeneration in relation to canopy density and tending strategy. Forst und Holz 63; 20-25. (In German with English and French summary) Runkle JR 1992. Guidelines and Sample Protocol for Sampling Forest Gaps. Forest Service, General Technical Report PNW-GTR-283.
Saksa T, Valkonen S 2011. Dynamics of seedling establishment and survival in uneven-aged boreal forests. Forest Ecology and Management 261; 1409-1414.
Shorohova E., Kuuluvainen T., Kangur A., Jogiste K. 2009. Natural stand structures, disturbance regimes and successional dynamics in the Eurasian boreal forests. Annals of Forest Science 66; 201. Schütz J.-P. 1994. Geschichtlicher Hergang und aktuelle Bedeutung der Plenterung in Europa (Historical development and current state of single tree selection in Europe). Allgemeine Forst- und Jagdzeitung 165(5-6): 106-114. (In German).
Schütz J-P. 1997. Sylviculture 2: La gestion des forêts irrégulières et mélanges. Presses
Polytechniques et Universitaires Romandes, Lausanne. (In French)
Schütz J-P. 2001. Opportunities and strategies of transforming regular forests to irregular forests. Forest Ecology and Management 151; 87-94.
Schütz J-P. 2002. Silvicultural tools to develop irregular and diverse forest structures. Forestry 75; 329-337.
Schotte G. 1912. Om gallringsförsök (About thinning experiments). Meddelanden från Statens Skogsförsöksanstalt 9; 211-269. (In Swedish)
Söderberg U. 1992. Functions for forest management: height, form height and bark thickness of individual trees. Swedish University of Agricultural Sciences, Department of Forest Survey, Report 52. 87 pp. ISSN 0348-0496. (In Swedish with English summary)
Spellmann H. 1997. Zielstärkennutzung: Waldbauliche und ertragskundliche Aspekte. (Target diameter cutting: Silvicultural and yield aspects.) Berichte zur Jahrestagung der Sektion Ertragskunde, Deutscher Verband Forstlicher Forschungsanstalten, Grünberg, p. 186-198. (In German)
Stähr F, Peters T, Eisenhauer D-R. 2006. Räumliche Verteilung und waldbauliche Nutzung des Naturverjüngungspotentials von Stiel- und Trauben-Eiche im Land Brandenburg (Spatial distribution and silvicultural utilization of the natural regeneration potential of Pedunculate and Sessile oak in Brandenburg). Eberswalder Forstliche Schriftenreihe, Bd. XXV; 109-117. (In German)
Stähr F, Peters T. 2000. Hähersaat - Qualität und Vitalität natürlicher Eichenverjüngung im nordostdeutschen Tiefland (Seeding by Jay - Quality and vitality of natural oak regeneration in the northeastern German lowland. AFZ/Der Wald 23/2000; 1231-1234. (In German)
Staupendahl K. 1997. Ein neues Stichprobenverfahren zur Erfassung und Beschreibung von Naturverjüngung (A new sample method to measure and describe natural regeneration). In: Jahrestagung der Sektion Forstliche Biometrie und Informatik des Deutschen Verbandes Forstlicher Forschungsanstalten in Freiburg i. Br., 24.-26.9.1997; 32-49. (In German)
Sterba H, Zingg A. 2001. Target diameter harvesting - a strategy to convert even-aged forests. Forest Ecology and Management 151; 95-105.
Strand M, Ottoson-Löfvenius M, Bergsten U, Lundmark T, Rosvall O. 2006. Height growth of planted conifer seedlings in relation to solar radiation and position in Scots pine shelterwood. Forest Ecology and Management 224; 258-265.

Tahvonen O, Pukkala T, Laiho O, Lähde E, Niinimäki S. 2010. Optimal management of uneven-aged Norway spruce stands. Forest Ecology and Management 260; 106-115.
Tremer N. 2008. Untersuchungen zur Verjüngung von Waldbeständen in Nordwestdeutschland (Study on regeneration of forest stands i nortwestern Germany. Doctoral thesis, Faculty of Forest Sciences and Forest Ecology, University of Göttingen. ISBN 978-3-86727-805-8. (In German)
Tron E, Tuhus E. 2001. Models for individual tree mortality in Norway. Forest Ecology and Management 154; 69-84.
Valinger E, Fridman J. 2011. Factors affecting the probability of windthrow at stand level as a result of Gudrun winter storm in southern Sweden. Forest Ecology and Management 262; 398-403.
Wagner S. 1994. Strahlungsschätzung in Wäldern durch hemisphärische Fotos - Methode und
Anwendung (Estimation of radiation in forests by hemispherical photographs - Method and
application). Berichte des Forschungszentrums Waldökosysteme, Göttingen.Reihe A 123; 1-166. (In German)
Wikberg P-E. 2004. Occurrence, morphology and growth of understory saplings in Swedish forests. Doctoral thesis. Acta Universitatis agriculturae Sueciae. Silvestria 322.
Wikström P. 2008. Jämförelse av ekonomi och produktion mellan trakthyggesbruk och blädning i skiktad granskog (Comparison of economy and productivity between the clearfelling system and single tree selection system in multi-layered spruce forest. Rapport 24. Skogsstyrelsens förlag. Jönköping. (In Swedish)
Yue C, Kohnle U, Hein S. 2008. Combining tree- and stand-level growth models: a new approach to growth prediction. For.Sci. 54, (5), 553-566.

## Appendices

## Appendix A



Appendix A, Figure A1. Height curve of spruce


Appendix A, Figure A2. Height curve of pine


Appendix A, Figure A3. Height curve of birch


Appendix A, Figure A5. Height curve of beech


Appendix A, Figure A6. Relation between dbh and tree height of small spruce trees below 5 cm dbh


Appendix A, Figure A7. Relationship between dbh and tree height of small oak trees below 5 cm dbh


Appendix A, Figure A8. Relationship between dbh and tree height of small trees of birch, beech and the rest of other tree species below 5 cm dbh

## Appendix B

Ingrowth scenarios

An assumption for the general scenario construction was made by plots with basal area below $15 \mathrm{~m}^{2} /$ ha representing gaps, and plots with less than $10 \mathrm{~m}^{2} /$ ha representing large gaps. This simplification is rough, but also logical to combine the plot data with gap information (Table B1).

Appendix B, Table B1. Basal area on systematic sample plots and visually estimated gap percentage of the plot area without forest canopy after the first cutting

| gap area <br> on the plot | BA after <br> thinning |  | gap area <br> on the plot | BA after <br> thinning |
| :---: | :---: | :---: | :---: | :---: |
| $\mathbf{7 5 \%}$ | $\mathbf{1 0 . 6}$ |  | 0 | $\mathbf{1 4 . 3}$ |
| $\mathbf{7 5 \%}$ | $\mathbf{1 2 . 4}$ |  | 0 | 26.5 |
| $\mathbf{6 6 \%}$ | $\mathbf{1 4 . 4}$ |  | 0 | 25.0 |
| $\mathbf{5 0 \%}$ | $\mathbf{8 . 3}$ |  | 0 | 30.4 |
| $\mathbf{5 0 \%}$ | 21.1 |  | 0 | 25.2 |
| $\mathbf{4 5 \%}$ | 21.3 |  | 0 | 14.8 |
| $\mathbf{3 5 \%}$ | $\mathbf{1 4 . 5}$ |  | 0 | 24.4 |
| $\mathbf{3 5 \%}$ | 30.0 |  | 0 | 25.7 |
| $10 \%$ | $\mathbf{1 0 . 4}$ |  | 0 | 25.8 |
| $10 \%$ | 17.7 |  | 0 | 24.0 |
| $0 \%$ | 15.1 |  | 0 | 29.6 |
| $0 \%$ | 18.6 |  | 0 | 2.7 |
| $0 \%$ | 21.9 |  | 0 | 38.9 |
| $0 \%$ | 15.5 |  | 0 | 32.5 |
| $0 \%$ | 22.2 |  | 0 | 26.1 |
| $0 \%$ | 22.1 |  | 0 | 27.1 |
| $0 \%$ | 15.4 |  | 0 | 28.0 |
| $0 \%$ | 24.8 |  | 0 | 29.5 |

Appendix B, Table B2. Implemented ingrowth in the simulator on plots representing gaps without and with scarification

| scen. | BA on the plot |  | operational ingrowth on a gap-plot without scar. | operational ingrowth on a gap-plot after scarification |
| :---: | :---: | :---: | :---: | :---: |
| M | $10-15 \mathrm{~m}^{2}$ | spruce | 2 after 10 years | 31 after 40 years |
|  |  |  |  |  |
| N | < $10 \mathrm{~m}^{2}$ | pine <br> birch |  | 16 after 40 years 16 after 30 years |
| $\begin{gathered} \mathrm{M} \\ \mathrm{E} \end{gathered}$ | 10-15 m ${ }^{2}$ | spruce <br> birch | 6 after 5 years <br> 1 after 10 years | 63 after 25 years |
| $\begin{aligned} & \mathrm{A} \\ & \mathrm{~N} \end{aligned}$ | < $10 \mathrm{~m}^{2}$ | pine <br> birch |  | 31 after 25 years <br> 31 after 25 years |
| $\begin{gathered} \mathrm{M} \\ \mathrm{~A} \\ \mathrm{X} \end{gathered}$ | 10-15 m ${ }^{2}$ | spruce <br> birch <br> oak <br> beech | 22 after 5 years <br> 2 after 5 years <br> 1 after 10 years <br> 2 after 10 years | 63 after 20 years |
|  | < $10 \mathrm{~m}^{2}$ | pine <br> birch |  | 31 after 20 years 31 after 20 years |

See figure B1-B3 regarding the occurrence of plots with $B A<15 \mathrm{~m}^{2} /$ ha over the simulation period. (Plots with low BA after 45 years did not contribute with new trees. Under the minimum scenario, only the first 40 years could result in new trees under the simulation period.) Concerning treatment TS, large gaps were represented by 14 plots with $B A<10$ $\mathrm{m}^{2} /$ ha, while plots with $B A<15 \mathrm{~m}^{2} /$ ha were simulated 48 times within the next 50 years.


