

N₂O emissions from plants are reduced under photosynthetic activity

Klaus Schützenmeister¹ | Katharina H. E. Meurer^{1,2} | Marco Gronwald³ | Antonia B. D. Hartmann¹ | Dirk Gansert⁴ | Hermann F. Jungkunst¹

¹iES Landau, Institute for Environmental Sciences, University Koblenz-Landau, Landau, Germany

²Department of Soil & Environment, Swedish University of Agricultural Sciences-SLU, Uppsala, Sweden

³Osnabrueck University, Osnabrück, Germany

⁴Plant Ecology and Ecosystem Research, Georg-August University Göttingen, Göttingen, Germany

Correspondence

Katharina H. E. Meurer, Department of Soil & Environment, Swedish University of Agricultural Sciences-SLU, Lennart Hjelms väg 9, 750 07 Uppsala, Sweden.

Email: meurer-katharina@uni-landau.de; katharina.meurer@slu.se

Funding information

Ministry of Science and Culture Lower Saxony; Niedersächsisches Vorab

Abstract

- New plant functions in the exchange of greenhouse gases between ecosystems and atmosphere have recently been discovered. We tested whether photosynthetic activity has an effect on N₂O emission rates from incubated plant-soil systems.
- Two laboratory experiments were performed. One to unravel possible effect of photosynthetic activity on the net N₂O ecosystem exchange for two species (beech and ash saplings). The other to account for possible effects from rhizosphere and aboveground plant parts separately (ash sapling only).
- Total N₂O emissions from both plant and plant-soil systems were significantly lower under light than in darkness (31%–65%). The photosynthetic effect only applied to the aboveground plant parts.
- Underlying processes have now to be unraveled to improve our understanding of ecosystem functioning. This will improve modeling and budgeting of greenhouse gas exchanges between ecosystems and the atmosphere.

KEY WORDS

N₂O, photosynthesis, ash (*Fraxinus excelsior* L.), beech (*Fagus sylvatica* L.), rhizosphere, plant-mediated

1 | INTRODUCTION

Present research points to the importance of ecosystems to counteract human-enforced climate change, because under sufficient nitrogen (N) availability plants are capable to partly fix carbon dioxide (CO₂) emitted by human activity (Ainsworth & Long, 2005; Ainsworth & Rogers, 2007). Furthermore, interaction of plants and soils involved in GHG cycling are by-passes of the oxic sink zone in soils, for example, methane (CH₄) release to the atmosphere via aerenchyma (Li, Zhu, Bao, Wang, & Xu, 2016; Mosier, Mohanty, Bhadrachalam, & Chakravorti, 1990). This, however, mostly accounts for wetland

plants, while most upland plants lack aerenchyma tissue. The effect of plants on N₂O fluxes has received some attention in recent years and mainly shows how plants can consistently emit N₂O even without aerenchyma (e.g., Chen, Boeckx, Shen, & Van Cleemput, 1999; Díaz-Pinés et al., 2016; Goshima et al., 1999; Lenhart et al., 2015; Pihlatie, Ambus, Rinne, Pilegaard, & Vesala, 2005; Rochester, Wood, & Macdonald, 2015; Rueckauf, Augustin, Russow, & Merbach, 2004; Wen, Corre, Rachow, Chen, & Veldkamp, 2017; Zou, Huang, Sun, Zheng, & Wang, 2005). Moreover, there are some studies showing that plant leaves can actively emit N₂O (Cheng, Sakai, Nishimura, Yagi, & Hasegawa, 2010; Hakata, Takahashi, Zumft, Sakamoto, &

This is an open access article under the terms of the Creative Commons Attribution License, which permits use, distribution and reproduction in any medium, provided the original work is properly cited.

© 2020 The Authors. *Journal of Plant-Environment Interactions* published by New Phytologist and John Wiley & Sons Ltd

Morikawa, 2003; Lenhart et al., 2019; Machacova et al., 2016; Smart & Bloom, 2001). According to Zou et al. (2005), there are two mechanisms for the efflux of N_2O from plants: N_2O is either derived from the soil and transported by the plant or directly produced by the plant itself during N assimilation. So far, it is not always clear which of the mechanisms is dominant or whether they occur simultaneously under field conditions. Nevertheless, the transport of dissolved N_2O through the plant happens via the transpiration stream (Chang, Janzen, Cho, & Nakonechny, 1998; Díaz-Pinés et al., 2016; Pihlatie et al., 2005; Yu, Wang, & Chen, 1997) and the efflux can be from the plant stomata (Zou et al., 2005) and stem lenticels (Díaz-Pinés et al., 2016; McBain, Warland, McBride, & Wagner-Riddle, 2004). However, most laboratory studies in that research area solely focussed on the plants and ignored the contribution of soil. Nowadays, most measurements of N_2O from plant-soil continua are derived from closed chambers with opaque walls (e.g., Jungkunst, Freibauer, Neufeldt, & Bareth, 2006; Kesik et al., 2005; Meurer et al., 2016). If the plants are small enough to fit into the chamber, then photosynthesis is not included as a potential driving force. Considering adult forest trees, measurements were done on the ground or the stem (Barba, Poyatos, & Vargas, 2019; Díaz-Pinés et al., 2016; Machacova et al., 2016; Wen et al., 2017), inevitably excluding the forest canopy and therewith possible effects of photosynthesis on N_2O fluxes. Additionally to their crucial role in buffering anthropogenic CO_2 emissions, forests and forest soils have been shown to be of high importance for N_2O exchange with the atmosphere (Butterbach-Bahl & Kiese, 2005; Kesik et al., 2005; Reay, Dentener, Smith, Grace, & Feely, 2008; Stocker et al., 2013). However, an influence of photosynthesis was hardly even considered possible as the influence of abiotic factors such as soil temperature, bulk density, pH value, and soil moisture are much more evident. The exclusion of photosynthesis appears justifiable because photosynthetic carbon assimilation by leaves is remote from the process of N_2O release by soil microorganisms to the atmosphere. Though Smart and Bloom (2001) showed that N_2O can be emitted by wheat leaves during photosynthesis and Bruhn, Albert, Mikkelsen, and Ambus (2014) found that natural UV irradiation caused the ecosystem N_2O emission to be ~30% higher than otherwise assumed using dark chambers as usual. Eddy covariance measurements always include the effect of photosynthesis but we are only aware of a few eddy covariance measurement that revealed diurnal N_2O patterns of correlations to gross primary production (e.g., Zona et al., 2013). In general, a light-dependent gas transport or N_2O production mechanisms have been suggested in the literature (Jørgensen, Struwe, & Elberling, 2012), meaning that changes in illumination can directly affect N_2O effluxes (Yu & Chen, 2009). In an extensive study including 32 plant species, Lenhart et al. (2019) found a strong relation between CO_2 respiration and N_2O effluxes, which was consistent over a broad range of changing environmental conditions (temperature and N supply). Though, they could not confirm the effect of light in their study.

