

# Bryosphere Loss Impairs Litter Decomposition Consistently Across Moss Species, Litter Types, and Micro-Arthropod Abundance

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## ABSTRACT

The bryosphere (that is, ground mosses and their associated biota) is a key driver of nutrient and carbon dynamics in many terrestrial ecosystems, in part because it regulates litter decomposition. However, we have a poor understanding of how litter decomposition responds to changes in the bryosphere, including changes in bryosphere cover, moss species, and bryosphere-associated biota. Specifically, the contribution of micro-arthropods to litter decomposition in the bryosphere is unclear. Here, we used a 16-month litterbag field experiment in two boreal forests to investigate bryosphere effects on litter decomposition rates among two moss species (*Pleurozium schreberi* and *Hylocomium splendens*), and two litter types (higher-quality *Betula pendula* litter and lower-quality *P. schreberi* litter). Additionally, we counted all micro-arthropods in the litterbags and identified them to functional groups. We found that bryosphere removal reduced litter decomposition rates by 28% and micro-arthropod abundance by 29% and led to a colder micro-climate. Litter decomposition rates

and micro-arthropod abundance were uncorrelated overall, but were positively correlated in *B. pendula* litterbags. Bryosphere effects on litter decomposition rates were consistent across moss species, litter types, and micro-arthropod abundances and community compositions. These findings suggest that micro-arthropods play a minor role in litter decomposition in the boreal forest floor, suggesting that other factors (for example, micro-climate, nutrient availability) likely drive the positive effect of the bryosphere on decomposition rates. Our results point to a substantial and consistent impairment of litter decomposition in response to loss of moss cover, which could have important implications for nutrient and carbon cycling in moss-dominated ecosystems.

**Key words:** collembola; mites; moss; litter mass loss; litterbag; litter quality; meso-fauna; boreal forest; *Pleurozium schreberi*; *Hylocomium splendens*.

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## HIGHLIGHTS

- Moss removal impaired litter decomposition and micro-arthropod abundance
- Moss effects on decomposition were unrelated to

micro-arthropod abundance

- Moss species and litter quality did not alter moss removal effects on decomposition

## INTRODUCTION

Mosses are particularly abundant in high latitude regions, where they often dominate the ground layer of peatlands, boreal forests, and tundra (Nilsson and Wardle 2005; Turetsky and others 2010; Street and others 2013). The bryosphere, that is, the ground bryophyte layer including senesced moss and its associated food web (Lindo and Gonzalez 2010), plays a key role in controlling carbon inputs to ecosystems due to its high net primary productivity, production of recalcitrant litter, and nutrient supply from moss-associated N-fixing cyanobacteria (Turetsky and others 2010; Lindo and others 2013; Street and others 2013). Moreover, the bryosphere can also control rates of ecosystem carbon loss through effects on microclimate (Gornall and others 2007), which in turn affect biological activity and thus litter decomposition (De Long and others 2016). Given that moss cover in high-latitude ecosystems is susceptible to changes in climate (Elmendorf and others 2012; Alatalo and others 2020) and land-use (Vitt and others 2019), understanding how litter decomposition may respond to changes in the bryosphere is needed to more accurately predict the effect of environmental change on carbon cycling.

In boreal forests, the bryosphere often promotes litter decomposition relative to bare ground (Jackson and others 2011, 2013; De Long and others 2016). The proposed mechanism is that the bryosphere improves environmental conditions for the decomposer community through regulating temperature and moisture and increasing nutrient supply (Jackson and others 2011). Importantly, the bryosphere can host high abundances of micro-arthropods, especially mites and springtails (Lindo and Gonzalez 2010; Glime 2017), and loss of bryosphere cover has been observed to decrease micro-arthropod abundance in the ground of alpine heaths (Mitchell and others 2016) and boreal forests (Salmane and Brumelis 2008; Bokhorst and others 2014). Micro-arthropods can enhance litter decomposition directly by feeding on detritus (Maraun and others 2011; Magilton and others 2019) and indirectly by stimulating microbes (which are the primary agents of decomposition) through litter fragmentation and faecal production (Briones 2014; Coleman and others 2017). While some studies have reported neutral effects (Wall

and others 2008; Barreto and Lindo 2018; Zhou and others 2020), numerous studies have shown that micro-arthropods can have important positive effects on litter decomposition (Schädler and Brandl 2005; Wall and others 2008; Wang and others 2009; Makkonen and others 2012; García-Palacios and others 2013; Fujii and others 2018). Therefore, increased micro-arthropod abundance can potentially contribute to the positive effects of the bryosphere on litter decomposition. However, little is known about the importance of micro-arthropod abundance to litter decomposition in boreal forests (Makkonen and others 2012; García-Palacios and others 2013). Moreover, whether and how the micro-arthropods impact on litter decomposition in the presence versus absence of a bryosphere remains unexplored.

The two feather moss species *Pleurozium schreberi* and *Hylocomium splendens* often dominate the forest floor in well-drained boreal forests (Nilsson and Wardle 2005). Physiological and morphological differences among moss species determine microhabitat structure and climate in the bryosphere, and this could influence the activity of the decomposer community. For example, patches dominated by *P. schreberi* are denser and better able to hold moisture than are patches dominated by *H. splendens* (Elumeeva and others 2011), thus potentially making them capable of supporting a greater abundance of micro-arthropods. Further, litter quality may mediate how micro-arthropods affect decomposition rates (García-Palacios and others 2013), and micro-arthropods may have a greater influence on decomposition of lower-quality litter in the bryosphere environment because lability of low-quality litter is further enhanced by soil fauna through the physical and chemical changes that they cause via litter fragmentation and partial digestion (Yang and Chen 2009; Milcu and Manning 2011; Joly and others 2020). Therefore, consideration of moss species identity and litter quality is needed to understand micro-arthropod effects on decomposition across a wider range of habitats, such as those varying in their types of bryospheres and litter inputs.