Appendix B, Figure B1. Number of plots with basal area < $15 \mathrm{~m}^{2} /$ ha at different simulation steps where ingrowth was implemented, according to treatment T , stand age-independent single-tree ages, and different ingrowth scenarios


Appendix B, Figure B2. Number of plots with basal area $<15 \mathrm{~m}^{2} /$ ha at different simulation steps, according to treatment TS, stand age-independent single-tree ages, and different ingrowth scenarios. Full color indicates plots when ingrowth was implemented.


Appendix B, Figure B3. Number of plots with basal area $<15 \mathrm{~m}^{2} /$ ha at different simulation steps where ingrowth was implemented, according to treatment TN, stand age-independent single-tree ages, and different ingrowth scenarios

Appendix C-Simulation output data

Appendix C, Table C1. Simulated initial stand characteristics per hectare of the unmanaged control treatment as starting point for the 50 years simulation on all 48 sample plots

|  | Species | N | BA <br> $\left[\mathrm{m}^{2} / \mathrm{ha}\right]$ | dg <br> $[\mathrm{cm}]$ | Vol <br> $\left[\mathrm{m}^{3} / \mathrm{ha}\right]$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| S | Spruce | 489 | 13.77 | 18.9 | 131 |
| T | Pine | 164 | 15.44 | 34.6 | 145 |
| A | Birch | 54 | 2.59 | 24.6 | 21 |
| R | Beech | 21 | 0.37 | 15.1 | 3 |
| T | Oak | 271 | 2.98 | 11.8 | 21 |
|  | Total | 999 | 35.15 | 21.2 | 322 |

Appendix C, Table C2. Simulated stand characteristics, mean annual increment, and mortality after 50 years without management (control). Figures were calculated per hectare, based on 48 sample plots, stand age-independent single-tree age, and three different ingrowth scenarios (see explanations in the text)


Appendix C, Table C3. Simulated stand characteristics, mean annual increment, and mortality after 50 years without management (control). Figures were calculated per hectare, based on 48 sample plots, stand age-dependent single-tree age, and three different ingrowth scenarios (see explanations in the text).

|  | Species | N | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2}\right]} \\ \hline \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\begin{gathered} \text { Vol } \\ {\left[\mathrm{m}^{3}\right]} \end{gathered}$ | MAI |  | nat. <br> mort. $\left[\mathrm{m}^{3}\right]$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2}\right]} \\ \hline \end{gathered}$ | $\begin{gathered} \mathrm{Vol} \\ {\left[\mathrm{~m}^{3}\right]} \end{gathered}$ |  |
|  | Spruce | 392 | 23.00 | 27.4 | 261 | 0.25 | 3.3 | 35 |
| M | Pine | 139 | 21.97 | 44.6 | 234 | 0.19 | 2.4 | 30 |
|  | Birch | 38 | 2.92 | 31.1 | 26 | 0.02 | 0.2 | 5 |
| N | Beech | 15 | 0.84 | 26.5 | 9 | 0.01 | 0.2 | 1 |
|  | Oak | 181 | 4.81 | 18.4 | 43 | 0.06 | 0.7 | 12 |
|  | Total | 765 | 53.54 | 29.8 | 573 | 0.53 | 6.7 | 83 |
|  | Spruce | 393 | 22.87 | 27.2 | 258 | 0.25 | 3.2 | 34 |
| M | Pine | 138 | 21.88 | 44.9 | 237 | 0.19 | 2.5 | 33 |
| E | Birch | 37 | 2.90 | 31.4 | 26 | 0.02 | 0.2 | 6 |
| A | Beech | 15 | 0.86 | 26.6 | 10 | 0.01 | 0.2 | 1 |
| N | Oak | 181 | 4.86 | 18.5 | 43 | 0.06 | 0.7 | 11 |
|  | Total | 765 | 53.37 | 29.8 | 574 | 0.53 | 6.7 | 85 |
|  | Spruce | 394 | 22.95 | 27.2 | 259 | 0.25 | 3.2 | 34 |
| M | Pine | 140 | 22.15 | 45.0 | 240 | 0.19 | 2.5 | 30 |
| A | Birch | 38 | 2.98 | 31.6 | 26 | 0.02 | 0.2 | 5 |
| X | Beech | 16 | 0.88 | 26.6 | 10 | 0.01 | 0.2 | 1 |
|  | Oak | 183 | 4.85 | 18.4 | 43 | 0.06 | 0.7 | 11 |
|  | Total | 770 | 53.82 | 29.8 | 579 | 0.53 | 6.8 | 81 |

Appendix C, Table C4. Simulated initial stand characteristics per hectare according to treatment $\mathbf{T}$ and $\mathbf{T S}$ as starting point for the 50 years simulation on all 48 sample plots

|  |  |  | BA <br> $\left[\mathrm{m}^{2}\right]$ | dg <br> $[\mathrm{cm}]$ | Vol <br> $\left[\mathrm{m}^{3}\right]$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| S | Spruce | 440 | 8.97 | 16.1 | 77 |
| T | Pine | 109 | 8.58 | 31.7 | 76 |
| A | Birch | 26 | 0.36 | 13.4 | 3 |
| R | Beech | 21 | 0.37 | 15.1 | 3 |
| T | Oak | 271 | 2.98 | 11.8 | 21 |
|  | Total | 866 | 21.26 | 17.7 | 179 |

Appendix C, Table C5. Stand characteristics of treatment T after 50 years simulation, annual increment and harvest (figures per ha, based on 48 sample plots, stand age-independent single-tree ages, three different ingrowth scenarios and minimum $\mathbf{1 0 0} \mathbf{m}^{\mathbf{3}} /$ ha standing volume after cut)

|  |  | N | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\begin{gathered} \text { Vol } \\ {\left[\mathrm{m}^{3}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{BA} \\ \mathrm{MAI} \\ {\left[\mathrm{~m}^{2}\right]} \\ \hline \end{gathered}$ | Vol <br> MAI $\left[\mathrm{m}^{3}\right]$ | Time of cutting [years] | 1th cut removal [ $\mathrm{m}^{3}$ ] | 2nd cut removal $\left[\mathrm{m}^{3}\right]$ | 3rd cut removal $\left[\mathrm{m}^{3}\right]$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Spruce | 312 | 11.83 | 22.0 | 117 | 0.27 | 3.1 |  | 52 | 43 | 64 |
| M | Pine | 21 | 2.33 | 37.5 | 22 | 0.12 | 1.3 |  | 69 | 62 | 57 |
| 1 | Birch | 12 | 0.28 | 17.2 | 2 | 0.01 | 0.0 |  | 18 | 2 | 1 |
| N | Beech | 8 | 0.31 | 21.9 | 3 | 0.02 | 0.2 |  | 0 | 4 | 7 |
|  | Oak | 199 | 7.05 | 21.2 | 55 | 0.13 | 1.1 |  | 0 | 2 | 18 |
|  | Total | 552 | 21.80 | 22.4 | 198 | 0.54 | 5.8 | 0/25/50 | 139 | 113 | 147 |


| Spruce | 410 | 13.03 | 20.1 | 127 |  | 0.30 | 3.3 |  |  | 52 | 46 | 52 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| M | Pine | 21 | 2.36 | 37.5 | 22 |  | 0.11 | 1.3 |  |  | 69 | 60 |
| E | Birch | 32 | 0.41 | 12.8 | 3 |  | 0.01 | 0.1 |  |  | 18 | 2 |
| A | Beech | 8 | 0.32 | 22.1 | 3 | 0.02 | 0.2 |  |  | 0 | 4 | 5 |
| N | Oak | 200 | 7.06 | 21.2 | 55 | 0.13 | 1.1 |  |  | 0 | 2 | 12 |
|  | Total | 672 | 23.18 | 21.0 | 209 |  | 0.56 | 5.9 |  | $0 / 25 / 45$ | 139 | 114 |


| Spruce | 555 | 14.89 | 18.5 | 143 |  | 0.33 | 3.7 |  |  | 52 | 45 | 51 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| M | Pine | 21 | 2.28 | 37.4 | 21 |  | 0.11 | 1.3 |  |  | 69 | 60 |
| A | Birch | 26 | 0.37 | 13.4 | 3 |  | 0.01 | 0.1 |  |  | 18 | 2 |
| X | Beech | 19 | 0.40 | 16.4 | 3 | 0.02 | 0.2 |  |  | 0 | 4 | 5 |
|  | Oak | 211 | 7.11 | 20.7 | 55 | 0.13 | 1.1 |  |  | 0 | 2 | 12 |
| Total | 832 | 25.06 | 19.6 | 225 |  | 0.60 | 6.2 |  | $0 / 25 / 45$ | 139 | 113 | 119 |