Investigations of tree-mediated N_2O fluxes are rare and mostly restricted to seedlings and/or saplings under laboratory conditions (e.g., Machacova, Papen, Kreuzwieser, & Rennenberg, 2013; Pihlatie

et al., 2005; Rusch & Rennenberg, 1998). So far, the highly likely influence of soil on plant-derived N_2O emission has hardly been considered and CO_2 and N_2O emissions from plants were measured under sterile conditions or using nutrient solutions to substitute the soil. Yet, the primary biogenic N_2O sources are from soils (70%) and involve the microbial N transformations brought about by nitrification and denitrification (Mosier et al., 1998). Therefore, we intended to make the next step by measuring N_2O emissions from trees in soil. This setup includes the competition for nutrients with microorganisms, explaining why net N_2O emissions are lower from plant-soil systems than from soil alone. Despite the fact that N_2O is obviously emitted by plants and it is therefore likely that photosynthetic activity will have some effect on the emissions, we hypothesized that photosynthetic activity will not have a relevant impact on net ecosystem N_2O emissions. We expected the impacts to be small (<10%) and close or below detection limits. Therefore, field measurement campaigns could still neglect the photosynthesis effect.

2 | METHODS

2.1 | Net ecosystem experiment

2.1.1 | Plant and soil material

The soil used for the soil column experiment (stagnic Luvisol) was gathered from a mixed deciduous broad-leaved forest in the Hainich National Park, Thuringia, Germany ($51^{\circ}04'N$ $10^{\circ}30'E$). The soil was sampled from the upper 10 cm (A_h -horizon) and homogenized by passing it through a 2 mm mesh sieve. Ash (*Fraxinus excelsior* L.) and beech (*Fagus sylvatica* L.) saplings (3- to 6-years old) of approximately identical biomass (plant height 15–20 cm) were sampled in the same forest. The choice of the two tree species is based on the fact that they are the two most common broad-leaved tree species in Central Europe and relatives to both species occur throughout the temperate zone. Moreover, ash and beech have differences in litter quality and water balance, which is why we expected them to behave differently throughout the experiment.

90 days after the beginning of the experiment, the soil was fertilized with 25 kg KNO_3 /ha. This fertilization was done to ensure sufficient N supply for N_2O emissions, as fluxes decelerate in the mesocosms if not fertilized. Another KNO_3 addition of 100 kg/ha was done 63 days later. The return of organic matter (mainly C) to the soil was simulated by applying 100 ml of a dissolved carbon (DOC) solution (5 g/L powdered ash litter). This C:N fertilization assured identical starting conditions between the treatments (C:N 11.8 with 1.82% organic carbon (C_{org})). To determine the concentrations of nitrate (NO_3^-) and ammonium (NH_4^+), soil samples were analyzed with the continuous flow injection colorimetry (SAN + Continuous Flow Analyzer, Skalar Instruments) at two dates within the experiment (Table 1). Nitrate was determined with the copper-cadmium-reduction method (ISO 13395), and NH_4^+ with the Berthelot reaction method (ISO 11732).

TABLE 1 Soil properties of the soil used in the column experiment. NO_3^- and NH_4^+ contents are given as means \pm SE

Treatment	Sand g/kg	Silt	Clay	pH_{KCl}	Prior to 1st fertilization		End of experiment	
					NO_3^- mg/L	NH_4^+ mg/L	NO_3^- mg/L	NH_4^+ mg/L
Ash (<i>Fraxinus excelsior</i>)	2.9	56.5	40.6	5.3	6.0 \pm 1.4	0.1 \pm 0.0	6.2 \pm 0.0	0.2 \pm 0.0
Beech (<i>Fagus sylvatica</i>)					6.3 \pm 2.1	0.1 \pm 0.0	9.4 \pm 0.0	0.1 \pm 0.0
Bare soil					11.3 \pm 2.9	0.2 \pm 0.0	11.2 \pm 2.0	0.3 \pm 0.1

2.2 | Experimental setup

Fifteen acrylic glass cylinders ($h = 50$ cm; diameter, $d = 17$ cm) were filled with 5 kg of freshly sieved (2 mm) soil and planted with the saplings. The cylinders were transparent in order to enable photosynthetic active radiation (PAR) to pass. This resulted in the three treatments: ash, beech, and bare soil. The soil columns were placed randomly in a greenhouse with a steady air temperature of 20°C and a relative air humidity of 80%. Illumination (12 hr) was maintained by lamps (PAR: $203 \pm 10 \mu\text{mol m}^{-2} \text{s}^{-1}$ PPF; Eye Lighting, Clean Ace). The water-filled pore space (WFPS) of each column was adjusted to 75%–80% and controlled once a week. The setup allowed the determination of net CO_2 uptake (plant assimilation) under illuminated conditions and dark CO_2 emission (all related to the soil surface area). The measurements were performed under (a) PAR conditions and (b) under dark conditions using a black cloth. The order of light before dark measurements was changed every day to avoid a bias by the order of dark to light measurements. Measurements were performed biweekly at 8 a.m. The experiment ran for 183 days in total and N_2O fluxes were low and close to the resolution of the gas chromatograph in all treatments. To ensure the accuracy of the data and be able to detect differences between the treatments, the dataset used in this study has been reduced to 14 samplings and a timespan of 58 days (35 days after the first fertilization; 25 kg KNO_3/ha).

2.2.1 | Trace gas sampling

We collected headspace gas samples (30–32 cm above soil surface) at 0, 10, and 20 min, after chamber closing using 60 ml syringes. The gas concentrations were analysed with an auto-sample, computer-controlled (Probe 64+1, V1.31) gas chromatograph (Shimadzu GC-14B). N_2O was detected by a ^{63}Ni electron capture detector. A linear regression was used to calculate the increase or decrease in gas concentrations, following Lessard, Rochette, Gregorich, Desjardins, and Pattey (1997). The influence of illumination was identified by the differences between N_2O fluxes under light (PAR) and dark conditions, whereby negative Δ indicated a reduction in N_2O fluxes under PAR conditions. The gross plant assimilation (ΔCO_2) was calculated as $\Delta \text{CO}_2 = \text{CO}_2 \text{ light}$

(photosynthesis)– CO_2 dark (respiration). The relationship between $\Delta \text{N}_2\text{O}$ and ΔCO_2 was tested by linear regression after correcting the dataset for outliers by considering each data point below or above 1.5 times the interquartile range as being too far from the central values to be reasonable. This led to exclusion of a total of 5 data points for the ash and 3 data points for the beech treatment. To generalize our findings, in the further, Δ will be expressed as a percentage reduction, that is, as $((\text{PAR-dark})/\text{dark}) * 100$.