Here, we used a litterbag field experiment to investigate the effect of the bryosphere and of micro-arthropods on litter decomposition in the boreal forest floor. For this, we placed litterbags containing each of two types of litter (that is, higher-quality birch [*Betula pendula*] or lower-quality moss [*P. schreberi*] litter) in bryosphere plots dominated by either *P. schreberi* or *H. splendens*, and in adjacent plots where the bryosphere had been

removed. We then measured micro-arthropod abundance in the litterbags and litter mass loss after 4 and 16 months. We hypothesised that: (1) Compared to bare ground, litter in the bryosphere has greater abundance of micro-arthropods and decomposes faster. This is because of improved micro-climatic conditions (Bokhorst and others 2014; De Long and others 2016); (2) bryosphere effects are greater in *P. schreberi* than in *H. splendens* bryospheres. This is because *P. schreberi* has a higher water holding capacity (Elumeeva and others 2011), and thus, decomposition is less moisture-limited; and (3) bryosphere effects are greater for lower-quality (moss) litter than for higher-quality (birch) litter, because lability of low-quality litter is further enhanced by micro-arthropods (Yang and Chen 2009), which are more abundant in bryosphere-covered plots. By examining the effect of bryosphere removal on micro-arthropod abundance and decomposition of two litters of contrasting quality, and across two widespread feather moss species, we aimed to better understand how the bryosphere controls litter decomposition in boreal forests.

## METHODS

### Study Site

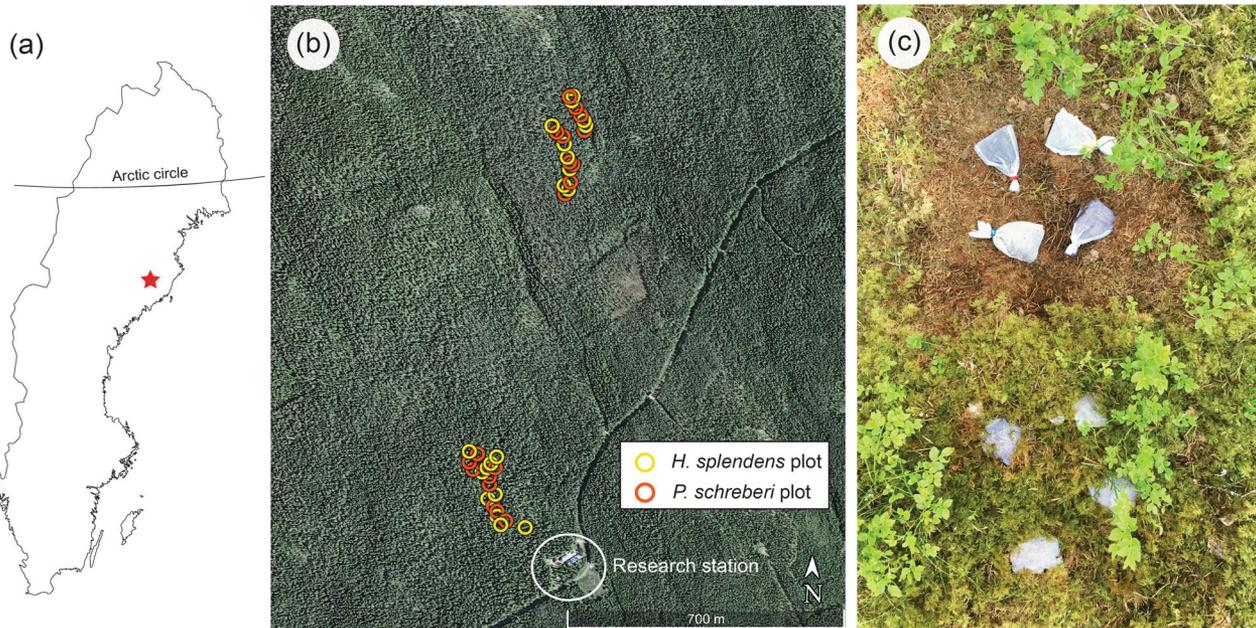
The study was carried out in Svartberget experimental forest, a mixed coniferous forest in northern Sweden (64°15' N, 19°46' E, 270 m a.s.l.; Figure 1a). The mean annual air temperature is 1.8 °C (mean temperature in January is – 10.3 °C and, in July, 15.3 °C), the mean annual precipitation is 620 mm, and the ground is covered by snow for about 170 days/year (data period 1991–2019; Svartberget Research Station 2020). We used two nearby sites located about 1 km apart. The southernmost site (64°14'43.83" N, 19°45'42.95" E) is a 110-year-old forest dominated by *Picea abies* (80%) with some *Pinus sylvestris* (20%), and the northernmost site (64°15'12.84" N, 19°45'56.37" E) is a 120-year-old forest composed almost exclusively of *Pinus sylvestris* (Figure 1b). *Betula* spp. is present at both sites. The understory vascular plant vegetation is similar at both sites and is overall dominated by *Vaccinium myrtillus* (11% cover), *Vaccinium vitis-idaea* (4%) and graminoids (1%). Continuous, mono-specific layers of the feather mosses *Pleurozium schreberi* and *Hylocomium splendens* are ubiquitous at both sites. The feather moss *Ptilium crista-castrensis* and the acrocarpous moss *Dicranum* sp. are also common but rarely dominate. *P. schreberi* and *H. splendens* bryospheres

had similar thickness (that is,  $7.2 \pm 0.4$  cm [mean  $\pm$  SE] and  $7.7 \pm 0.4$  cm, respectively), but *P. schreberi* bryospheres were denser (that is,  $16.9 \pm 1.6$  mg cm<sup>-3</sup> versus  $11.3 \pm 0.6$  mg cm<sup>-3</sup>, dry weight basis).