Appendix C, Table C6. Stand characteristics of treatment T after 50 years simulation, annual increment and harvest (figures per ha, based on 48 sample plots, stand age-independent single-tree ages, three different ingrowth scenarios and minimum $\mathbf{1 5 0} \mathbf{m}^{\mathbf{3}}$ /ha standing volume after cut)

|  |  | N | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2}\right]} \\ \hline \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\begin{gathered} \text { Vol } \\ {\left[\mathrm{m}^{3}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{BA} \\ \mathrm{MAI} \\ {\left[\mathrm{~m}^{2}\right]} \\ \hline \end{gathered}$ | Vol <br> MAI $\left[\mathrm{m}^{3}\right]$ | Time of cutting [years] | 1th cut removal $\left[\mathrm{m}^{3}\right]$ | 2nd cut removal $\left[\mathrm{m}^{3}\right]$ | 3rd cut removal $\left[\mathrm{m}^{3}\right]$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Spruce | 308 | 22.74 | 22.0 | 116 | 0.27 | 3.2 |  | 52 | 46 | 52 |
| M | Pine | 21 | 2.28 | 37.5 | 21 | 0.11 | 1.3 |  | 69 | 60 | 51 |
| 1 | Birch | 12 | 0.27 | 16.8 | 2 | 0.01 | 0.0 |  | 18 | 2 | 0 |
| N | Beech | 8 | 0.31 | 22.4 | 3 | 0.02 | 0.2 |  | 0 | 4 | 6 |
|  | Oak | 197 | 6.99 | 21.3 | 54 | 0.13 | 1.1 |  | 0 | 3 | 12 |
|  | Total | 545 | 21.59 | 22.5 | 196 | 0.54 | 5.8 | 0/25/45 | 139 | 115 | 121 |


| Spruce |  | 412 | 13.05 | 20.1 | 127 |  | 0.30 | 3.4 |  |  | 52 | 45 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| M | Pine | 21 | 2.29 | 37.4 | 21 |  | 0.11 | 1.3 |  |  | 69 | 60 |
| E | Birch | 29 | 0.39 | 13.0 | 3 |  | 0.01 | 0.1 |  |  | 18 | 2 |
| A | Beech | 8 | 0.32 | 22.2 | 3 |  | 0.02 | 0.2 |  |  | 0 | 4 |
| N | Oak | 201 | 7.06 | 21.1 | 55 |  | 0.13 | 1.1 |  |  | 0 | 2 |
|  | Total | 672 | 23.11 | 20,9 | 208 |  | 0.56 | 5.9 |  | $0 / 25 / 45$ | 139 | 113 |


| Spruce | 560 | 14.91 | 18.4 | 143 |  | 0.33 | 3.7 |  |  | 52 | 45 | 51 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| M | Pine | 21 | 2.30 | 37.5 | 21 |  | 0.11 | 1.3 |  |  | 69 | 60 |
| A | Birch | 26 | 0.36 | 13.2 | 2 |  | 0.01 | 0.1 |  |  | 18 | 2 |
| X | Beech | 19 | 0.40 | 16.4 | 3 | 0.02 | 0.2 |  |  | 0 | 4 | 5 |
|  | Oak | 211 | 7.07 | 20.7 | 55 | 0.13 | 1.1 |  |  | 0 | 3 | 12 |
| Total | 837 | 25.04 | 19.5 | 225 |  | 0.60 | 6.2 |  | $0 / 25 / 45$ | 139 | 114 | 118 |

Appendix C, Table C7. Harvest and initial stand characteristics per hectare on block 1 according to treatment $\mathbf{T}$

| Species | N | removed |  |  | after cutting |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\begin{aligned} & \text { Vol } \\ & {\left[\mathrm{m}^{3}\right]} \end{aligned}$ | N | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\begin{aligned} & \mathrm{Vol} \\ & {\left[\mathrm{~m}^{3}\right]} \end{aligned}$ |
| Spruce | 34 | 3.36 | 35.6 | 35 | 358 | 6.44 | 15.1 | 53 |
| Pine | 34 | 4.25 | 40.0 | 42 | 101 | 8.18 | 32.0 | 73 |
| Birch | 46 | 3.57 | 31.5 | 30 | 32 | 0.63 | 15.8 | 5 |
| Beech |  |  |  |  | 26 | 0.48 | 15.3 | 4 |
| Oak |  |  |  |  | 448 | 4.81 | 11.7 | 34 |
| Total | 113 | 11.19 | 35.4 | 107 | 965 | 20.53 | 16.5 | 168 |

Appendix C, Table C8. Stand characteristics, annual increment and harvest on block 1 according to treatment $\mathbf{T}$ after 50 years simulation (figures per ha, based on stand ageindependent single-tree ages, with the moderate ingrowth scenario and minimum 100 $\mathrm{m}^{3} /$ ha standing volume after cut)

| Species | N | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2}\right]} \\ \hline \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\begin{gathered} \text { Vol } \\ {\left[\mathrm{m}^{3}\right]} \\ \hline \end{gathered}$ | $\begin{gathered} \mathrm{BA} \\ \mathrm{MAI} \\ {\left[\mathrm{~m}^{2}\right]} \\ \hline \end{gathered}$ | Vol <br> MAI <br> [m ${ }^{3}$ ] | Time of cutting [years] | 1th cut removal [ $\mathrm{m}^{3}$ ] | 2nd cut removal $\left[\mathrm{m}^{3}\right]$ | 3rd cut removal [ $\mathrm{m}^{3}$ ] |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Spruce | 241 | 8.17 | 20.8 | 77 | 0.24 | 2.7 |  | 35 | 38 | 56 |
| Pine | 13 | 1.26 | 35.6 | 11 | 0.11 | 1.2 |  | 42 | 62 | 55 |
| Birch | 20 | 0.27 | 13.2 | 2 | 0.01 | 0.1 |  | 30 | 5 | 0 |
| Beech | 7 | 0.19 | 18.2 | 2 | 0.02 | 0.3 |  | 0 | 5 | 9 |
| Oak | 328 | 10.99 | 20.6 | 86 | 0.20 | 1.7 |  | 0 | 2 | 21 |
| Total | 609 | 20.88 | 20.9 | 177 | 0.58 | 5.9 | 0/25/50 | 107 | 112 | 142 |

Appendix C, Table C9. Harvest and initial stand characteristics per hectare on block 2 according to treatment T

| removed |  |  |  |  |  |  | after cutting |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | N | BA <br> $\left[\mathrm{m}^{2}\right]$ | dg <br> $[\mathrm{cm}]$ | Vol <br> $\left[\mathrm{m}^{3}\right]$ | N | BA <br> $\left[\mathrm{m}^{2}\right]$ | dg <br> $[\mathrm{cm}]$ | Vol <br> $\left[\mathrm{m}^{3}\right]$ |  |  |  |
| Spruce | 48 | 3.77 | 31.7 | 40 | 342 | 6.76 | 15.9 | 56 |  |  |  |
| Pine | 84 | 10.51 | 40.0 | 105 | 127 | 10.25 | 32.0 | 90 |  |  |  |
| Birch | 16 | 1.36 | 33.0 | 11 | 26 | 0.21 | 10.2 | 1 |  |  |  |
| Beech |  |  |  |  | 28 | 0.57 | 16.2 | 5 |  |  |  |
| Oak |  |  |  |  | 247 | 2.92 | 12.3 | 21 |  |  |  |
| Total | 147 | 15.65 | 36.8 | 156 | 770 | 20.72 | 18.5 | 173 |  |  |  |

Appendix C, Table C10. Stand characteristics, annual increment and harvest on block 2 according to treatment $\mathbf{T}$ after 50 years simulation (figures per ha, based on stand ageindependent single-tree ages, with the moderate ingrowth scenario and minimum 100 $\mathrm{m}^{\mathbf{3}} / \mathrm{ha}$ standing volume after cut)

| Species | N | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\begin{gathered} \mathrm{Vol} \\ {\left[\mathrm{~m}^{3}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{BA} \\ \mathrm{MAI} \\ {\left[\mathrm{~m}^{2}\right]} \\ \hline \end{gathered}$ | Vol <br> MAI $\left[\mathrm{m}^{3}\right]$ | Time of cutting [years] | 1th cut removal $\left[\mathrm{m}^{3}\right]$ | 2nd cut removal $\left[\mathrm{m}^{3}\right]$ | 3rd cut removal $\left[\mathrm{m}^{3}\right]$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Spruce | 339 | 12.45 | 21.6 | 123 | 0.24 | 2.7 |  | 40 | 47 |  |
| Pine | 70 | 9.51 | 41.7 | 93 | 0.14 | 1.6 |  | 105 | 68 |  |
| Birch | 34 | 0.38 | 12.0 | 3 | 0.01 | 0.0 |  | 11 | 0 |  |
| Beech | 18 | 1.27 | 29.6 | 13 | 0.03 | 0.3 |  | 0 | 6 |  |
| Oak | 192 | 7.65 | 22.5 | 63 | 0.12 | 1.1 |  | 0 | 5 |  |
| Total | 654 | 31.27 | 24.7 | 296 | 0.54 | 5.8 | 0/25 | 156 | 126 |  |

Appendix C, Table C11. Harvest and initial stand characteristics per hectare on block 3 according to treatment $\mathbf{T}$

| removed |  |  |  |  |  |  | after cutting |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | N | BA <br> $\left[\mathrm{m}^{2}\right]$ | dg <br> $[\mathrm{cm}]$ | Vol <br> $\left[\mathrm{m}^{3}\right]$ | N | BA <br> $\left[\mathrm{m}^{2}\right]$ | dg <br> $[\mathrm{cm}]$ | Vol <br> $\left[\mathrm{m}^{3}\right]$ |  |  |  |
| Spruce | 68 | 7.27 | 37.0 | 81 | 619 | 13.71 | 16.8 | 121 |  |  |  |
| Pine | 48 | 5.82 | 39.4 | 60 | 97 | 7.32 | 30.9 | 64 |  |  |  |
| Birch | 24 | 1.75 | 30.5 | 14 | 20 | 0.26 | 12.8 | 2 |  |  |  |
| Beech |  |  |  |  | 8 | 0.05 | 9.3 | 0 |  |  |  |
| Oak |  |  |  |  | 119 | 1.21 | 11.3 | 8 |  |  |  |
| Total | 139 | 14.83 | 36.8 | 155 | 863 | 22.54 | 18.2 | 196 |  |  |  |