2.3 | Day-night test

To negate a possible diurnal trend in the effect of photosynthesis, we used the ash treatment (columns 11, 12, 14, 16, and 17) for a one-time 24-hr measurement campaign with samplings in 3-hr intervals. WFPS was adjusted to 75%–80% and calculated for every gas sampling time, based on the column weights after the experiment and the assumption that WFPS decreased linearly during the experiment by evapotranspiration. As it had been shown that the sequence of the measurements did not have an impact on the N_2O fluxes, samplings were first made under exclusion of photosynthesis and afterwards during photosynthesis. Gas samples were taken exactly as described above (see *Trace gas sampling*), that is, air was sampled via syringes and samples were analyzed in the gas chromatograph.

2.4 | Plant flux experiment

To be able to negate that the observed reduction in N_2O emissions was primarily soil-driven, we measured emissions from above- and belowground parts of ash saplings separately in a shorter experiment (10 samplings in total with bi-daily measurements). In this experiment, we only added a very small amount of mineral soil to mimic the rhizosphere. For that reason, the experimental period was kept to 10 days.

2.4.1 | Experimental setup

Soil material was filled into six acrylic cylinders ($h = 10$ cm, $d = 5$ cm) and planted with ash saplings (height 8–14 cm). To separate the

belowground and aboveground plant parts, an adjusted lid with a gap for the stem and a tube for belowground gas extraction was on top of the acrylic cylinder (Figure 1). The lid was sealed airtight to the cylinder. To avoid light penetration and, consequently, the growth of further biomass, for example, algae, in the space between soil and lid (3.0–3.6 cm), the acrylic cylinder was wrapped in aluminum foil. The ash saplings were placed randomly in a greenhouse with 20°C and 9.5 hr illumination per day. For gas measurements, the ash saplings were put into the columns used in the previous experiment ($h = 50$ cm; $d = 17$ cm; Figure 1) and gas samples for N_2O were taken manually and using syringes and stored in 12.5 ml Labco vials until analysis at the gas chromatograph. A total number of three gas samples were taken over a period of 30 min (at 0, 15, and 30 min). Gas samplings were taken every other day and fluxes were calculated using linear regression (Lessard et al., 1997). The measurements started one day after all ash saplings had been fertilized with 200 kg N/ha as KNO_3 . Seven days after the fertilization, the ash saplings were irrigated until the soil became waterlogged. Measurements were done under (a) PAR conditions and (b) dark conditions. Cumulative fluxes were calculated by first linearly interpolating between the days and summarizing over the experimental period.

Correction for outliers, that is, data points below or above 1.5 times the interquartile range, led to the exclusion of 15 and 11 data points for the aboveground and rhizosphere treatment, respectively. Just like the net ecosystem experiment, $\Delta\text{N}_2\text{O}$ is expressed as $(\text{dark-PAR})/\text{dark}) \times 100$.

2.5 | Statistical analyses

Statistical analyses were done in R (R Core Team, 2018) and data were processed using the *openxlsx* (Walker, 2019) and *plyr* (Wickham, 2011) packages. Figures were made with the *ggplot* (*ggplot2* package; Wickham, 2016) and *plot_grid* functions (*cowplot* package; Wilke, 2019).

The design of all experiments allowed for pairwise statistics because identical tree-soil systems ($n = 5$) were measured in darkness and thereafter again under light conditions. To avoid a systematic error, the order of light and dark measurements was switched from each measurement day to the other. Differences in N_2O fluxes between treatments or diurnal and nocturnal periods (Day-Night test) were calculated using the Wilcoxon test for paired samples (paired = T, alternative = "greater"). Differences were regarded significant for $p \leq .05$.

For the Day-Night test, the relationship between N_2O fluxes and WFPS was tested using linear regression (*lm* function from the *stats* package).

3 | RESULTS AND DISCUSSION

Light conditions had a reducing effect on observed N_2O emissions for both net ecosystem fluxes from ash as well as beech and respective average fluxes were 65% and 34% lower compared to observations made under dark conditions. In both cases, the reduction (Δ) over the 58 days of measurements was significantly different from zero ($p < .01$). For the bare soil treatment, Δ was positive (5%) but no significant differences were found between light and dark measurements ($p = .72$). The average N_2O emissions observed throughout the individual experiments are shown in Table 2.

The metabolism of these tree species causes a strong reduction of net ecosystem N_2O fluxes. Most likely, the N-metabolism of the plant is involved, but C shortage by reduced rhizo-deposition and water changes in the rhizosphere may also influence N_2O emission from the soil. Considering only the rhizosphere of ash, as was done in the plant flux experiment, observed total N_2O emissions were negative and this uptake was higher under PAR compared to dark conditions ($\Delta = 104\%$, $p = .17$; Figure 2b). A strong reduction of N_2O fluxes by PAR was found from the shoots ($\Delta = -66\%$; $p = .05$), indicating that the photosynthetic effect mainly applies to the aboveground plant parts.



FIGURE 1 Setup for gas measurements from belowground (left) and aboveground plant parts under light (middle) and dark (right) conditions

TABLE 2 Average N_2O fluxes [$\mu\text{g}/\text{m}^2$] under dark and PAR conditions observed during the three experiments. For the net ecosystem experiment, both the daily average and average fluxes over the 2-month period are presented. The data presented in this study originate from a running experiment and cover a period of 58 days (35 days after the first fertilization with 25 kg KNO_3/ha) and includes 14 measurements. For the 24-hr measurements, average values for each hour of the day (HOD) are shown. For the plant experiment, the cumulative fluxes over the 10-day period are presented (see also Figure 2). Δ stands for the $\text{N}_2\text{O}_{\text{PAR}} - \text{N}_2\text{O}_{\text{dark}}$). Different letters represent significant differences between the treatments