### Experimental Design and Litterbag Deployment

We used the litterbag method to estimate decomposition rates from litter mass loss. To test the effect of different litter qualities on decomposition, we used two litter types of contrasting lability: senesced *Betula pendula* leaves (higher-quality litter; C:N ratio =  $49.0 \pm 0.9$  [mean  $\pm$  SE], N = 6 independent sets of leaves), and senesced *P. schreberi* shoots (lower-quality litter; C:N ratio =  $70.2 \pm 1.4$ , N = 6 independent sets of shoots). The choice of litter types was based on previous research showing that litter quality (including N content) and mass loss values differed greatly between *B. pendula* and *P. schreberi* (Wardle and others 2003). Both litter types were collected on 12 May 2019 (that is, shortly after snowmelt) from the forest floor in Umeå (63° 50' N, 20° 20' E), about 50 km south-east of the study site. We selected freshly senesced *B. pendula* leaves (that is, those produced in the most recent autumn) on the forest floor. For *P. schreberi* litter, we harvested an intermediate 3–4 cm brown section of the shoot which was clearly separate from the upper green part and the lower, highly decomposed part, in line with Wardle and others (2003). Each of the two litter types was separately dried at 60 °C in a fan-assisted oven and homogenised. Nylon mesh litterbags (10 × 5 cm) were then filled with 1.7 g (measured to three decimal places) of oven-dry litter (either *B. pendula* or *P. schreberi* litter). Some litter fragmentation occurred during the process of filling the litterbags with whole leaves or shoots. To achieve a wider range of micro-arthropod abundances among the litterbags, we used litterbags of two different mesh sizes: 1.3 × 1.1 mm, and 1.0 × 0.2 mm (Bokhorst and Wardle 2013). Although both mesh sizes allow micro-arthropod access, the greater physical impediment in the small size litterbags could potentially influence the micro-arthropod community through slowing colonisation or excluding large individuals. In total, we prepared 320 litterbags (that is, 80 for each combination of two mesh sizes and two litter types).

On 25 May 2019, eight litterbags were deployed in each of 40 randomly chosen plots (20 in each site) where the bryosphere was dominated (that is, cover > 90%) by either *P. schreberi* (10 plots in



**Figure 1.** **a** Location of the study site in northern Sweden (map source: Vemaps.com). **b** Distribution of the plots in the two sites (image source: Google Earth Pro). **c** Example of a *Hylocomium splendens* plot. Each plot was divided in two subplots, one where the bryosphere had been removed and one where the bryosphere had not been removed. In each subplot, there were four litterbags, that is, a combination of two litter types (birch leaves or moss shoots), and two mesh sizes ( $1.3 \times 1.1$  mm, or  $1.0 \times 0.2$  mm).

each site) or by *H. splendens* (10 plots in each site). To minimise litter loss during transport, the litterbags were previously wetted using a spray bottle. At each plot, we placed four of the eight litterbags (that is, the factorial combination of two litter types and two mesh sizes) on a subplot where the bryosphere was not altered, and four of the eight litterbags on an adjacent  $40 \times 40$  cm subplot where the bryosphere had been removed (Figure 1c). The bryosphere was removed by hand, taking care to minimise disturbance to other plants and to the surface humus, on 22 May 2019. In bryosphere-covered subplots, the litterbags were embedded in the centre of the bryosphere profile. To do this, we manually parted some moss shoots and placed the litterbags flat about 2–4 cm above the humus surface (Figure 1c). Litter intercepted by the bryosphere is first decomposed in the upper and middle bryosphere before reaching the humus interface later in the decomposition process (Startsev and others 2008; De Long and others 2016). Thus, the placement of our litterbags in the centre of the bryosphere profile is intended to mimic conditions typical of the early stages of litter decomposition, in line with the relatively short-term nature of the study (De Long and others 2016). We used metal pins to ensure the litterbags remained embedded in the moss layer or made good contact with the soil.

At each plot, we installed two temperature loggers ( $N = 80$ ; iButtons, Maxim Integrated) on 27 July 2019, that is, one in each bryosphere-covered subplot and one in each bryosphere-removed subplot. The loggers were placed underneath the finer-mesh moss litterbag and set to record the temperature every 4 h over the entire incubation period. For each logger, we calculated mean daily temperature and daily range for each of four observation periods: spring (March–May), summer (June–August), autumn (September–November), and the snow-covered period between 19 November 2019 and 10 May 2020. We then calculated the change in mean daily temperature (and, separately, in daily range) due to bryosphere removal as the mean daily temperature (and daily range) in bryosphere-removed subplots minus mean daily temperature (and daily range) in bryosphere-covered subplots, for each observation period and for each plot.

### Litter Mass Loss and Micro-Arthropod Abundance

We collected the litterbags from half of the plots (that is, 10 *P. schreberi* plots and 10 *H. splendens* plots from each site) between 26 September 2019 and 7 October 2019 (that is, about 4 months after the

litterbags were deployed). Because our micro-arthropod extraction system could only fit 60 litterbags at a time, the 160 litterbags were collected in three sets of 48 to 56 litterbags (comprising whole plots of both moss species and from both sites). The litterbags were transported in air-tight plastic bags to SLU-Umeå, where they were immediately placed in Tullgren funnels for micro-arthropod extractions. The litterbags were carefully opened and placed upside down to facilitate micro-arthropod extraction. The extraction lasted for four days, and the extracted fauna was collected in 70% ethanol solution. The litter materials were then oven-dried at 60 °C for 24 h and weighed to calculate litter mass loss as a percentage of the initial mass of dry litter. The second half of the litterbags (N = 160) was collected between 28 August 2020 and 7 September 2020 (that is, about 16 months after the litterbags were deployed), and we followed the same procedure to extract micro-arthropods and calculate litter mass loss. Damage to the litterbags in the field and sample mishandling meant that data of litter mass loss were available for 319 out of 320 litterbags, and data from micro-arthropod extractions were available for 302 litterbags.