Appendix C, Table C12. Stand characteristics, annual increment and harvest on block 3 according to treatment $\mathbf{T}$ after 50 years simulation (figures per ha, based on stand ageindependent single-tree ages, with the moderate ingrowth scenario and minimum 100 $\mathbf{m}^{\mathbf{3}}$ /ha standing volume after cut)

| Species | N | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\begin{gathered} \text { Vol } \\ {\left[\mathrm{m}^{3}\right]} \end{gathered}$ | $\begin{array}{r} \mathrm{BA} \\ \mathrm{MAI} \\ {\left[\mathrm{~m}^{2}\right]} \\ \hline \end{array}$ | Vol <br> MAI $\left[\mathrm{m}^{3}\right]$ | Time of cutting [years] | 1th cut removal [ $\mathrm{m}^{3}$ ] | 2nd cut removal $\left[\mathrm{m}^{3}\right]$ | 3rd cut removal [m ${ }^{3}$ ] |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Spruce | 594 | 20.59 | 21.0 | 206 | 0.41 | 4.7 |  | 81 | 51 | 75 |
| Pine | 23 | 2.37 | 36.2 | 22 | 0.10 | 1.1 |  | 60 | 50 | 41 |
| Birch | 39 | 0.56 | 13.4 | 4 | 0.01 | 0.1 |  | 14 | 0 | 0 |
| Beech | 5 | 0.06 | 13.1 | 0 | 0.00 | 0.0 |  | 0 | 0 | 2 |
| Oak | 88 | 2.84 | 20.3 | 22 | 0.05 | 0.4 |  | 0 | 0 | 5 |
| Total | 748 | 26.42 | 21.2 | 255 | 0.57 | 6.3 | 0/25/45 | 155 | 102 | 123 |

Appendix C, Table C13. Stand characteristics of the treatment TS after 50 years simulation, annual increment and harvest (figures per ha, based on stand age-independent single-tree ages, three different ingrowth scenarios and minimum $\mathbf{1 0 0} \mathbf{~ m}^{\mathbf{3}}$ /ha standing volume after cut)

|  |  | $\begin{gathered} \mathrm{N} \\ {\left[\mathrm{ha}^{-1}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\begin{gathered} \mathrm{Vol} \\ {\left[\mathrm{~m}^{3}\right]} \end{gathered}$ | BA <br> MAI $\left[\mathrm{m}^{2}\right]$ | Vol <br> MAI $\left[\mathrm{m}^{3}\right]$ | Time of cutting [years] | 1th cut removal $\left[\mathrm{m}^{3}\right]$ | 2nd cut removal $\left[\mathrm{m}^{3}\right]$ | 3rd cut removal $\left[\mathrm{m}^{3}\right]$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Spruce | 415 | 10.88 | 18.3 | 104 | 0.28 | 3.1 |  | 53 | 46 | 64 |
| M | Pine | 44 | 1.75 | 22.4 | 16 | 0.12 | 1.3 |  | 69 | 60 | 59 |
| 1 | Birch | 36 | 0.33 | 10.7 | 2 | 0.01 | 0.0 |  | 19 | 2 | 1 |
| N | Beech | 7 | 0.23 | 19.8 | 2 | 0.02 | 0.2 |  | 0 | 4 | 7 |
|  | Oak | 190 | 6.23 | 20.5 | 48 | 0.13 | 1.1 |  | 0 | 3 | 18 |
|  | Total | 692 | 19.41 | 18.9 | 172 | 0.55 | 5.8 | 0/25/50 | 141 | 114 | 149 |


|  | Spruce | 689 | 14.20 | 16.2 | 134 |  | 0.32 | 3.4 |  |  | 53 | 44 | 51 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Pine | 104 | 2.77 | 18.4 | 24 |  | 0.12 | 1.3 |  |  | 69 | 60 | 51 |
| E | Birch | 108 | 0.64 | 8.7 | 4 |  | 0.01 | 0.1 |  |  | 19 | 2 | 1 |
| A | Beech | 8 | 0.32 | 22.1 | 3 |  | 0.02 | 0.2 |  |  | 0 | 4 | 5 |
| N | Oak | 203 | 7.13 | 21.1 | 55 | 0.13 | 1.1 |  |  | 0 | 2 | 11 |  |
|  | Total | 1113 | 25.06 | 16.9 | 220 |  | 0.60 | 6.1 |  | $0 / 25 / 45$ | 141 | 113 | 119 |


| Spruce | 760 | 16.02 | 16.4 | 151 |  | 0.35 | 3.7 |  |  | 53 | 44 | 50 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| M | Pine | 105 | 2.91 | 18.8 | 25 |  | 0.12 | 1.3 |  |  | 69 | 60 |
| A | Birch | 109 | 0.72 | 9.1 | 4 |  | 0.02 | 0.1 |  |  | 19 | 2 |
|  | Beech | 15 | 0.36 | 17.3 | 3 | 0.02 | 0.2 |  |  | 0 | 4 | 5 |
|  | Oak | 212 | 7.18 | 20.7 | 55 | 0.13 | 1.0 |  |  | 0 | 2 | 11 |
| Total | 1201 | 27.18 | 17.0 | 239 | 0.64 | 6.4 |  | $0 / 25 / 45$ | 141 | 112 | 118 |  |

Appendix C, Table C14. Stand characteristics of the treatment TS after 50 years simulation, annual increment and harvest (figures per ha, based on stand age-dependent single-tree ages, three different ingrowth scenarios and minimum $\mathbf{1 0 0} \mathbf{m}^{\mathbf{3}}$ /ha standing volume after cut)

|  |  | N | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\begin{gathered} \text { Vol } \\ {\left[\mathrm{m}^{3}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{BA} \\ \mathrm{MAI} \\ {\left[\mathrm{~m}^{2}\right]} \\ \hline \end{gathered}$ | Vol <br> MAI $\left[\mathrm{m}^{3}\right]$ | Time of cutting [years] | 1th cut removal $\left[\mathrm{m}^{3}\right]$ | 2nd cut removal $\left[\mathrm{m}^{3}\right]$ | 3rd cut removal [m ${ }^{3}$ ] |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Spruce | 426 | 11.25 | 18.3 | 107 | 0.27 | 3.1 |  | 53 | 43 | 63 |
| M | Pine | 44 | 1.85 | 23.2 | 17 | 0.12 | 1.4 |  | 69 | 62 | 58 |
| 1 | Birch | 37 | 0.33 | 10.8 | 2 | 0.01 | 0.0 |  | 19 | 2 | 1 |
| N | Beech | 7 | 0.24 | 20.2 | 2 | 0.02 | 0.2 |  | 0 | 4 | 7 |
|  | Oak | 195 | 6.40 | 20.5 | 50 | 0.13 | 1.1 |  | 0 | 2 | 17 |
|  | Total | 709 | 20.06 | 19.0 | 178 | 0.55 | 5.8 | 0/25/50 | 141 | 113 | 145 |


|  | Spruce | 660 | 14.65 | 16.8 | 139 |  | 0.31 | 3.4 |  |  | 53 | 42 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| M | Pine | 125 | 2.85 | 17.0 | 25 |  | 0.12 | 1.3 |  |  | 69 | 62 |
| E | Birch | 129 | 0.65 | 8.0 | 4 |  | 0.01 | 0.1 |  |  | 19 | 2 |
| A | Beech | 9 | 0.33 | 22.1 | 3 | 0.02 | 0.2 |  |  | 0 | 4 | 5 |
| N | Oak | 203 | 7.08 | 21.1 | 55 | 0.13 | 1.1 |  |  | 0 | 2 | 12 |
|  | Total | 1125 | 25.56 | 17.0 | 226 |  | 0.60 | 6.0 |  | $0 / 25 / 45$ | 141 | 112 |


| Spruce | 717 | 15.87 | 16.8 | 150 |  | 0.34 | 3.6 |  |  | 53 | 41 | 48 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| M | Pine | 122 | 2.88 | 17.3 | 25 |  | 0.13 | 1.4 |  |  | 69 | 62 |
| A | Birch | 125 | 0.72 | 8.5 | 4 |  | 0.02 | 0.1 |  |  | 19 | 2 |
|  | Beech | 15 | 0.37 | 17.7 | 3 |  | 0.02 | 0.2 |  |  | 0 | 4 |
|  | Oak | 211 | 7.15 | 20.8 | 55 | 0.13 | 1.0 |  |  | 0 | 2 | 11 |
|  | Total | 1190 | 26.99 | 17.0 | 237 |  | 0.63 | 6.3 |  | $0 / 25 / 45$ | 141 | 112 |

