











## RESEARCH ARTICLE

# Towards rainy high Arctic winters: How experimental icing and summer warming affect tundra plant phenology, productivity and reproduction

Mathilde Le Moullec<sup>1,2</sup>  | Anna-Lena Hendel<sup>3</sup>  | Matteo Petit Bon<sup>4,5</sup>  |  
 Ingibjörg Svala Jónsdóttir<sup>6,7</sup>  | Øystein Varpe<sup>6,8,9</sup>  | René van der Wal<sup>10</sup>  |  
 Larissa Teresa Beumer<sup>6</sup>  | Kate Layton-Matthews<sup>1,11</sup>  | Ketil Isaksen<sup>12</sup>  |  
 Brage Bremset Hansen<sup>1,13,14</sup> 

<sup>1</sup>Department of Biology, Norwegian University of Science and Technology, Trondheim, Norway; <sup>2</sup>Greenland Institute of Natural Resources, Nuuk, Greenland; <sup>3</sup>Department of Wildlife Ecology and Management, University of Freiburg, Freiburg, Germany; <sup>4</sup>Department of Wildland Resources and Ecology Center, Utah State University, Logan, Utah, USA; <sup>5</sup>Department of Applied Ecology, North Carolina State University, Raleigh, North Carolina, USA; <sup>6</sup>Department of Arctic Biology, The University Centre in Svalbard, Longyearbyen, Norway; <sup>7</sup>Institute of Life and Environmental Sciences, University of Iceland, Reykjavík, Iceland; <sup>8</sup>Department of Biological Sciences, University of Bergen, Bergen, Norway; <sup>9</sup>Norwegian Institute for Nature Research, Bergen, Norway; <sup>10</sup>Department of Ecology, Swedish University of Agricultural Sciences, Uppsala, Sweden; <sup>11</sup>Norwegian Institute for Nature Research, Oslo, Norway; <sup>12</sup>Climate and Environment Department, Norwegian Meteorological Institute, Oslo, Norway; <sup>13</sup>Norwegian Institute for Nature Research, Trondheim, Norway and <sup>14</sup>Gjørevoll Centre for Biodiversity Foresight Analysis, Norwegian University of Science and Technology, Trondheim, Norway

**Correspondence**

Mathilde Le Moullec

Email: [mathilde.lemoullec@ntnu.no](mailto:mathilde.lemoullec@ntnu.no); [malm@natur.gl](mailto:malm@natur.gl)**Funding information**

The Research Council of Norway, Grant/Award Number: 223257 and 276080; Svalbard Environmental Protection Fund, Grant/Award Number: 16/113

Handling Editor: Sergey Rosbakh

**Abstract**

1. The Arctic is warming rapidly and much faster in winter than in summer. Warm spells in winter lead to more frequent extreme rain-on-snow events that alter snowpack conditions and can encapsulate plants in 'basal ice' ('icing') for months. Yet, how icing affects plant communities, especially over multiple winters and under warmer summers, remains largely unstudied.
2. We investigated winter icing and summer warming effects on vascular plants' productivity, reproduction and phenology in mesic dwarf shrub heath, an important reindeer habitat in high Arctic Svalbard, where winter temperatures have been rising particularly fast. In a full-factorial field experiment, rain-on-snow and resultant icing were simulated in five consecutive winters, and each followed by experimentally increased summer temperatures. Vascular plant responses at the community level, with particular attention to the dominant dwarf shrub *Salix polaris*, were assessed throughout each subsequent growing season.
3. Icing alone increased community-level primary productivity, but only late in the growing season and reduced inflorescence production. Accordingly, *S. polaris* showed delayed early leaf phenophases, but accelerated subsequent development, resulting in smaller, thinner leaves. This compensatory growth response apparently occurred at the cost of delayed seed maturation. The phenological delay

This is an open access article under the terms of the [Creative Commons Attribution](https://creativecommons.org/licenses/by/4.0/) License, which permits use, distribution and reproduction in any medium, provided the original work is properly cited.

© 2026 The Author(s). *Journal of Ecology* published by John Wiley & Sons Ltd on behalf of British Ecological Society.

was associated with icing-induced delays in spring soil warm-up, possibly favouring resource allocation to primary productivity over reproduction. Experimental summer warming (on average 0.8°C) largely counteracted the effects of icing, enhancing community productivity throughout the growing season, offsetting *S. polaris* leaf size reductions and turning around its delayed phenophases, including seed dispersal. Effect sizes of icing and warming combined could be larger than those under warming alone. Yet, summer warming did not negate the reduction in community inflorescence production caused by icing.

4. **Synthesis.** Extreme rain-on-snow events encapsulating plants in ice can influence high Arctic plant communities in mesic habitats to similar extents as—the better-studied—summer warming. Nevertheless, the absence of magnified icing effects over the years indicates community resistance to icing, particularly under warmer summers, which contrasts with earlier documented ice-induced die-offs in communities dominated by evergreen shrubs. As warm spells during winter become the rule rather than exception, we call for similar experiments in coordinated circumpolar monitoring programmes across the tundra biome.