We assessed the abundance of mites (Acari) and Collembola in each sample using a Wild Heerbrugg M400 dissecting microscope with 6–32 × magnification. Mites were assigned to one of four major taxonomic and trophic groups: Oribatida (fungivorous/omnivorous), Mesostigmata (predaceous), Prostigmata belonging to the family Tydeidae (fungivorous), and all other Prostigmata (predaceous), following Krantz and Walter (2009), and Walter and Proctor (2013). We did not find any Astigmata (fungivorous) in our samples. Because micro-arthropod activity (and, therefore, its effect on decomposition) depends on micro-arthropod biomass (Wood and Lawton 1973; Petersen and Luxton 1982), and given the high abundance of juvenile mites in our samples, we applied a correction factor to the abundance of juvenile mites to account for their smaller biomass relative to adults. Specifically, bigger juvenile mites (that is, those that could be confidently assigned to a morphologic group) were applied a correction factor of 0.238 based on published overall juvenile to adult mass ratios (Block 1966; Wood and Lawton 1973; Mercer and others 2001). Smaller (unidentified) juvenile mites were considered larvae and applied a correction factor of 0.077, after Wood and Lawton (1973) and Mercer and others (2001). All Collembola were considered fungivorous, as we did not find any individuals belonging to the predaceous

family Frieseinae (Holtkamp 2008). In each sample, we added the number of mites and Collembola from all groups to calculate a measure of total micro-arthropod abundance. By collecting the litterbags in sets, and over a few days in each year, we sought to minimise the effect of time of harvest on abundance and composition of micro-arthropods and to ensure that the effects of the experimental treatments prevailed.

## Statistical Analysis

We used R version 4.0.3 (R Core Team 2020) for all analyses and figures. Micro-arthropod community composition was analysed using the R package *vegan* (Oksanen and others 2020). Micro-arthropod community composition was analysed at the functional group level, which included the following groups: Oribatida mites, Mesostigmatida mites, Prostigmatida (Tydeidae) mites, other Prostigmatida mites, unidentified juvenile mites, and Collembola. We tested the effect of litter type, bryosphere removal, moss species and litterbag mesh size on composition using permutational multivariate analysis of variance (PERMANOVA; Anderson 2001), as implemented in the function ‘*adonis2*’, restricting permutations to within plots. PERMANOVA was carried out on a distance matrix computed using the ‘Bray–Curtis’ dissimilarity index (Faith and others 1987). We used non-metrical multidimensional scaling (NMDS; Clarke 1993) on the untransformed data to visualize the data. To test the association between micro-arthropod community composition and litter decomposition rates, we used the function ‘*envfit*’ to fit the variable litter mass loss onto the ordination and assess the statistical significance of the association through permutation tests (restricted to within plots).

Univariate data analysis was based on linear mixed effects models, as implemented in the R package *nlme* (Pinheiro and others 2020). Total micro-arthropod abundance was analysed by fitting an interaction between the factors ‘bryosphere presence’ (bryosphere present or removed) and ‘litter type’ (birch or moss) as fixed effects. The dominant species of the bryosphere (*P. schreberi* or *H. splendens*), mesh size of the litterbags, and time of incubation (4 or 16 months) were also included in the model as additive fixed effects and allowed to interact with the bryosphere presence and litter type interaction. Site was included as an additive fixed effect. To account for the nested structure of the experimental design (that is, litterbags were grouped in plots), plot was included as a random

effect. Litter decomposition (estimated as % mass loss) was analysed by fitting an interaction between the covariate ‘micro-arthropod abundance’ and the factors ‘bryosphere presence’ and ‘litter type’ as fixed effects. As in the previous model, the dominant species of the bryosphere, mesh size of the litterbags, and time of incubation were also included as additive fixed effects and allowed to interact with the previous variables, and site was included as an additive fixed effect. Plot was included as a random effect.

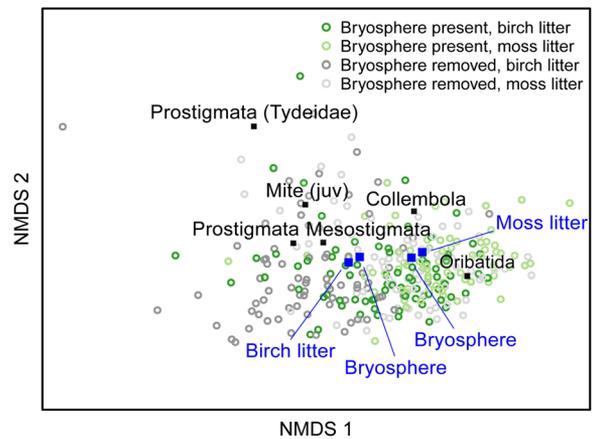
Change in mean daily temperature due to bryosphere removal was analysed by fitting an interaction between season and moss species as fixed effects. Site was added as an additive fixed effect, and plot as a random effect. In all models, we used a constant variance function (‘varIdent’) to account for variance heterogeneity among factor variables (Zuur and others 2009). Marginal  $R^2$  values, that is, the proportion of the total variance explained by the fixed effects, and conditional  $R^2$  values, that is, the variance explained by both fixed and random effects (Nakagawa and others 2017), were calculated using the package *MuMIn* (Barton 2020). Statistical significance of slopes for specific covariate and factor combinations was computed using the function ‘emtrends’ in the package *emmeans* (Lenth 2020).

## RESULTS

### Micro-Arthropod Abundance

Most identified micro-arthropods extracted from the litterbags were Oribatid mites (67%), followed by Mesostigmatid mites (14%), Collembola (13%), Prostigmatid mites belonging to the family Tydeidae (4%) and other Prostigmatida (1%) (Table S1). Micro-arthropod community composition responded most strongly to litter type ( $R^2 = 0.124$ ) and bryosphere removal ( $R^2 = 0.067$ ) (Table S2). Bryosphere presence and moss litter were associated with higher abundances of Oribatid mites and Collembola, while bryosphere removal and birch litter were associated with higher abundances of Mesostigmatid, Prostigmatid, and unidentified juvenile mites (Figure 2). Micro-arthropod composition was not significantly associated with litter mass loss ( $R^2 = 0.032$ ,  $P = 0.674$ ).