Time	Dark	PAR	Δ	Dark	PAR	Δ	Dark	PAR	Δ
	<i>Bare soil</i>			<i>F. excelsior</i>			<i>F. sylvatica</i>		
Net ecosystem experiment									
<i>Sampling</i> <i>Daily measurements</i>									
1	61 (65)	85 (56)	24 (26)	23 (19)	6 (20)	-17 (13)	29 (31)	-1 (33)	-30 (14)
2	241 (225)	237 (209)	-4 (23)	19 (17)	11 (11)	-8 (11)	99 (27)	68 (31)	-31 (7)
3	84 (80)	75 (66)	-9 (20)	13 (9)	-9 (11)	-22 (13)	40 (23)	15 (31)	-25 (14)
4	253 (246)	272 (244)	19 (32)	19 (16)	0 (11)	-19 (8)	122 (94)	113 (128)	-10 (34)
5	43 (56)	38 (41)	-5 (18)	11 (7)	-18 (15)	-29 (19)	14 (8)	-17 (11)	-30 (4)
6	188 (213)	165 (200)	-23 (20)	27 (24)	-6 (16)	-34 (26)	36 (23)	6 (25)	-30 (15)
7	40 (23)	44 (39)	4 (27)	23 (33)	-2 (14)	-25 (20)	27 (27)	3 (25)	-23 (17)
8	184 (186)	170 (173)	-14 (51)	154 (279)	112 (236)	-43 (43)	98 (80)	68 (76)	-30 (25)
9	69 (65)	65 (50)	-4 (17)	12 (13)	-14 (14)	-26 (13)	32 (24)	10 (21)	-22 (16)
10	278 (256)	294 (269)	16 (28)	102 (71)	51 (53)	-52 (22)	164 (142)	142 (136)	-22 (12)
11	66 (56)	67 (53)	1 (18)	17 (9)	-23 (8)	-40 (12)	69 (69)	39 (59)	-29 (32)
12	200 (169)	200 (173)	1 (21)	109 (108)	71 (87)	-38 (25)	323 (358)	248 (263)	-74 (118)
13	95 (144)	43 (44)	-52 (112)	84 (139)	37 (101)	-47 (179)	39 (40)	20 (42)	-19 (14)
14	72 (78)	203 (285)	131 (275)	33 (23)	14 (41)	-19 (47)	98 (82)	68 (104)	-30 (36)
<i>Average fluxes</i>									
	134 (162)	140 (172)	6 (85)	46 (94)	16 (78)	-30 (50)	85 (129)	56 (112)	-29 (37)
<i>cumulative fluxes</i>									
	1873 (1689)	1958 (1547)	84 (340)	646 (368)	228 (253)	-418 (280)	1,188 (793)	782 (756)	-406 (184)
<i>HOD</i> <i>24-hr measurements</i>									
5 a.m.				29 (13)	13 (17)	-15 (8)			
8 a.m.				45 (15)	23 (17)	-30 (4)			
11 a.m.				37 (25)	15 (20)	-29 (13)			
2 p.m.				40 (24)	18 (23)	-25 (6)			
5 p.m.				50 (31)	20 (24)	-23 (13)			
8 p.m.				44 (29)	36 (27)	-28 (9)			
11 p.m.				60 (38)	36 (37)	-23 (8)			
2 a.m.				48 (42)	31 (38)	-17 (9)			
<i>Plant flux experiment</i>									
<i>F. excelsior</i>									
	<i>Aboveground</i>				<i>Rhizosphere</i>				
	1,494 (1517)	508 (665)	-986 (852)		-50 (408)	-100 (355)	-51 (91)		

Generally, the cumulative N_2O efflux from planted soil was lower than from bare soil and showed species-specific differences confirming previous findings (e.g., Fender, Leuschner, Schützenmeister, Gansert, & Jungkunst, 2013; Fender et al., 2012). However, it mattered whether measurements were performed in darkness or under light. Ash caused a pronounced reduction of 66% in the dark and even 88% during the photosynthetic period in the light compared

to bare soil. For beech, the corresponding decrease was less pronounced but still 37% in the dark and 60% in the light (Figure 2a). If photosynthetic N-assimilation alone were responsible for the decrease in N_2O efflux, this effect should cease in the dark, reaching similar rates as the bare soil. However, in the dark, both tree species revealed highly relevant reductions of N_2O emissions compared to the bare soil, whereas ash showed a stronger effect than beech.

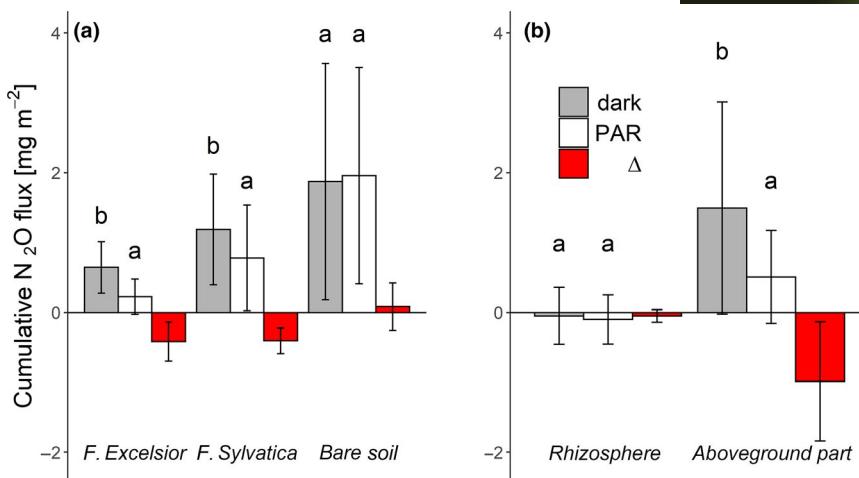
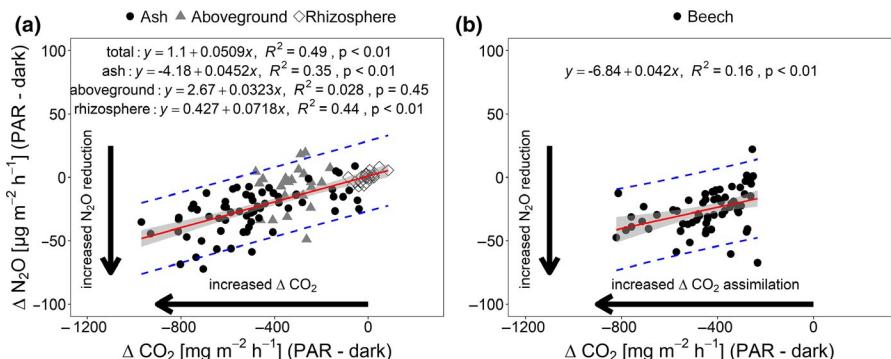


FIGURE 2 Mean net cumulative N₂O fluxes under dark and light conditions and N₂O reduction (Δ) from (a) ash, beech, and bare soil after the measuring period of 58 days and (b) from rhizosphere and aboveground biomass of ash after 4 days of measurements. Bars represent the standard deviation (SD). Different letters indicate significant differences ($p < .05$) between dark and PAR conditions. Differences between tree species (*Fraxinus excelsior*, *Fagus sylvatica*, and bare soil) and plant parts (rhizosphere and aboveground parts), respectively, were not significant ($p > .05$).