Appendix C, Table C15. Stand characteristics of the treatment TS after 50 years simulation, annual increment and harvest (figures per ha, based on stand age-independent single-tree ages, three different ingrowth scenarios and minimum $\mathbf{1 5 0} \mathbf{m}^{\mathbf{3}} / \mathrm{ha}$ standing volume after cut)

|  |  | $\begin{gathered} \mathrm{N} \\ {\left[\mathrm{ha}^{-1}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2}\right]} \\ \hline \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\begin{aligned} & \text { Vol } \\ & {\left[\mathrm{m}^{3}\right]} \end{aligned}$ | BA <br> MAI $\left[\mathrm{m}^{2}\right]$ | Vol MAI [m ${ }^{3}$ ] | Time of cutting [years] | 1th cut removal $\left[\mathrm{m}^{3}\right]$ | 2nd cut remova $\left[\mathrm{m}^{3}\right]$ | 3rd cut removal $\left[\mathrm{m}^{3}\right]$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Spruce | 340 | 10.78 | 20.1 | 104 | 0.27 | 3.0 |  | 53 | 43 | 63 |
| M | Pine | 27 | 1.80 | 29.1 | 16 | 0.12 | 1.4 |  | 69 | 62 | 58 |
| 1 | Birch | 20 | 0.26 | 12.8 | 2 | 0.01 | 0.0 |  | 19 | 2 | 1 |
| N | Beech | 8 | 0.25 | 20.7 | 2 | 0.02 | 0.2 |  | 0 | 4 | 7 |
|  | Oak | 190 | 6.27 | 20.5 | 49 | 0.13 | 1.1 |  | 0 | 3 | 17 |
|  | Total | 586 | 19.36 | 20.5 | 173 | 0.54 | 5.7 | 0/25/50 | 141 | 113 | 146 |


|  | Spruce | 691 | 14.32 | 16.2 | 136 |  | 0.32 | 3.4 |  |  | 53 | 45 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | M | Pine | 105 | 2.84 | 18.5 | 25 |  | 0.12 | 1.3 |  |  | 69 |
| E | Birch | 108 | 0.64 | 8.7 | 4 |  | 0.01 | 0.1 |  |  | 19 | 2 |
| A | Beech | 9 | 0.33 | 21.9 | 3 |  | 0.02 | 0.2 |  |  | 0 | 4 |
| N | Oak | 203 | 7.14 | 21.1 | 55 | 0.13 | 1.1 |  |  | 0 | 2 | 12 |
|  | Total | 1117 | 25.27 | 17.0 | 223 | 0.60 | 6.1 |  | $0 / 25 / 45$ | 141 | 113 | 119 |


| Spruce | 774 | 15.96 | 16.2 | 150 |  | 0.35 | 3.7 |  |  | 53 | 44 | 50 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Pine | 103 | 2.90 | 18.9 | 25 |  | 0.12 | 1.3 |  |  | 69 | 60 |
| A | Birch | 109 | 0.71 | 9.1 | 4 |  | 0.02 | 0.1 |  |  | 19 | 2 |
| X | Beech | 15 | 0.34 | 17.1 | 3 |  | 0.02 | 0.2 |  |  | 0 | 4 |
|  | Oak | 209 | 7.11 | 20.8 | 55 | 0.13 | 1.0 |  |  | 0 | 2 | 11 |
| Total | 1210 | 27.02 | 16.9 | 238 |  | 0.64 | 6.4 |  | $0 / 25 / 45$ | 141 | 113 | 117 |

Appendix C, Table C16. Simulated initial stand characteristics per hectare according to treatment TN as starting point for the 50 years simulation on all 48 sample plots

|  |  |  | BA <br> $\left[{ }^{2}\right]$ | dg <br> $[\mathrm{cm}]$ | Vol <br> $\left[\mathrm{m}^{3}\right]$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| S | Spruce | 422 | 7.80 | 15.3 | 65 |
| T | Pine | 136 | 11.15 | 32.3 | 100 |
| A | Birch | 38 | 0.98 | 18.2 | 8 |
| R | Beech | 21 | 0.37 | 15.1 | 3 |
| T | Oak | 271 | 2.98 | 11.8 | 21 |
|  | Total | 887 | 23.28 | 18.3 | 197 |

Appendix C, Table C17. Stand characteristics of treatment TN after 50 years simulation, annual increment and harvest (figures per ha, based on stand age-independent single-tree ages, three different ingrowth scenarios and minimum $\mathbf{1 0 0} \mathbf{m}^{\mathbf{3}} / \mathrm{ha}$ standing volume after cut)

|  |  | N | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\begin{gathered} \text { Vol } \\ {\left[\mathrm{m}^{3}\right]} \end{gathered}$ | BA <br> MAI $\left[\mathrm{m}^{2}\right]$ | Vol <br> MAI $\left[\mathrm{m}^{3}\right]$ | Time of cutting [years] | 1th cut removal $\left[\mathrm{m}^{3}\right]$ | 2nd cut removal $\left[\mathrm{m}^{3}\right]$ | 3rd cut removal [ $\mathrm{m}^{3}$ ] |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Spruce | 227 | 6.18 | 18.6 | 55 | 0.24 | 2.6 |  | 64 | 53 | 71 |
| M | Pine | 29 | 3.18 | 37.4 | 29 | 0.14 | 1.6 |  | 44 | 61 | 80 |
| 1 | Birch | 17 | 0.52 | 20.0 | 4 | 0.01 | 0.1 |  | 13 | 4 | 2 |
| N | Beech | 18 | 1.46 | 32.3 | 16 | 0.02 | 0.3 |  | 0 | 0 | 0 |
|  | Oak | 214 | 8.36 | 22.3 | 66 | 0.13 | 1.1 |  | 0 | 1 | 1 |
|  | Total | 505 | 19.72 | 22.3 | 169 | 0.54 | 5.6 | 0/20/45 | 122 | 119 | 156 |


|  | Spruce | 340 | 7.34 | 16.6 | 64 |  | 0.26 | 2.8 |  |  | 64 | 53 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| M | Pine | 29 | 3.17 | 37.3 | 29 |  | 0.14 | 1.6 |  |  | 44 | 61 |
| E | Birch | 33 | 0.61 | 15.3 | 4 |  | 0.01 | 0.1 |  |  | 13 | 5 |
| A | Beech | 17 | 1.37 | 32.1 | 15 | 0.02 | 0.3 |  |  | 0 | 0 | 0 |
| N | Oak | 212 | 8.29 | 22.3 | 65 | 0.13 | 1.1 |  |  | 0 | 1 | 2 |
|  | Total | 631 | 20.78 | 20.5 | 177 |  | 0.56 | 5.8 |  | $0 / 20 / 45$ | 122 | 119 |


| Spruce | 533 | 10.76 | 16.0 | 96 |  | 0.30 | 3.1 |  |  | 64 | 53 | 55 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| M | Pine | 36 | 4.14 | 38.4 | 39 |  | 0.13 | 1.5 |  |  | 44 | 60 |
| A | Birch | 33 | 0.69 | 16.2 | 5 |  | 0.01 | 0.1 |  |  | 13 | 5 |
| X | Beech | 33 | 1.55 | 24.4 | 16 |  | 0.03 | 0.3 |  |  | 0 | 0 |
|  | Oak | 232 | 8.66 | 21.8 | 68 | 0.13 | 1.1 |  |  | 0 | 1 | 1 |
|  | Total | 867 | 25.79 | 19.5 | 224 |  | 0.61 | 6.1 |  | $0 / 20 / 40$ | 122 | 119 |

Appendix C, Table C18. Simulated initial stand characteristics per hectare according to the treatment $\mathbf{T}$ with $\mathbf{5 c m}$ higher target diameters as starting point for the 50 years simulation on all 48 sample plots

|  |  |  | BA <br> N | dg <br> $[\mathrm{cm}]$ | Vol <br> $\left[\mathrm{m}^{3}\right]$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| S | Spruce | 489 | 13.77 | 18.9 | 131 |
| T | Pine | 164 | 15.44 | 34.6 | 145 |
| A | Birch | 54 | 2.59 | 24.6 | 21 |
| R | Beech | 21 | 0.37 | 15.1 | 3 |
| T | Oak | 271 | 2.98 | 11.8 | 21 |
|  | Total | 999 | 35.15 | 21.2 | 322 |

Appendix C, Table C19. Stand characteristics, annual increment and harvest according to treatment $\mathbf{T}$ with $5 \mathbf{c m}$ higher target diameters after 50 years simulation (figures per ha, based on stand age-independent single-tree ages, three different ingrowth scenarios and minimum $100 \mathbf{m}^{\mathbf{3}}$ /ha standing volume after cut)

|  |  | N | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\begin{gathered} \text { Vol } \\ {\left[\mathrm{m}^{3}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{BA} \\ \mathrm{MAI} \\ {\left[\mathrm{~m}^{2}\right]} \\ \hline \end{gathered}$ | Vol <br> MAI <br> [ $\mathrm{m}^{3}$ ] | Time of cutting [years] | 1th cut removal [ $\mathrm{m}^{3}$ ] | 2nd cut removal $\left[\mathrm{m}^{3}\right]$ | 3rd cut removal $\left[\mathrm{m}^{3}\right]$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Spruce | 318 | 13.78 | 23.5 | 140 | 0.27 | 3.2 |  | 44 | 39 | 46 |
| M | Pine | 51 | 6.34 | 39.8 | 62 | 0.15 | 1.8 |  | 52 | 59 | 52 |
| 1 | Birch | 14 | 0.33 | 17.6 | 3 | 0.01 | 0.1 |  | 16 | 4 | 1 |
| N | Beech | 10 | 0.43 | 22.8 | 4 | 0.02 | 0.2 |  | 0 | 2 | 5 |
|  | Oak | 198 | 6.57 | 20.6 | 53 | 0.10 | 0.9 |  | 1 | 1 | 5 |
|  | Total | 591 | 27.45 | 24.3 | 261 | 0.55 | 6.2 | 5/30/50 | 114 | 106 | 110 |


|  | Spruce | 347 | 14.09 | 22.7 | 143 |  | 0.27 | 3.3 |  |  | 44 | 39 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 47 |  |  |  |  |  |  |  |  |  |  |  |  |
| M | Pine | 51 | 6.40 | 39.8 | 62 |  | 0.15 | 1.8 |  |  | 52 | 58 |
| E | Birch | 21 | 0.36 | 14.8 | 3 |  | 0.01 | 0.1 |  |  | 16 | 4 |
| A | Beech | 11 | 0.46 | 23.2 | 4 | 0.02 | 0.2 |  |  | 0 | 2 | 5 |
| N | Oak | 199 | 6.52 | 20.4 | 52 | 0.10 | 0.9 |  |  | 1 | 1 | 5 |
|  | Total | 629 | 27.83 | 23.7 | 265 | 0.55 | 6.2 |  | $5 / 30 / 50$ | 114 | 105 | 110 |


| 0 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Spruce | 402 | 14.62 | 21.5 | 147 |  | 0.28 | 3.3 |  |  | 45 | 38 | 47 |
| M | Pine | 51 | 6.44 | 39.9 | 63 |  | 0.15 | 1.8 |  |  | 51 | 58 |
| A | Birch | 18 | 0.36 | 15.7 | 3 |  | 0.01 | 0.1 |  |  | 16 | 4 |
|  | Beech | 16 | 0.49 | 20.0 | 5 |  | 0.02 | 0.2 |  |  | 0 | 2 |
|  | Oak | 203 | 6.53 | 20.2 | 52 |  | 0.10 | 0.9 |  |  | 1 | 5 |
| Total | 691 | 28.44 | 22.9 | 269 |  | 0.56 | 6.3 |  | $5 / 30 / 50$ | 113 | 104 | 110 |