Total micro-arthropod abundance was 29% lower in litterbags placed on the soil in bryosphere-removed subplots than in litterbags embedded in the bryosphere layer (that is, in bryosphere-covered subplots) (Figure 3, Table S3). The negative effect of bryosphere removal on total micro-



**Figure 2.** Non-metric multidimensional scaling of micro-arthropod abundance. Micro-arthropods were grouped by the morphological and trophic groups Collembola, Oribatida, Mesostigmata, Prostigmata belonging to the family Tydeidae, other Prostigmata, and unidentified juvenile mites. Open circles represent individual samples. Blue symbols represent treatment centroids. The NMDS had a stress value of 0.14.

arthropod abundance depended on the moss species that dominated the bryosphere (Figure 3, Table S3): effects were stronger for *H. splendens* bryospheres (where abundance decreased by 41%) than for *P. schreberi* bryospheres (where abundance decreased by 14%). Further, the decrease in abundance due to bryosphere removal was greater in litterbags collected after 16 months than after 4 months (Table S3). Litter type had the strongest main effect on micro-arthropod abundance: moss litterbags had 82% higher abundance than birch litterbags, but litter type did not mediate the effect of bryosphere removal on abundance. Finally, the mesh size of the litterbags had no significant effect on micro-arthropod abundance.

### Litter Mass Loss

Removal of the bryosphere significantly reduced litter mass loss, by on average 28% (Figure 4, Table S4). Moss species had a marginally non-significant effect on overall litter mass loss ( $P = 0.065$ ), through being 12% higher in *H. splendens* plots than in *P. schreberi* plots. Litter type also affected litter mass loss, through being 42% higher for birch than moss litter. However, neither moss species nor litter type mediated the effect of bryosphere removal on litter mass loss (Table S4). Mesh size also had no significant effects on litter mass loss. The duration of the incubation mediated both the effect of bryosphere removal and of litter type on litter mass loss, because both effects were

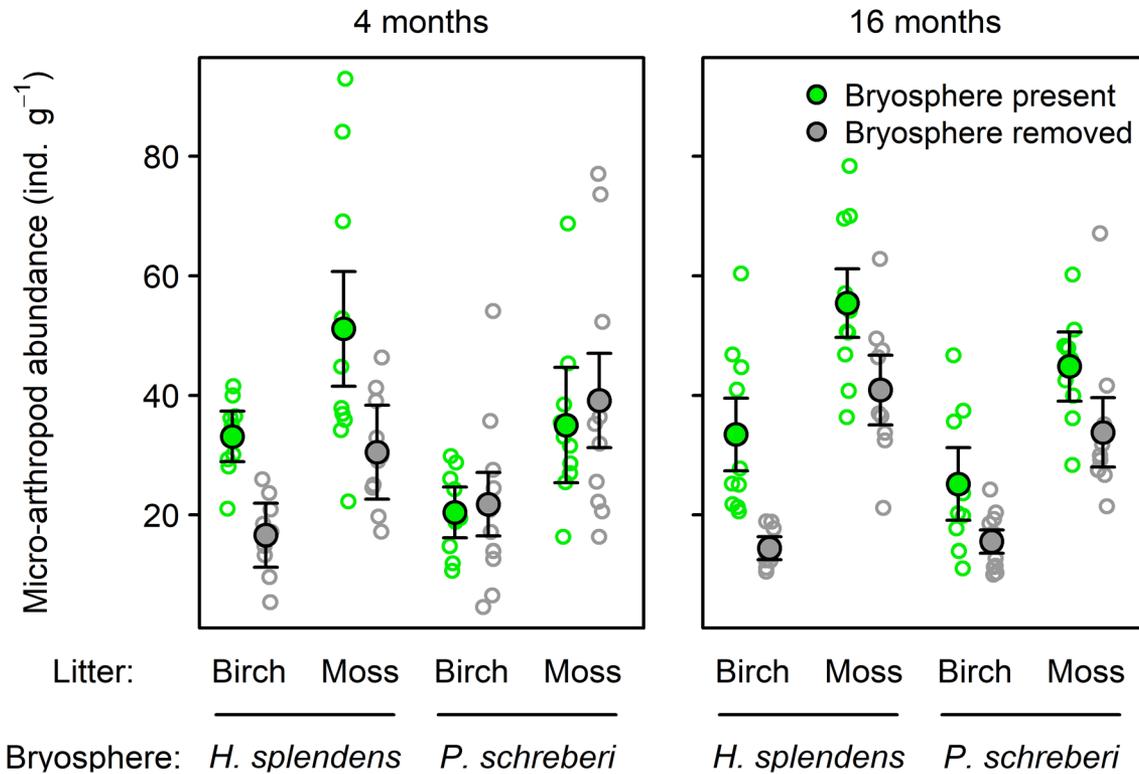


Figure 3. Micro-arthropod abundance on a dry litter weight basis in litterbags collected in (left) September 2019 (that is, after a 4-month field incubation) and (right) August 2020 (that is, after a 16-month field incubation) in bryospheres dominated by *H. splendens* or by *P. schreberi*, in birch or moss litter, and in bryosphere-removed or bryosphere-present subplots. Within the same subplot and litter type, the data of litterbags of both mesh sizes were aggregated because mesh size had no significant effect on total micro-arthropod abundance (Table S3). Open symbols indicate individual observations, solid symbols are means, and bars are 95% confidence intervals.

significantly greater after 16 months than after 4 months (Table S4). Overall, mass loss was 64% higher in litterbags that were collected after 16 months compared to litterbags collected after 4 months.

Micro-arthropod abundance was not correlated with litter mass loss overall (Table S4). However, litter type mediated the correlation between micro-arthropod abundance and litter mass loss, because micro-arthropod abundance was on average positively correlated with mass loss in birch litter ( $P = 0.001$ ) but not in moss litter ( $P = 0.13$ ) (Figure 5). Bryosphere removal and moss species did not alter the relationship between micro-arthropod abundance and litter mass loss (Table S4).