FIGURE 3 Relationship between plant CO₂ assimilation (Δ CO₂) and N₂O reduction (Δ N₂O) for ash (a) and beech (b). The linear regression line (red), the 95% confidence interval (gray-shaded area), and the prediction interval (dashed blue lines) are shown. $N_{\text{ash}} = 64$, $N_{\text{aboveground}} = 23$, $N_{\text{tot}} = 87$, $N_{\text{beech}} = 64$



Since NO₃⁻ uptake by plant roots was found to start with a lag time of 4 hr after transition from dark to light (Delhon, Gojon, Tillard, & Passama, 1996), it appears unlikely that N uptake by plants has an immediate effect on N₂O emission from the soil. The short-term experiment rather shows that aboveground plant parts may have a direct effect on N₂O emissions. Both ash and beech with soil in the net ecosystem experiment showed a strong negative correlation between plant CO₂ assimilation and decrease in N₂O efflux (Figure 3a,b). However, a closer look at the aboveground plant part during the plant experiment revealed no such correlation (Figure 3a).

Photosynthesis has an instantaneous and species-specific effect on the reduction of N₂O emissions from the soil. To prove whether this effect might be a function of the trees' photoperiodism, we measured N₂O efflux rates of eight dark-light intervals throughout a diurnal course over 24 hr. The amount of the photosynthetically induced reduction of N₂O efflux was influenced by the time of the day (Figure 4), and Δ was higher during the diurnal (~56%) compared to the nocturnal (~34%) period ($p = .01$). The reducing effect of photosynthesis was apparent throughout the day and Δ was significantly different from zero ($p = .03$) for all but the last measurements ($p = .06$; Table 2; Figure 4). An observed overall trend of increasing

N₂O efflux can be related to the decreasing water content in the soil columns during the 24-hr experimental period ($R^2 = 0.84$; $p < .01$ and $.93$; $p < .01$, in the dark and in the light, respectively), which most likely explains the diurnal effect best.

This study provides evidence that tree photosynthesis can have a considerable and instantaneous reducing effect on N₂O emission from ecosystems. This effect was highly repeatable in our experiments, and tree species-specific differences were found. The effect was persistent independent of the order of dark before light or dark after light measurement. The plant-induced reduction of N₂O efflux under light may hold key information in understanding the underlying processes driven by N assimilation of tree roots during darkness. Our results showed that the photosynthetic effect does not apply to the rhizosphere but primarily to the aboveground plant parts. However, the correlation between N₂O reduction and CO₂ assimilation (Figure 3) of the net ecosystem experiment suggests that there might be an effect on the rhizosphere that has not come to light in the experiments presented here. In contrast to leaves, root-specific plasma membrane-bound nitrate reductase (PM-NR) is not down-regulated in the dark so that apoplastic reduction of NO₃⁻ to NO₂⁻ can take place at similar rates day and night (Duncanson, Ip, Sherman, Kirk, & Wray, 1992; Eick & Stöhr, 2012; Stöhr & Mäck, 2012). In the apoplast,

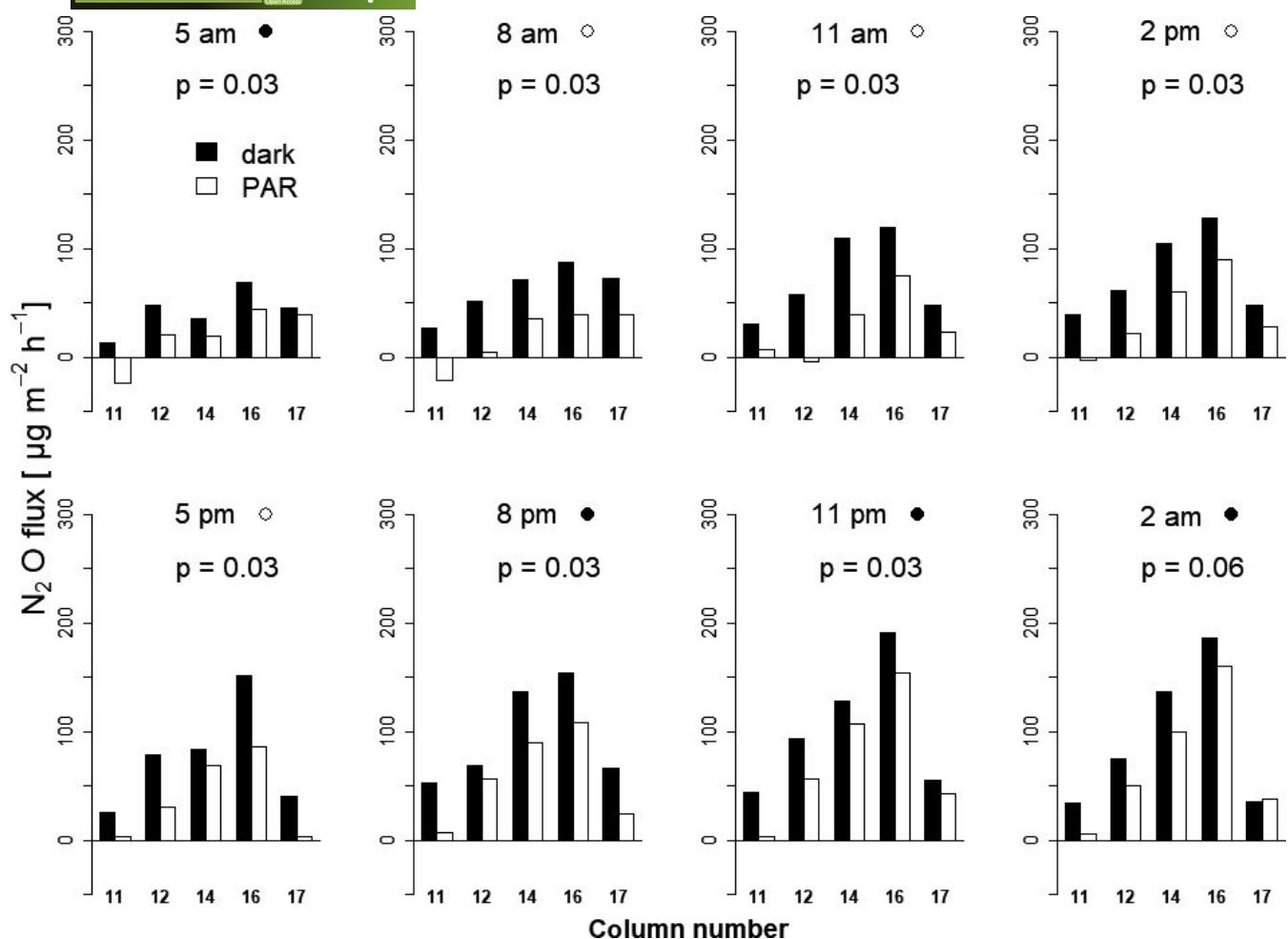


FIGURE 4 N₂O fluxes ($\mu\text{g m}^{-2} \text{hr}$) of each ash planted column under dark and PAR conditions. ●, nighttime, ○, daytime