Appendix C, Table C20. Simulated initial stand characteristics per hectare according to the treatment $\mathbf{T}$ with $\mathbf{5 c m}$ lower target diameters as starting point for the 50 years simulation on all 48 sample plots

|  |  |  | BA <br> $\left[\mathrm{m}^{2}\right]$ | dg <br> $[\mathrm{cm}]$ | Vol <br> $\left[\mathrm{m}^{3}\right]$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| S | Spruce | 401 | 7.18 | 15.1 | 59 |
| T | Pine | 71 | 4.81 | 29.4 | 40 |
| A | Birch | 22 | 0.19 | 10.6 | 1 |
| R | Beech | 19 | 0.30 | 14.0 | 2 |
| T | Oak | 271 | 2.94 | 11.8 | 21 |
|  | Total | 783 | 15.43 | 15.8 | 124 |

Appendix C, Table C21. Stand characteristics, annual increment and harvest according to treatment $\mathbf{T}$ with 5 cm lower target diameters after 50 years simulation (figures per ha, based on stand age-independent single-tree ages, three different ingrowth scenarios and minimum $100 \mathbf{m}^{\mathbf{3}} /$ ha standing volume after cut)

|  |  | N | $\begin{gathered} \mathrm{BA} \\ {\left[\mathrm{~m}^{2}\right]} \end{gathered}$ | $\begin{gathered} \mathrm{dg} \\ {[\mathrm{~cm}]} \end{gathered}$ | $\begin{gathered} \text { Vol } \\ {\left[\mathrm{m}^{3}\right]} \end{gathered}$ | BA <br> MAI <br> [ $\mathrm{m}^{2}$ ] | Vol <br> MAI $\left[\mathrm{m}^{3}\right]$ | Time of cutting [years] | 1th cut removal $\left[\mathrm{m}^{3}\right]$ | 2nd cut removal $\left[\mathrm{m}^{3}\right]$ | 3rd cut removal $\left[\mathrm{m}^{3}\right]$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Spruce | 324 | 9.72 | 19.6 | 90 | 0.27 | 3.0 |  | 69 | 98 |  |
| M | Pine | 6 | 0.51 | 33.9 | 4 | 0.08 | 0.9 |  | 104 | 74 |  |
| 1 | Birch | 8 | 0.14 | 15.1 | 1 | 0.01 | 0.0 |  | 20 | 1 |  |
| N | Beech | 6 | 0.20 | 20.2 | 2 | 0.02 | 0.2 |  | 1 | 9 |  |
|  | Oak | 178 | 6.20 | 21.1 | 47 | 0.14 | 1.2 |  | 0 | 26 |  |
|  | Total | 521 | 16.78 | 20.2 | 145 | 0.51 | 5.3 | 0/40 | 194 | 209 |  |



|  | Spruce | 763 | 16.76 | 16.7 | 154 | 0.39 | 4.0 |  | 69 | 85 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| M | Pine | 10 | 0.98 | 36.1 | 9 | 0.07 | 0.8 |  | 104 | 67 |  |
| A | Birch | 33 | 0.34 | 11.4 | 2 | 0.01 | 0.1 |  | 20 | 1 |  |
| X | Beech | 35 | 0.51 | 13.6 | 4 | 0.02 | 0.2 |  | 1 | 7 |  |
|  | Oak | 226 | 7.44 | 20.5 | 57 | 0.15 | 1.2 |  | 0 | 19 |  |
|  | Total | 1067 | 26.04 | 17.6 | 226 | 0.64 | 6.2 | 0/35 | 194 | 179 |  |

Appendix D - Variation of simulated future basal area on plots


Appendix, figure D2. Box plot with median, $10^{\text {th }}, 25^{\text {th }}, 75^{\text {th }}$, and $90^{\text {th }}$ percentiles of the basal area on single plots at year $0,25,45$, and 50 of the simulation (treatment TS).


Appendix D, figure D3. Box plot with median, $10^{\text {th }}, 25^{\text {th }}, 75^{\text {th }}$, and $90^{\text {th }}$ percentiles of the basal area on single plots at year $0,25,45$, and 50 of the simulation (according to treatment TN).


Appendix D, figure D4. Box plot with median, $10^{\text {th }}, 25^{\text {th }}, 75^{\text {th }}$, and $90^{\text {th }}$ percentiles of the basal area on single plots at year $0,25,45$, and 50 of the simulation (according to treatment T with 5 cm larger target diameters).


Appendix, figure D7. Box plot with median, $10^{\text {th }}, 25^{\text {th }}, 75^{\text {th }}$, and $90^{\text {th }}$ percentiles of the basal area on single plots at year $0,25,45$, and 50 of the simulation (according to treatment $T$ with 5 cm reduced target diameters).

## Appendix E

Table E1. Yield table and cost projections for treatment T. Figures were calculated per hectare. 0\% interest rate.

|  |  | Stand before thinning |  |  | Removed |  |  |  |  |  |  | Net |  | Annual |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Time |  | BA | V |  | BA | V | MAI |  | Income | Cost | income | PV | revenue | NPV |
|  | [yrs.] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | [ $\mathrm{m}^{3}$ ] | Treatment | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] |
| Spruce | 0 | 489 | 13.8 | 129 | 50 | 4.8 | 52 |  |  |  |  |  |  |  |  |
| Pine | 0 | 164 | 15.4 | 144 | 55 | 6.9 | 69 |  |  |  |  |  |  |  |  |
| Broadleaves | 0 | 346 | 5.9 | 45 | 29 | 2.2 | 18 |  |  |  |  |  |  |  |  |
| Total | 0 | 999 | 35.2 | 318 | 134 | 13.9 | 140 |  | Harvest | 49518 | 14000 | 35518 | 35518 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 25 | 459 | 15.6 | 152 | 61 | 4.2 | 45 |  |  |  |  |  |  |  |  |
| Pine | 25 | 103 | 11.8 | 113 | 44 | 6.0 | 60 |  |  |  |  |  |  |  |  |
| Broadleaves | 25 | 287 | 6.6 | 51 | 16 | 0.9 | 8 |  |  |  |  |  |  |  |  |
| Total | 25 | 849 | 34.1 | 316 | 121 | 11.0 | 113 | 6.1 | Harvest | 39845 | 11300 | 28545 | 28545 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 45 | 423 | 16.2 | 163 | 62 | 4.7 | 53 |  |  |  |  |  |  |  |  |
| Pine | 45 | 57 | 7.3 | 71 | 36 | 5.1 | 51 |  |  |  |  |  |  |  |  |
| Broadleaves | 45 | 276 | 8.8 | 71 | 28 | 1.9 | 17 |  |  |  |  |  |  |  |  |
| Total | 45 | 756 | 32.3 | 305 | 126 | 11.7 | 121 | 5.9 | Harvest | 43290 | 12100 | 31190 | 31190 | 2185 | '95253 |

$\mathrm{N}=$ number of stands, BA = basal area, $\mathrm{V}=$ total volume over bark, $\mathrm{PV}=$ present value, NPV = net present value

Table E2. Yield table and cost projections for treatment TS. Figures were calculated per hectare. 0\% interest rate.