### Ground Temperature

Mean temperature was significantly lower (by on average 0.32 °C) on the ground in bryosphere-removed subplots compared to in the bryosphere of the bryosphere-covered subplots in summer and

autumn, but was similar during the snow period and during the snow-free spring period (Figure 6, Table S5). Daily temperature range did not differ significantly between bryosphere-removed and bryosphere-covered subplots, although we did find some evidence of a lower temperature range in bryosphere-removed subplots during the summer (that is, by  $-0.35$  °C;  $P = 0.066$ ). Differences in mean daily temperature or daily temperature range between bryosphere-covered and bryosphere-removed subplots were similar in *H. splendens* and in *P. schreberi* bryospheres (Table S5).

### DISCUSSION

Litter decomposition at small spatial scales is mainly controlled by micro-climate (which is largely driven by habitat structure), litter quality, and the composition and activity of the decomposer biota (Prescott 2010; Makkonen and others 2012). Here, we used a litterbag experiment to investigate the effect of forest floor structure (that is, different

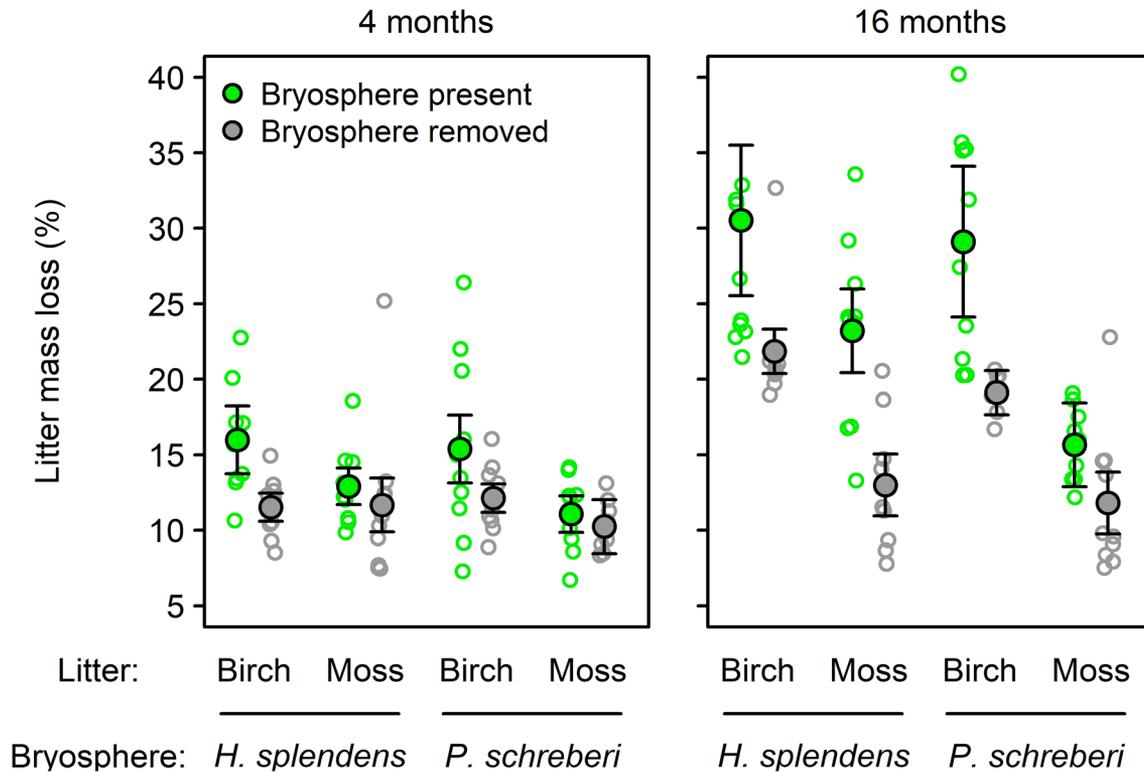
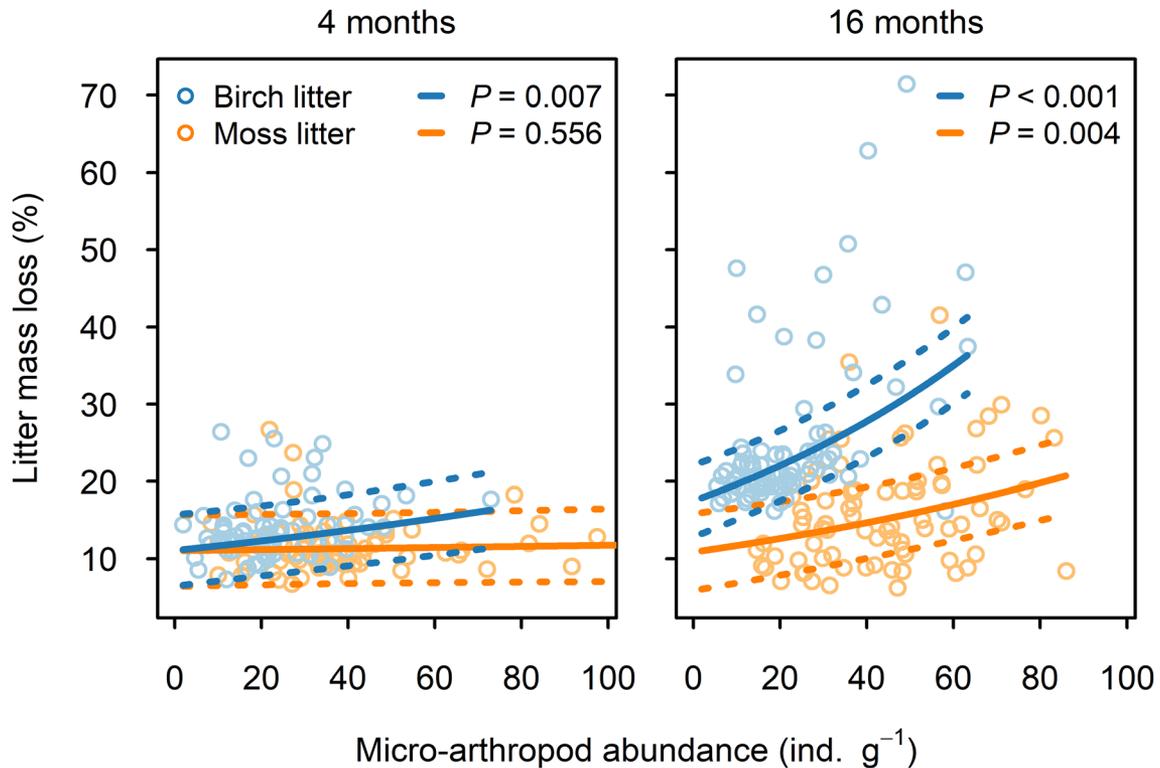