#### KEYWORDS

basal ice, extreme events, multi-year tundra experiment, open top chamber, rain-on-snow, *Salix polaris*, seasonality, Svalbard

## 1 | INTRODUCTION

Global warming comes with stronger and more frequent extreme climate events, prompting research activity aimed at determining their ecological and evolutionary consequences for terrestrial biota (Harris et al., 2018; IPCC, 2021; Trisos et al., 2020). In the Arctic, warming occurs three times as fast as the global average, with marked climate variability, including for instance extreme droughts, floods, heat and cold waves (AMAP, 2021; Panchen et al., 2022; Van Beest et al., 2022; You et al., 2021). Arctic winter warming is even more pronounced—up to seven times faster than in summer in some regions, resulting in a drastic increase in the frequency of extreme warm spells affecting terrestrial ecosystem functioning (Bintanja & Andry, 2017; Graham et al., 2017; Lemaire et al., 2025; Serreze et al., 2021). Most studies, however, have primarily focused on rising summer temperatures and changes in snow depth (Bjorkman et al., 2020; Frei & Henry, 2022; Henry et al., 2022; Rixen et al., 2022), although winter temperatures are increasingly recognised as important drivers of fine-scale regional vegetation patterns (Bjorkman et al., 2018; Niittynen et al., 2020). Given the key role tundra ecosystems play in global carbon cycling and climate feedback (Maes et al., 2024; Zhang et al., 2020), it is now urgent to understand how the increased frequency of warm spells, combined with warmer summers, influences Arctic plant communities.

Warm spells during winter often come with 'rain-on-snow' events, transforming the snowpack (Bokhorst et al., 2016; Pan et al., 2018; Rasmus et al., 2018). The snowpack may completely

melt, exposing plants to thaw-freeze cycles (Bokhorst et al., 2009, 2011) or—especially in the high Arctic—meltwater freezes to the ground, forming a layer of basal ice (hereafter referred to as 'icing') up to several decimetres thick (Langlois et al., 2017; Peeters et al., 2019; Putkonen & Roe, 2003). Low-growing Arctic plants can remain entirely encapsulated in such ice for months. Both winter warm spells and icing have been identified as the cause of regional Arctic vegetation 'browning' (i.e. a decrease in expected 'Arctic greening' under climate warming; Berner et al., 2020; Bjerke et al., 2017; Myers-Smith et al., 2020; Phoenix et al., 2025; Vickers et al., 2016). However, changes in winter conditions can also interact with summer warming (Frei & Henry, 2022; Kelsey et al., 2021), adding complexity and spatial variability to vegetation productivity trends, highlighting the need to consider the combined effect of winter and summer effects.

Icing has the potential for ecosystem-wide consequences by directly or indirectly affecting several trophic levels, for example by reducing soil arthropod abundance (Coulson et al., 2000), disrupting snow conditions critical for small mammals' survival and reproduction (Kausrud et al., 2008), and causing starvation-induced die-offs in large herbivores, which then affect predators and scavengers (Hansen et al., 2013; Serreze et al., 2021; Sokolov et al., 2016). However, the immediate and long-term effects of icing on high Arctic plant productivity, reproduction and phenology—fuelling the tundra food web—remain largely unknown.

Icing can directly damage vegetation through frost injury, especially after possible de-acclimatation ('de-hardening') to cold during a warm

spell (Bokhorst et al., 2010, 2018) or through stress-induced anoxia from anaerobic respiration (Crawford et al., 1994; Preece & Phoenix, 2014). Indirect effects include altered soil thermal properties (Putkonen & Roe, 2003), delayed nutrient availability and possibly shifted phenology in ways similar to late snowmelt (Assmann et al., 2019; Darrouzet-Nardi et al., 2019; Frei & Henry, 2022; Semenchuk et al., 2016). Icing may also protect plants from winter herbivory. In response, plants may modify life-history trade-offs by altering resource allocation towards growth and survival (Bazzaz et al., 1987; Jónsdóttir, 2011) or leach nutrients into the soil for use by surviving plants and their microbial networks (Darrouzet-Nardi et al., 2019).