toxic levels of nitrite are avoided by further metabolism to NO via a plasma membrane-bound nitrite/NO reductase (NI-NOR), which is spatially associated to PM-NR (Delhon et al., 1996). Whether the symplastic pathway of NO_3^- reduction to NH_4^+ in the roots also contributes to NO_3^- uptake in the dark requires further investigations. Nevertheless, the ATP and reductant that are required for the N assimilation during which N₂O is being produced may be provided by photosynthesis, respiration or both—in any case does N assimilation depend on increased respiratory carbon flow (Turpin, Weger, & Huppe, 1997). As summarized by Huppe and Turpin (1994), in the presence of an alternative carbon source, the inhibition of NO_3^- assimilation can be overcome. This has been demonstrated by darkened photosynthetic tissues of higher plants, showing that N₂O production during N assimilation also occurs in the absence of plant photosynthesis. This finding has been confirmed by Zou et al. (2005) who found a correlation between N₂O emissions from plants and a plant respiratory coefficient, indicating that N₂O emissions from plants might be associated with plant respiration. Maintaining a high but physiologically tolerable level of apoplastic NO_2^- in the dark appears to provide the physiological prerequisite for the correlation between increasing photosynthesis and decreasing denitrification in the light because nitrite reductase (NIR) in the chloroplast is not substrate

limited, particularly not after the onset of light. Moreover in relation to the soil, plant-mediated denitrification (Delhon et al., 1996) in the dark is of minor significance, if at all, because otherwise a decrease in N₂O efflux would not occur. The coupling of N metabolizing processes in the dark and in the light is ecologically meaningful, because in nature, ash is distinguished by a strong growth under light-limited conditions. The fact that beech also caused a considerable reduction of N₂O emission in the light, but with a weaker, albeit significant, correlation between photosynthesis and N₂O decrease, gives rise to the assumption that, in the dark, NO_3^- reduction in beech roots might be less pronounced than in ash. The lower reduction of N₂O emission from beech-planted columns in the dark supports this assumption.

We are aware of the fact that the observed N₂O reduction potential under illuminated conditions of 65% by ash and 31% by beech does not allow for global projections yet, but we consider these results as an important step into further investigations on the disentanglement of N metabolizing pathways in plants and its coupling with N₂O release from the soil. This will considerably improve our understanding of the temporal dynamics of N₂O emissions. Studies on N₂O emissions using eddy covariance technique are still scarce, especially for forest sites, but diurnal patterns of N₂O emissions have been found from a poplar plantation (Zona

et al., 2013). Our results moreover show that our assumption of a non-relevant effect on net ecosystem fluxes does not apply. The importance of considering photosynthesis in N₂O budgeting has been suggested before. According to Bruhn et al. (2014), ecosystem N₂O emissions were 30% higher under natural UV radiation compared to darkness. Mueller (2003) stated that emissions could be even twice as high compared with dark chambers.

Investigations on N₂O fluxes from shoots of adult trees are difficult and measurements in the laboratory are mainly performed using saplings, while field measurements focus on the stems (Barba et al., 2019; Díaz-Pinés et al., 2016; Machacova et al., 2013, 2016; Wen et al., 2017). Even though the importance of the latter as a N₂O source is not clear and stems were found to be both a source (Díaz-Pinés et al., 2016; Machacova et al., 2013, 2016; Wen et al., 2017) and a sink (Barba et al., 2019), excluding the contribution of trees to N₂O exchange with the atmosphere might result in a systematic underestimation of the total forest ecosystem fluxes (Machacova et al., 2016).

Certainly, the influence of plants on the exchange of N₂O between terrestrial ecosystems and the atmosphere is not limited to the competition with soil microorganisms for N without species-specific differences. Obviously, the interaction between root-mediated N-metabolism and photosynthetic N-assimilation of plants is still poorly understood and our knowledge on terrestrial ecosystem-atmosphere exchanges of N₂O is to be expanded. Our results proved us partly wrong and have shown that we cannot predict ecosystem net fluxes by measuring plant-derived and soil-derived N₂O fluxes separate from each other.

ACKNOWLEDGMENTS

The study was funded by the Ministry of Science and Culture of Lower Saxony and the "Niedersächsisches Vorab."

CONFLICTS OF INTEREST

The authors declare no conflict of interest.

AUTHORS' CONTRIBUTIONS

K.S. and H.F.J. conceptualized the lab studies; K.S., H.F.J., and D.G. designed the lab studies; K.S., M.G., K.H.E.M., and A.B.D.H. performed the studies, analyzed the samples and data; K.S., K.H.E.M., M.G., D.G., A.B.D.H., and H.F.J. wrote the paper. All authors discussed the results and commented on the manuscript.

ORCID

Katharina H. E. Meurer  <https://orcid.org/0000-0002-8880-9650>
 Marco Gronwald  <https://orcid.org/0000-0002-5081-4390>
 Hermann F. Jungkunst  <https://orcid.org/0000-0002-9807-9401>