Figure 4. Litter mass loss in litterbags collected in (left) September 2019 (that is, after a 4-month field incubation) and (right) August 2020 (that is, after a 16-month field incubation) in bryospheres dominated by *H. splendens* or by *P. schreberi*, in birch or moss litter, and in bryosphere-removed or bryosphere-present subplots. Within the same subplot and litter type, data of litterbags of both mesh sizes were aggregated because mesh size had no effect on mass loss (Table S4). Open symbols indicate individual observations, solid symbols are means, and bars are 95% confidence intervals. One extreme observation (67% mass loss from a birch litterbag in a *H. splendens* bryosphere-present plot collected in 2020) was not plotted to improve clarity.

bryosphere species, and bare ground), litter type, and micro-arthropod abundance on litter decomposition in boreal forests. We show that bryosphere loss substantially impairs litter decomposition, and that this bryosphere effect is generally not altered by moss species or litter type, or associated with micro-arthropod abundance or composition.

Both micro-arthropod abundance and decomposition rates were higher in litterbags embedded in the bryosphere compared to litterbags on bare ground, in line with our first hypothesis. This is also consistent with previous studies, which have attributed the positive effects of the bryosphere on micro-arthropod abundance and on litter decomposition rates to improved micro-climatic conditions (Jackson and others 2013, De Long and others 2016). We note that our litterbags were placed in the centre of the bryosphere profile, and that even stronger bryosphere effects are likely to occur in the lower parts of the bryosphere (De Long and others 2016). Increased temperature and moisture

in the bryosphere may be important in promoting the decomposer subsystem (Lindo and Winchester 2007; Makkonen and others 2012). Although we did not measure moisture content, our finding of higher mean temperatures in summer and autumn in the bryosphere compared to bare ground indicates that improved micro-climate in the bryosphere could have enhanced soil biota activity and thus decomposition. Given the lower density of the bryosphere compared to soil, a lower thermal inertia in the bryosphere compared to on bare ground could explain the higher temperatures that we found in the bryosphere. Further, our finding that total micro-arthropod abundance was generally uncorrelated with litter decomposition is in line with previous studies frequently showing small or negligible effects of soil fauna on litter decomposition in boreal ecosystems (Wall and others 2008; Makkonen and others 2012; García-Palacios and others 2013; Barreto and Lindo 2018) and indicates that other biotic or abiotic changes in-



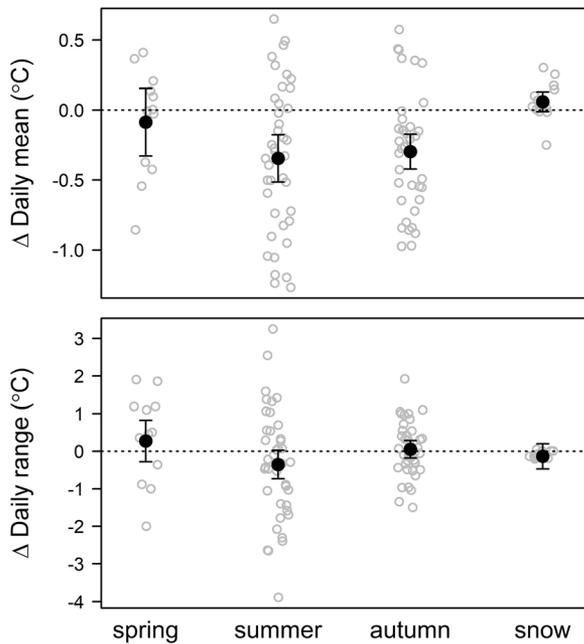
**Figure 5.** Litter mass loss against micro-arthropod abundance in litterbags collected in (left) September 2019 (that is, after a 4-month field incubation) and (right) August 2020 (that is, after a 16-month field incubation), in birch *versus* moss litterbags. The effect of micro-arthropod abundance on litter mass loss was not mediated by bryosphere removal, moss species, or litterbag mesh size, and data were therefore pooled across those variables (Table S4). Open circles are individual observations. Fitted slopes (solid lines) and their 95% confidence intervals (dashed lines) and *P* values are based on a linear mixed effects model including the logarithm of mass loss as a response variable, the interaction between micro-arthropod abundance, litter type and sampling year as fixed effects, and plot as a random effect. One extreme observation (a 2019 moss litter sample with an abundance of 149 ind. g<sup>-1</sup> and a mass loss of 11.5%) was not plotted to improve clarity.

duced by the bryosphere are the main drivers of the observed higher litter decomposition rates in the bryosphere.

Bryosphere removal also altered the composition of the micro-arthropod community. The presence *versus* absence of the bryosphere can potentially impact faunal composition and therefore their trophic interactions, which could in turn impact on microbial activity (Kardol and others 2016) and litter decomposition (Tan and others 2020). However, we found no evidence that changes in micro-arthropod community composition due to bryosphere removal influenced decomposition. This is despite our finding that litterbags in the bryosphere had greater abundance of Oribatid mites and Collembola, which can potentially facilitate microbial decomposition through breakdown of litter (Briones 2014). Further, litterbags on bare ground had greater abundance of predatory mites (e.g., Mesostigmatida), which could potentially

impair the effects of prey species of Oribatid mites and Collembola (Hedlund and Öhrn 2000). Instead, we found that variation in micro-arthropod community composition in response to important environmental variables (that is, litter type, bryosphere presence, and bryophyte species) had no effect on litter decomposition rates.