|  |  | Stand before thinning |  |  | Removed |  |  |  |  |  |  | Net |  | Annual |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Time |  | BA | V |  | BA | V | MAI |  | Income | Cost | income | PV | revenue | NPV |
|  | [yrs.] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | [ $\mathrm{m}^{3}$ ] | Treatment | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] |
| Spruce | 0 | 489 | 13.8 | 129 | 50 | 4.8 | 53 |  |  |  |  |  |  |  |  |
| Pine | 0 | 164 | 15.4 | 144 | 55 | 6.9 | 69 |  |  |  |  |  |  |  |  |
| Broadleaves | 0 | 346 | 5.9 | 45 | 29 | 2.2 | 19 |  |  |  |  |  |  |  |  |
| Total | 0 | 999 | 35.2 | 318 | 134 | 13.9 | 140 |  | Harvest | 49518 | 14000 | 30518 | 30518 |  |  |
|  | 0 |  |  |  |  |  |  |  | Soil prep. |  | 2500 |  |  |  |  |
|  | 0 |  |  |  |  |  |  |  | Cleaning |  | 2500 |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 25 | 641 | 15.8 | 151 | 63 | 4.2 | 45 |  |  |  |  |  |  |  |  |
| Pine | 25 | 191 | 12.0 | 113 | 44 | 6.0 | 60 |  |  |  |  |  |  |  |  |
| Broadleaves | 25 | 307 | 6.7 | 51 | 17 | 0.9 | 8 |  |  |  |  |  |  |  |  |
| Total | 25 | 1139 | 34.5 | 315 | 124 | 11.1 | 113 | 6.0 | Harvest | 39845 | 11300 | 23545 | 23545 |  |  |
|  | 25 |  |  |  |  |  |  |  | Soil prep. |  | 2500 |  |  |  |  |
|  | 25 |  |  |  |  |  |  |  | Cleaning |  | 2500 |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 45 | 555 | 16.6 | 165 | 63 | 4.5 | 51 |  |  |  |  |  |  |  |  |
| Pine | 45 | 137 | 7.8 | 75 | 38 | 5.2 | 51 |  |  |  |  |  |  |  |  |
| Broadleaves | 45 | 283 | 8.9 | 71 | 28 | 1.9 | 18 |  |  |  |  |  |  |  |  |
| Total | 45 | 975 | 33.3 | 311 | 129 | 11.6 | 120 | 6.1 | Harvest | 43290 | 12000 | 26290 | 26290 | 1607 | 80353 |
|  | 45 |  |  |  |  |  |  |  | Soil prep. |  | 2500 |  |  |  |  |
|  | 45 |  |  |  |  |  |  |  | Cleaning |  | 2500 |  |  |  |  |

[^2]Table E3. Yield table and cost projections for treatment TN. Figures were calculated per hectare. 0\% interest rate.

|  | Stand before thinning |  |  |  | Removed |  |  |  |  |  |  | Net |  | Annual |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Time |  | BA | V |  | BA | V | MAI |  | Income | Cost | income | PV | revenue | NPV |
|  | [yrs.] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | [ $\mathrm{m}^{3}$ ] | Treatment | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] |
| Spruce | 0 | 489 | 13.8 | 129 | 68 | 6.0 | 64 |  |  |  |  |  |  |  |  |
| Pine | 0 | 164 | 15.4 | 144 | 28 | 4.3 | 44 |  |  |  |  |  |  |  |  |
| Broadleaves | 0 | 346 | 5.9 | 45 | 17 | 1.6 | 13 |  |  |  |  |  |  |  |  |
| Total | 0 | 999 | 35.2 | 318 | 113 | 11.9 | 121 |  | Harvest | 30886 | 12100 | 18786 | 18786 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 20 | 436 | 12.7 | 119 | 84 | 5.1 | 53 |  |  |  |  |  |  |  |  |
| Pine | 20 | 131 | 14.4 | 136 | 42 | 6.0 | 60 |  |  |  |  |  |  |  |  |
| Broadleaves | 20 | 303 | 6.6 | 51 | 16 | 0.7 | 6 |  |  |  |  |  |  |  |  |
| Total | 20 | 870 | 33.7 | 306 | 142 | 11.7 | 119 | 6.0 | Harvest | 46843 | 11900 | 34943 | 34943 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 45 | 379 | 12.7 | 123 | 103 | 6.6 | 70 |  |  |  |  |  |  |  |  |
| Pine | 45 | 84 | 11.0 | 108 | 55 | 8.0 | 81 |  |  |  |  |  |  |  |  |
| Broadleaves | 45 | 280 | 9.7 | 80 | 11 | 0.4 | 4 |  |  |  |  |  |  |  |  |
| Total | 45 | 743 | 33.4 | 311 | 169 | 15.0 | 155 | 5.7 | Harvest | 59851 | 15500 | 44351 | 44351 | 1962 | 98080 |

$\mathrm{N}=$ number of stands, $\mathrm{BA}=$ basal area, $\mathrm{V}=$ total volume over bark, $\mathrm{PV}=$ present value, $\mathrm{NPV}=$ net present value

Table E4. Yield table and cost projections for the clearfelling strategy, planting of spruce on this site, and typical even-aged forest management with 60 years rotation. Figures calculated per hectare with the forest simulator DT (Nilsson \& Fahlvik, 2006). $r=0 \%$.

|  |  | Stand before thinning |  |  | Removed |  |  |  |  |  |  | Net |  | Annual |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Time |  | BA | V |  | BA | V | MAI |  | Income | Cost | income | PV | revenue | NPV |
|  | [yrs.] | N | [ $\mathrm{m}^{2}$ ] | $\left[\mathrm{m}^{3}\right]$ | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | [ $\mathrm{m}^{3}$ ] | Treatment | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] |
| Spruce | 0 | 489 | 13.8 | 129 | 489 | 13.8 | 129 |  |  | 32904 |  |  |  |  |  |
| Pine | 0 | 164 | 15.4 | 144 | 164 | 15.4 | 144 |  |  | 72079 |  |  |  |  |  |
| Broadleaves | 0 | 346 | 5.9 | 45 | 346 | 5.9 | 45 |  |  | 9223 |  |  |  |  |  |
| Total | 0 | 999 | 35.2 | 318 | 999 | 35.2 | 318 |  | Harvest | 114206 | 31800 | 64406 | 64406 |  |  |
|  | 0 |  |  |  |  |  |  |  | Soil prep. |  | 3000 |  |  |  |  |
|  | 0 |  |  |  |  |  |  |  | Planting |  | 15000 |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 7 |  |  |  |  |  |  |  | Cleaning |  | 1000 | -1000 | -1000 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 15 |  |  |  |  |  |  |  | Cleaning |  | 1500 | -1500 | $-1500$ |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 30 | 1783 | 27.5 | 165 | 796 | 10.7 | 64 | 5.5 | Thinning | 11954 | 8240 | 3714 | 3714 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 40 | 637 | 21.3 | 177 | 318 | 9.3 | 76 | 7.9 | Thinning | 19660 | 6251 | 13409 | 13409 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 60 | 573 | 40.4 | 454 | 573 | 40.4 | 454 | 9.9 | Harvest | 160584 | 15325 | 127259 | 127259 | 2365 | '206288 |
|  | 60 |  |  |  |  |  |  |  | Soil prep. |  | 3000 |  |  |  |  |
|  | 60 |  |  |  |  |  |  |  | Planting |  | 15000 |  |  |  |  |

$\mathrm{N}=$ number of stands, $\mathrm{BA}=$ basal area, $\mathrm{V}=$ total volume over bark, $\mathrm{PV}=$ present value, $\mathrm{NPV}=$ net present value

Table E5. Yield table and cost projections for treatment T. Figures were calculated per hectare. $4 \%$ interest rate.

|  |  | Stand before thinning |  |  | Removed |  |  |  |  |  |  | Net |  | Annual |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Time |  | BA | V |  | BA | V | MAI |  | Income | Cost | income | PV | revenue | NPV |
|  | [yrs.] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | [ $\mathrm{m}^{3}$ ] | Treatment | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] |
| Spruce | 0 | 489 | 13.8 | 129 | 50 | 4.8 | 52 |  |  |  |  |  |  |  |  |
| Pine | 0 | 164 | 15.4 | 144 | 55 | 6.9 | 69 |  |  |  |  |  |  |  |  |
| Broadleaves | 0 | 346 | 5.9 | 45 | 29 | 2.2 | 18 |  |  |  |  |  |  |  |  |
| Total | 0 | 999 | 35.2 | 318 | 134 | 13.9 | 140 |  | Harvest | 49518 | 14000 | 35518 | 35518 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 25 | 459 | 15.6 | 152 | 61 | 4.2 | 45 |  |  |  |  |  |  |  |  |
| Pine | 25 | 103 | 11.8 | 113 | 44 | 6.0 | 60 |  |  |  |  |  |  |  |  |
| Broadleaves | 25 | 287 | 6.6 | 51 | 16 | 0.9 | 8 |  |  |  |  |  |  |  |  |
| Total | 25 | 849 | 34.1 | 316 | 121 | 11.0 | 113 | 6.1 | Harvest | 39845 | 11300 | 28545 | 10708 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 45 | 423 | 16.2 | 163 | 62 | 4.7 | 53 |  |  |  |  |  |  |  |  |
| Pine | 45 | 57 | 7.3 | 71 | 36 | 5.1 | 51 |  |  |  |  |  |  |  |  |
| Broadleaves | 45 | 276 | 8.8 | 71 | 28 | 1.9 | 17 |  |  |  |  |  |  |  |  |
| Total | 45 | 756 | 32.3 | 305 | 126 | 11.7 | 121 | 5.9 | Harvest | 43290 | 12100 | 31190 | 5340 | 2185 | 「51565 |

$\mathrm{N}=$ number of stands, $\mathrm{BA}=$ basal area, $\mathrm{V}=$ total volume over bark, $\mathrm{PV}=$ present value, $\mathrm{NPV}=$ net present value

Table E6. Yield table and cost projections for treatment TS. Figures were calculated per hectare. $4 \%$ interest rate.