REFERENCES

- Ainsworth, E. A., & Long, S. P. (2005). What have we learned from 15 years of free-air CO₂ enrichment (FACE)? A meta-analytic review of the responses of photosynthesis, canopy properties and plant production to rising CO₂. *New Phytologist*, 165, 351–372. <https://doi.org/10.1111/j.1469-8137.2004.01224.x>
- Ainsworth, E. A., & Rogers, A. (2007). The response of photosynthesis and stomatal conductance to rising CO₂: Mechanisms and environmental interactions. *Plant Cell and Environment*, 30, 258–270. <https://doi.org/10.1111/j.1365-3040.2007.01641.x>
- Barba, J., Poyatos, R., & Vargas, R. (2019). Automated measurements of greenhouse gases fluxes from tree stems and soils: Magnitudes, patterns and drivers. *Scientific Reports*, 9, 4005. <https://doi.org/10.1038/s41598-019-39663-8>
- Bruhn, D., Albert, K. R., Mikkelsen, T. N., & Ambus, P. (2014). UV-induced N₂O emission from plants. *Atmospheric Environment*, 99, 206–214. <https://doi.org/10.1016/j.atmosenv.2014.09.077>
- Butterbach-Bahl, K., & Kiese, R. (2005). Significance of forests as sources for N₂O and NO. In D. Binkley & O. Menyailo (Eds.), *Tree species effects on soils: Implications for global change* (pp. 173–191). The Netherlands: Springer.
- Chang, C., Janzen, H. H., Cho, C. M., & Nakonechny, E. M. (1998). Nitrous oxide emission through plants. *Soil Science Society of America Journal*, 62, 35–38. <https://doi.org/10.2136/sssaj1998.03615995006200010005x>
- Chen, X., Boeckx, P., Shen, S., & Van Cleemput, O. (1999). Emission of N₂O from rye grass (*Lolium perenne* L.). *Biology and Fertility of Soils*, 28, 393–396. <https://doi.org/10.1007/s003740050510>
- Cheng, W., Sakai, H., Nishimura, S., Yagi, K., & Hasegawa, T. (2010). The lowland paddy weed *Monochoria vaginalis* emits N₂O but not CH₄. *Agriculture, Ecosystems and Environment*, 137, 219–221. <https://doi.org/10.1016/j.agee.2010.01.011>
- Delhon, P., Gojon, A., Tillard, P., & Passama, L. (1996). Diurnal regulation of NO₃⁻ uptake in soybean plants. IV. Dependence on current photosynthesis and sugar availability to the roots *Journal of Experimental Botany*, 47(300), 893–900.
- Díaz-Pinés, E., Heras, P., Gasche, R., Rubio, A., Rennenberg, H., Butterbach-Bahl, K., & Kiese, R. (2016). Nitrous oxide emissions from stems of ash (*Fraxinus angustifolia* Vahl) and European beech (*Fagus sylvatica* L.). *Plant & Soil*, 398, 35–45. <https://doi.org/10.1007/s11104-015-2629-8>
- Duncanson, E., Ip, S.-M., Sherman, A., Kirk, D. W., & Wray, J. L. (1992). Synthesis of nitrite reductase is regulated differently in leaf and root of barley (*Hordeum vulgare* L.). *Plant Science*, 87, 151–160. [https://doi.org/10.1016/0168-9452\(92\)90146-D](https://doi.org/10.1016/0168-9452(92)90146-D)
- Eick, M., & Stöhr, C. (2012). Denitrification by plant roots? New aspects of plant plasma membrane-bound nitrate reductase. *Protoplasma*, 249, 909–918. <https://doi.org/10.1007/s00709-011-0355-5>
- Fender, A. C., Leuschner, C., Schützenmeister, K., Gansert, D., & Jungkunst, H. F. (2013). Rhizosphere effects of tree species – Large reduction of N₂O emission by saplings of ash, but not of beech, in temperate forest soil. *European Journal of Soil Biology*, 54, 7–15. <https://doi.org/10.1016/j.ejsobi.2012.10.010>
- Fender, A. C., Pfeiffer, B., Gansert, D., Jungkunst, H. F., Fiedler, S., Beyer, F., ... Leuschner, C. (2012). Root-induced tree species effects on the source/sink strength for greenhouse gases (CH₄, N₂O and CO₂) of a temperate deciduous forest soil. *Soil Biology and Biochemistry*, 57, 587–597. <https://doi.org/10.1016/j.soilbio.2012.08.004>
- Goshima, N., Mukai, T., Suemori, M., Takahashi, M., Caboche, M., & Morikawa, H. (1999). Emission of nitrous oxide (N₂O) from transgenic tobacco expressing antisense NiR mRNA. *The Plant Journal*, 19(1), 75–80. <https://doi.org/10.1046/j.1365-313X.1999.00494.x>
- Hakata, M., Takahashi, M., Zumft, W., Sakamoto, A., & Morikawa, H. (2003). Conversion of the nitrate nitrogen and nitrogen dioxide to nitrous oxide in plants. *Acta Biotechnologica*, 23, 249–257.
- Huppe, H. C., & Turpin, D. H. (1994). Integration of carbon and nitrogen metabolism in plant and algae cells. *Annual Review of Plant Physiology and Plant Molecular Biology*, 45, 577–607.
- Jørgensen, C. J., Struwe, S., & Elberling, B. (2012). Temporal trends in N₂O flux dynamics in a Danish wetland – effects of plant-mediated gas transport of N₂O and O₂ following changes in water level and soil