We found no evidence that bryosphere effects on litter decomposition differed between *P. schreberi* and *H. splendens* bryospheres, which is inconsistent with our second hypothesis. Further, a greater decrease in micro-arthropod abundance occurred following removal of *H. splendens* than *P. schreberi* bryospheres, suggesting that *H. splendens* creates a more favourable habitat for micro-arthropods, also contrary to our second hypothesis. Thermal dynamics within the bryosphere did not differ between *P. schreberi* and *H. splendens* bryospheres, as indicated by their similar mean temperatures and daily temperature range. It is, however, possible



**Figure 6.** Change in (top) mean daily temperature and (bottom) daily range at the ground surface in bryosphere-removed subplots, relative to the bryosphere in adjacent bryosphere-covered subplots (that is, negative values indicate that bryosphere-removed subplots were [top] colder and [bottom] had lower temperature variation than bryosphere-covered subplots) for four observation periods: spring (March–May), summer (June–August), autumn (September–November), and the snow-covered period between 19 November 2019 and 10 May 2020. Grey open symbols are observations averaged by observation period and plot. Black solid symbols are means and 95% confidence intervals.

that although *P. schreberi* bryospheres have higher water holding capacity than *H. splendens* bryospheres (Elumeeva and others 2011), *H. splendens* may prefer wetter environments (Busby and Whitfield 1978), which could have promoted abundance of micro-arthropods in *H. splendens* bryospheres. Additionally, biological nitrogen fixation has been observed to be higher in *H. splendens* than in *P. schreberi* bryospheres in boreal forests (Lagerström and others 2007). Therefore, higher nitrogen availability could have contributed to the observed higher micro-arthropod abundance in litterbags on *H. splendens* bryospheres through supporting greater abundance of microbes and micro-fauna (that is, nematodes and tardigrades), which some micro-arthropods feed upon. This interpretation agrees with our observation (although marginally statistically non-significant) of increased decomposition in *H. splendens* compared to in *P. schreberi*, and with previous findings that

the abundance of micro-fauna is higher in *H. splendens* than in *P. schreberi* bryospheres (Jonsson and others 2015). However, differences in bryosphere effects on micro-arthropod abundance between moss species were not reflected in litter decomposition rates, which supports the observation that micro-arthropod abundance had little overall effect on litter decomposition.

Bryosphere effects on litter decomposition were not mediated by litter type, contrary to our third hypothesis. Specifically, our results indicate that the observed higher micro-arthropod abundance in the bryosphere compared to bare soil did not enhance decomposition of lower-quality *P. schreberi* litter more than decomposition of higher-quality *B. pendula* litter, in contrast both to our hypothesis and to previous studies showing greater soil faunal effects on decomposition of lower-quality litter (Yang and Chen 2009; Milcu and Manning 2011; Joly and others 2020). However, we found that the association between overall micro-arthropod abundance (that is, across all litterbags) and litter decomposition rates was mediated by litter type, because higher abundance of micro-arthropods was associated with enhanced decomposition of higher-quality *B. pendula* litter but not of lower-quality *P. schreberi* litter. Although our correlative analysis cannot rule out that changes in litter mass loss and micro-arthropod abundance were driven by covariation with unaccounted for variables, our results suggest that higher-quality litter may actually lead to stronger faunal effects on decomposition, contrary to our expectation, but in agreement with previous manipulative experiments (Schädler and Brandl 2005; Perez and others 2013; Fujii and others 2018; Tan and others 2020). One possible explanation is that the detritivorous activity of micro-arthropods is impaired in *P. schreberi* litter because of its lower quality, while in *B. pendula* litter, they could have a greater effect in enhancing litter decomposition through feeding on detritus and stimulating microbial communities.

Overall, we found that despite the higher micro-arthropod abundance (and, presumably, higher moisture availability; Wardle and others 2003) in *P. schreberi* litter, possibly as a consequence of its high pore space (Salmane and Brumelis 2008; Glime 2017), the higher-quality *B. pendula* litter decomposed faster, probably because it promoted microbial activity more because of its greater quality. Previous work on boreal ecosystems has identified higher C:N ratios and more recalcitrant compounds in moss litter compared to vascular plant litter as important drivers of the observed lower decomposition rates of moss litter (Wardle and others 2003;

Del Giudice and Lindo 2017; Palozzi and Lindo 2017). Moreover, the effects of litter type and of bryosphere presence on litter decomposition were greater after 16 months than after 4 months, which point to long-lasting effects of litter type and bryosphere presence that intensify over time. In total, our results indicate that while forest floor structure (that is, bryosphere *versus* bare ground) and litter type did not interact, they were both strong drivers of micro-arthropod abundance and litter decomposition rates.

## CONCLUSIONS

Our study shows that bryosphere loss in boreal forests impairs litter decomposition, and that this is unrelated to the concomitant decrease in micro-arthropod abundance, which point to other variables (for example, micro-climate, nutrient supply) driving bryosphere effects. Further, it demonstrates largely consistent effects of bryosphere loss (that is, across different moss species, litter types, and micro-arthropod abundances) on litter decomposition. This supports the view that the bryosphere buffers forest floor processes (Grau-Andrés and others 2021), possibly through regulation of thermal, moisture, or nutrient dynamics. Our results highlight the importance of the forest floor structure (that is, bryosphere *versus* bare ground) on controlling litter decomposition rates and suggest that loss of bryosphere cover, as is projected to occur in high-latitude ecosystems as a result of global change (Elmendorf and others 2012; Alatalo and others 2020), may impair litter decomposition. The reduced carbon and nutrient turnover that is associated with impaired litter decomposition could in the future impact on the capacity of boreal forest soils to supply nutrients for plant growth or to store carbon (Prescott 2010). Therefore, to improve our predictions of changes in litter decomposition and related ecosystem functions in moss-dominated ecosystems, future research should aim to better understand how global change will alter bryosphere cover.

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## DATA AVAILABILITY

Data are available at <https://doi.org/10.5061/dryad.8cz8w9grp>.

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