|  |  | Stand before thinning |  |  | Removed |  |  |  |  |  |  | Net |  | Annual |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Time |  | BA | V |  | BA | V | MAI |  | Income | Cost | income | PV | revenue | NPV |
|  | [yrs.] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | [ $\mathrm{m}^{3}$ ] | Treatment | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] |
| Spruce | 0 | 489 | 13.8 | 129 | 50 | 4.8 | 53 |  |  |  |  |  |  |  |  |
| Pine | 0 | 164 | 15.4 | 144 | 55 | 6.9 | 69 |  |  |  |  |  |  |  |  |
| Broadleaves | 0 | 346 | 5.9 | 45 | 29 | 2.2 | 19 |  |  |  |  |  |  |  |  |
| Total | 0 | 999 | 35.2 | 318 | 134 | 13.9 | 140 |  | Harvest | 49518 | 14000 | 30518 | 30518 |  |  |
|  | 0 |  |  |  |  |  |  |  | Soil prep. |  | 2500 |  |  |  |  |
|  | 0 |  |  |  |  |  |  |  | Cleaning |  | 2500 |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 25 | 641 | 15.8 | 151 | 63 | 4.2 | 45 |  |  |  |  |  |  |  |  |
| Pine | 25 | 191 | 12.0 | 113 | 44 | 6.0 | 60 |  |  |  |  |  |  |  |  |
| Broadleaves | 25 | 307 | 6.7 | 51 | 17 | 0.9 | 8 |  |  |  |  |  |  |  |  |
| Total | 25 | 1139 | 34.5 | 315 | 124 | 11.1 | 113 | 6.0 | Harvest | 39845 | 11300 | 23545 | 8832 |  |  |
|  | 25 |  |  |  |  |  |  |  | Soil prep. |  | 2500 |  |  |  |  |
|  | 25 |  |  |  |  |  |  |  | Cleaning |  | 2500 |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 45 | 555 | 16.6 | 165 | 63 | 4.5 | 51 |  |  |  |  |  |  |  |  |
| Pine | 45 | 137 | 7.8 | 75 | 38 | 5.2 | 51 |  |  |  |  |  |  |  |  |
| Broadleaves | 45 | 283 | 8.9 | 71 | 28 | 1.9 | 18 |  |  |  |  |  |  |  |  |
| Total | 45 | 975 | 33.3 | 311 | 129 | 11.6 | 120 | 6.1 | Harvest | 43290 | 12000 | 26290 | 4501 | 1607 | 43851 |
|  | 45 |  |  |  |  |  |  |  | Soil prep. |  | 2500 |  |  |  |  |
|  | 45 |  |  |  |  |  |  |  | Cleaning |  | 2500 |  |  |  |  |

[^3]Table E7. Yield table and cost projections for treatment TN. Figures were calculated per hectare. $4 \%$ interest rate.

|  |  | Stand before thinning |  |  | Removed |  |  |  |  |  |  | Net |  | Annual |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Time |  | BA | V |  | BA | V | MAI |  | Income | Cost | income | PV | revenue | NPV |
|  | [yrs.] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | [ $\mathrm{m}^{3}$ ] | Treatment | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] |
| Spruce | 0 | 489 | 13.8 | 129 | 68 | 6.0 | 64 |  |  |  |  |  |  |  |  |
| Pine | 0 | 164 | 15.4 | 144 | 28 | 4.3 | 44 |  |  |  |  |  |  |  |  |
| Broadleaves | 0 | 346 | 5.9 | 45 | 17 | 1.6 | 13 |  |  |  |  |  |  |  |  |
| Total | 0 | 999 | 35.2 | 318 | 113 | 11.9 | 121 |  | Harvest | 30886 | 12100 | 18786 | 18786 |  |  |
| Spruce | 20 | 436 | 12.7 | 119 | 84 | 5.1 | 53 |  |  |  |  |  |  |  |  |
| Pine | 20 | 131 | 14.4 | 136 | 42 | 6.0 | 60 |  |  |  |  |  |  |  |  |
| Broadleaves | 20 | 303 | 6.6 | 51 | 16 | 0.7 | 6 |  |  |  |  |  |  |  |  |
| Total | 20 | 870 | 33.7 | 306 | 142 | 11.7 | 119 | 6.0 | Harvest | 46843 | 11900 | 34943 | 15948 |  |  |
| Spruce | 45 | 379 | 12.7 | 123 | 103 | 6.6 | 70 |  |  |  |  |  |  |  |  |
| Pine | 45 | 84 | 11.0 | 108 | 55 | 8.0 | 81 |  |  |  |  |  |  |  |  |
| Broadleaves | 45 | 280 | 9.7 | 80 | 11 | 0.4 | 4 |  |  |  |  |  |  |  |  |
| Total | 45 | 743 | 33.4 | 311 | 169 | 15.0 | 155 | 5.7 | Harvest | 59851 | 15500 | 44351 | 7593 | 1962 | 42326 |

$\mathrm{N}=$ number of stands, $\mathrm{BA}=$ basal area, $\mathrm{V}=$ total volume over bark, $\mathrm{PV}=$ present value, $\mathrm{NPV}=$ net present value

Table E8. Yield table and cost projections for the clearfelling strategy, planting of spruce on this site, and typical even-aged forest management with 60 years rotation. Figures calculated per hectare with the forest simulator DT (Nilsson \& Fahlvik, 2006). $r=4 \%$.

|  |  | Stand before thinning |  |  | Removed |  | V | MAI |  | Income | Cost | Net income | PV | Annual revenue | NPV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Time |  | BA | V |  | BA |  |  |  |  |  |  |  |  |  |
|  | [yrs.] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | N | [ $\mathrm{m}^{2}$ ] | [ $\mathrm{m}^{3}$ ] | [ $\mathrm{m}^{3}$ ] | Treatment | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] | [SEK] |
| Spruce | 0 | 489 | 13.8 | 129 | 489 | 13.8 | 129 |  |  | 32904 |  |  |  |  |  |
| Pine | 0 | 164 | 15.4 | 144 | 164 | 15.4 | 144 |  |  | 72079 |  |  |  |  |  |
| Broadleaves | 0 | 346 | 5.9 | 45 | 346 | 5.9 | 45 |  |  | 9223 |  |  |  |  |  |
| Total | 0 | 999 | 35.2 | 318 | 999 | 35.2 | 318 |  | Harvest | 114206 | 31800 | 64406 | 64406 |  |  |
|  | 0 |  |  |  |  |  |  |  | Soil prep. |  | 3000 |  |  |  |  |
|  | 0 |  |  |  |  |  |  |  | Planting |  | 15000 |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 7 |  |  |  |  |  |  |  | Cleaning |  | 1000 | -1000 | -760 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 15 |  |  |  |  |  |  |  | Cleaning |  | 1500 | -1500 | -833 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 30 | 1783 | 27.5 | 165 | 796 | 10.7 | 64 | 5.5 | Thinning | 11954 | 8240 | 3714 | 1145 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 40 | 637 | 21.3 | 177 | 318 | 9.3 | 76 | 7.9 | Thinning | 19660 | 6251 | 13409 | 2793 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Spruce | 60 | 573 | 40.4 | 454 | 573 | 40.4 | 454 | 9.9 | Harvest | 160584 | 15325 | 127259 | 12097 | 2365 | 78849 |
|  | 60 |  |  |  |  |  |  |  | Soil prep. |  | 3000 |  |  |  |  |
|  | 60 |  |  |  |  |  |  |  | Planting |  | 15000 |  |  |  |  |

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[^1]:    Box 3. Prediction of future ingrowth of oak and other tree species
    The focus on spruce as the dominating tree species of ingrowth might be seen as too pessimistic, because future occurrence of other tree species can be underestimated. Especially oak and beech might have a certain potential with the indicated climatic changes (Hickler et al., 2009; Grundmann et al., 2011). However, heavy browsing damages to broadleaves as a general experience in south Sweden and as found in the stand gave reason for the chosen scenarios.
    The emergence of current oak trees is not completely clear. The Jay (Garrulus glandarius) is suspected for dispersal of acorns as described by Stähr et al. (2006). Accumulated over an unknown time period, Frost (1997) found 54 oak seedlings per ha on average in coniferous forest. On mesic, central European sites with average nutrient supply, Mosandl and Kleinert (1998) and Kätzel et al. (2005) reported 500-2000 oak seedlings per ha accumulated over time in pine stands. Bilke (2005) found 300 seedlings per ha in 100-200 m distances from 100 years old oak trees. He found considerable more individuals if the area was fenced.
    More conceptually, Reif \& Gärtner (2007) assume for natural oak forests, "that long-lasting phases without successful regeneration of young oaks change with phases of successful establishment of a new oak generation", depending on open canopy and low ground vegetation. More information about oak in southern Sweden can be found in Drössler et al. (2012).
    The prediction of tree species proportions at the stage of regeneration establishment is rather difficult. A naturally very high variation of determining factors (seed production, dispersal, germination conditions and diseases, climate conditions, soil moisture, browsing etc.) make it hard to predict the actual composition of future tree regeneration. Experienced surprises from natural regeneration experiments intending to establish particular tree species demonstrated how difficult the prediction of regeneration pattern in a particular stand can be (i.e. experiment MB99 Skogaby). After establishment, height growth and survival of established regeneration are still highly stochastic processes compared to single tree growth predictions. Unfortunately, there is no empirical data known by the authors beside the given references above to forecast other tree species. In the study stand, the initial proportion of spruce seedlings was ten times higher than oak. In addition, spruce was least susceptible to browsing, and the second most shade-tolerant tree species in the stand (after beech). Regarding the difficulties of prediction, the most appropriate interpretation approach seems to refer to total future seedling number assumed in the three scenarios without differentiation between the tree species

[^2]:    $\mathrm{N}=$ number of stands, $\mathrm{BA}=$ basal area, $\mathrm{V}=$ total volume over bark, $\mathrm{PV}=$ present value, $\mathrm{NPV}=$ net present value

[^3]:    $N=$ number of stands, $B A=$ basal area, $V=$ total volume over bark, $P V=$ present value, $N P V=$ net present value

[^4]:    $\mathrm{N}=$ number of stands, $\mathrm{BA}=$ basal area, $\mathrm{V}=$ total volume over bark, $\mathrm{PV}=$ present value, $\mathrm{NPV}=$ net present value