- mineral-N availability. *Global Change Biology*, 18, 210–222. <https://doi.org/10.1111/j.1365-2486.2011.02485.x>
- Jungkunst, H. F., Freibauer, A., Neufeldt, H., & Bareth, G. (2006). Nitrous oxide emissions from agricultural land use in Germany – a synthesis of available annual field data. *Journal of Plant Nutrition and Soil Science*, 169, 341–351. <https://doi.org/10.1002/jpln.200521954>
- Kesik, M., Ambus, P., Baritz, R., Brüggemann, N., Butterbach-Bahl, K., Damm, M., ... Leip, A. (2005). Inventories of N_2O and NO emissions from European forest soils. *Biogeoscience*, 2, 353–375. <https://doi.org/10.5194/bg-2-353-2005>
- Lenhart, K., Behrendt, T., Greiner, S., Steinkamp, J., Well, R., Giesemann, A., & Keppler, F. (2019). Nitrous oxide effluxes from plants as a potentially important source to the atmosphere. *New Phytologist*, 221, 1398–1408. <https://doi.org/10.1111/nph.15455>
- Lenhart, K., Weber, B., Elbert, W., Steinkamp, J., Clough, T., Crutzen, P., ... Keppler, F. (2015). Nitrous oxide and methane emissions from cryptogamic covers. *Global Change Biology*, 21, 3889–3900. <https://doi.org/10.1111/gcb.12995>
- Lessard, R., Rochette, P., Gregorich, E. G., Desjardins, R. L., & Pattey, E. (1997). CH_4 fluxes from a soil amended with dairy cattle manure and ammonium nitrate. *Canadian Journal of Soil Science*, 77, 179–186.
- Li, F., Zhu, R., Bao, T., Wang, Q., & Xu, H. (2016). Sunlight stimulates methane uptake and nitrous oxide emission from the High Arctic tundra. *Science of the Total Environment*, 572, 1150–1160. <https://doi.org/10.1016/j.scitotenv.2016.08.026>
- Machacova, K., Bäck, J., Vanhatalo, A., Halmeenmäki, E., Kolari, P., Mammarella, I., ... Pihlati, M. (2016). *Pinus sylvestris* as a missing source of nitrous oxide and methane in boreal forest. *Scientific Reports*, 6, 23410. <https://doi.org/10.1038/srep23410>
- Machacova, K., Papen, H., Kreuzwieser, J., & Rennenberg, H. (2013). Inundation strongly stimulates nitrous oxide emissions from stems of the upland tree *Fagus sylvatica* and the riparian tree *Alnus glutinosa*. *Plant & Soil*, 364, 287–301. <https://doi.org/10.1007/s11104-012-1359-4>
- McBain, M. C., Warland, J. S., McBride, R. A., & Wagner-Riddle, C. (2004). Laboratory-scale measurements of N_2O and CH_4 emissions from hybrid poplars (*Populus deltoids* × *Populus nigra*). *Waste Management & Research*, 22, 454–465.
- Meurer, K. H. E., Franko, U., Stange, C. F., Dalla Rosa, J., Madari, B. E., & Jungkunst, H. F. (2016). Direct nitrous oxide (N_2O) fluxes from soils under different land use in Brazil – A critical review. *Environmental Research Letters*, 11, 023001. <https://doi.org/10.1088/1748-9326/11/2/023001>
- Mosier, A., Kroeze, C., Nevison, C., Oenema, O., Seitzinger, S., & van Cleemput, O. (1998). Closing the global N_2O budget: Nitrous oxide emissions through the agricultural nitrogen cycle. *Nutrient Cycling in Agroecosystems*, 52, 225–248.
- Mosier, A. R., Mohanty, S. K., Bhadrachalam, A., & Chakravorti, S. P. (1990). Evolution of dinitrogen and nitrous oxide from the soil to the atmosphere through rice plants. *Biology and Fertility of Soils*, 9, 61–67. <https://doi.org/10.1007/BF00335863>
- Mueller, C. (2003). Plants affect the *in situ* N_2O emissions of temperate grassland ecosystem. *Journal of Plant Nutrition and Soil Science*, 166, 771–773.
- Pihlatie, M., Ambus, P., Rinne, J., Pilegaard, K., & Vesala, T. (2005). Plant-mediated nitrous oxide emissions from beech (*Fagus sylvatica*) leaves. *New Phytologist*, 168, 93–98. <https://doi.org/10.1111/j.1469-8137.2005.01542.x>
- R Core Team. (2018). *R: A language and environment for statistical computing*. Vienna, Austria: R Foundation for Statistical Computing. Retrieved from <https://www.R-project.org/>
- Reay, D. S., Dentener, F., Smith, P., Grace, J., & Feely, R. A. (2008). Global nitrogen deposition and carbon sinks. *Nature Geoscience*, 1, 430–437. <https://doi.org/10.1038/ngeo230>
- Rochester, I., Wood, C., & Macdonald, B. (2015). Quantifying nitrous oxide emissions from the foliage of cotton, maize and soybean crops. *Crop and Pasture Science*, 66, 689–695. <https://doi.org/10.1071/CP14301>
- Rueckauf, U., Augustin, J., Russow, R., & Merbach, W. (2004). Nitrate removal from drained and reflooded fen soils affected by soil N transformation processes and plant uptake. *Soil Biology & Biochemistry*, 36, 77–90. <https://doi.org/10.1016/j.soilbio.2003.08.021>
- Rusch, H., & Rennenberg, H. (1998). Black alder (*Alnus Glutinosa* (L.) Gaertn) trees mediate methane and nitrous oxide emission from the soil to the atmosphere. *Plant & Soil*, 201, 1–7.
- Smart, D. R., & Bloom, A. J. (2001). Wheat leaves emit nitrous oxide during nitrate assimilation. *Proceedings of the National Academy of Science*, 98(14), 7875–7878. <https://doi.org/10.1073/pnas.131572798>
- Stocker, T. F., Qin, D., Plattner, G.-K., Tignor, M., Allen, S. K., Boschung, J., ... Midgley, P. M. (2013). Climate change 2013: The physical science basis. *Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on climate change*. Cambridge, UK: Cambridge University Press.
- Stöhr, C., & Mäck, G. (2012). Diurnal changes in nitrogen metabolism of tobacco roots. *Journal of Experimental Botany*, 52, 1283–1289.
- Turpin, D. H., Weger, H. G., & Huppe, H. C. (1997). Interactions between photosynthesis, respiration and nitrogen assimilation. In D. T. Dennis, D. H. Turpin, & D. D. Lefebvre (Eds.), *Plant metabolism* (pp. 509–524). Edinburgh Gate, Harlow: Addison Wesley Longman.
- Walker, A. (2019). *openxlsx: Read, write and edit XLSX files*. R package version 4.1.0.1. Retrieved from <https://CRAN.R-project.org/package=openxlsx>
- Wen, Y., Corre, M. D., Rachow, C., Chen, L., & Veldkamp, E. (2017). Nitrous oxide emissions from stems of alder, beech and spruce on a temperate forest. *Plant & Soil*, 420, 423–434. <https://doi.org/10.1007/s11104-017-3416-5>
- Wickham, H. (2011). The split-apply-combine strategy for data analysis. *Journal of Statistical Software*, 40(1), 1–29. Retrieved from <http://www.jstatsoft.org/v40/i01/>
- Wickham, H. (2016). *ggplot2: Elegant graphics for data analysis*. New York, NY: Springer-Verlag. Retrieved from <https://ggplot2.tidyverse.org>
- Wilke, C. O. (2019). *cowplot: Streamlined plot theme and plot annotations for "ggplot2"*. R package version 1.0.0. Retrieved from <https://CRAN.R-project.org/package=cowplot>
- Yu, K., & Chen, G. (2009). Nitrous oxide emissions from terrestrial plants: observations, mechanisms and implications. In A. I. Sheldon & E. P. Barnhart (Eds.), *Nitrous oxide emissions research progress* (pp. 85–104).
- Yu, K. W., Wang, Z. P., & Chen, G. X. (1997). Nitrous oxide and methane transport through rice plants. *Biology and Fertility of Soils*, 24(3), 341–343. <https://doi.org/10.1007/s003740050254>
- Zona, D., Janssens, I. A., Gioli, B., Jungkunst, H. F., Marta, C. S., & Ceulemans, R. (2013). N_2O fluxes of a bio-energy poplar plantation during a two years rotation period. *Global Change Biology – Bioenergy*, 5, 536–547.
- Zou, J., Huang, Y., Sun, W., Zheng, X., & Wang, Y. (2005). Contribution of plants to N_2O emissions in soil-winter wheat ecosystem: Pot and field experiments. *Plant & Soil*, 269, 205–211. <https://doi.org/10.1007/s11104-004-0484-0>

How to cite this article: Schützenmeister K, Meurer KHE, Gronwald M, Hartmann ABD, Gansert D, Jungkunst HF. N_2O emissions from plants are reduced under photosynthetic activity. *Plant-Environment Interactions*. 2020;1:48–56. <https://doi.org/10.1002/pei3.10015>