Experimental studies on icing effects are limited and all short-term so far, including only a limited number of species of vascular plants and lichens. They reveal impacts on both productivity and reproduction, likely through mechanisms that link the two. Evergreen shrubs and some lichen species seem particularly sensitive to icing (Bjerke, 2011; Bjerke et al., 2017; Finne et al., 2025; Milner et al., 2016; Preece & Phoenix, 2014). Shoot mortality and tissue damage can reduce primary productivity and photosynthetic efficiency, while surviving shoots may exhibit compensatory growth (Bjerke et al., 2018; Finne et al., 2025; Milner et al., 2016; Preece et al., 2012). Such reallocation of resources seems to come at the cost of reduced flowering (e.g. in *Cassiope tetragona*; Milner et al., 2016) or fruit production (e.g. reduced berry yield in *Empetrum nigrum* after repeated icing; Preece & Phoenix, 2014). However, responses vary and some deciduous shrub species such as *Vaccinium myrtillus* showed accelerated leaf emergence with unclear reproductive costs (Preece et al., 2012), while seedlings of *Salix polaris* and the woodrush *Luzula confusa* appeared relatively tolerant, despite some shoot damage (Bjerke et al., 2018).

Warmer summers may counteract some of the icing effects through enhanced growth of surviving shoots and advanced phenology, particularly of the reproductive phenophases of late-flowering species (Collins et al., 2021; Prev  y et al., 2019). Because Arctic plants commonly form flower buds the previous summer (Billings & Mooney, 1968), increased bud initiation under warmer summers could compensate for bud loss from icing. Detecting such effects requires multi-year studies.

We assessed how year-on-year occurrence of icing and summer warming interact to shape high Arctic plant productivity, reproduction and phenology. To achieve this, we set up a full-factorial field experiment in which we simulated, for 5 years in a row, (i) plot-level rain-on-snow and basal ice formation in winter and (ii) warming using open top chambers during summer (treatment and control) following the protocol of the International Tundra Experiment (ITEX; Henry & Molau, 1997). Each growing season, we determined plant community responses and recorded phenology and leaf traits of the dominant species *S. polaris*. We expected icing to (1) delay early phenophases due to a slower soil warm-up; (2) reduce community productivity due to plant damage and shorter growing season; and (3) reduce flower production due to reallocation of resources to compensatory growth. We anticipated summer warming to offset some effects of icing, but with net outcomes to depend on species, plant part and phenophase.

## 2 | MATERIALS AND METHODS

### 2.1 | Study site

The experiment was located in high Arctic Svalbard, in the valley of Adventdalen (78°17'N, 16°02'E), with measurements collected for five consecutive years (January 2016–August 2020). At Svalbard Airport, 15 km from the study site, the annual mean air temperature was −5.9°C for the reference period 1971–2000 (winter: −13.9°C, summer: 4.5°C; Hanssen-Bauer et al., 2019), rising to −2.2°C during the study period (<https://seklima.met.no/>). The average annual precipitation was 196 mm for the reference period (winter: 51 mm, summer: 52 mm; Hanssen-Bauer et al., 2019), increasing to 223 mm during the study period. RCP8.5 projections for Svalbard anticipate increases of 7°C–9°C for annual temperature and 20%–40% for annual precipitation by 2100 (Hanssen-Bauer et al., 2019), with winter warming rates five to seven times faster than in summer (Isaksen, Nordli, et al., 2022). The proportion of annual precipitation falling as snow has already dropped from 50% to 30% over 1975–2015 (Hanssen-Bauer et al., 2019), resulting in more frequent rain-on-snow events (Hansen et al., 2014; Peeters et al., 2019). The basal ice layer formed can reach an average thickness of 15 cm (Peeters et al., 2019). Svalbard currently experiences conditions already seen, or soon expected, across large parts of the Arctic (Bintanja & Andry, 2017; Lemaire et al., 2025; Serreze et al., 2021).

Soil thermal conditions are influenced by the underlying permafrost. At the permafrost monitoring site Janssonhaugen, about 6 km east of the study site, temperature at the permafrost table (2 m depth) warmed at a rate of 1.4°C per decade (2002–2018; Etzelm  ller et al., 2020). At our study site, mean sub-surface soil temperatures (5 cm depth) were 4.7°C in June and 7.5°C in July (Table S1a), the two most important months for plant growth (Van der Wal & Stien, 2014). Volumetric water content (i.e. soil moisture) decreased from about 45% in early June to 25% by mid-August at 10 cm depth (Table S1a), with strong negative correlations between moisture and temperature (Table S1b).

The tundra plant growing season starts immediately upon the onset of soil thaw (Descals et al., 2020), around mid to late May at our site and lasts approximately 3 months (Table 1). Adventdalen lies in the bioclimatic subzone C ‘middle Arctic tundra’ (CAVM Team, 2003). The plant community studied here was a mesic grass and moss-rich dwarf shrub heath, dominated by the prostrate deciduous dwarf shrub *Salix polaris* and the grass *Alopecurus borealis*. Other abundant vascular plants were the horsetail *Equisetum arvense*, the woodrush *Luzula confusa*, the forb *Bistorta vivipara* and the grass *Poa arctica* (Figure S1). The most abundant bryophytes were *Sanionia uncinata*, *Tomentypnum nitens* and *Polytrichastrum* spp. Nomenclature follows <http://panarcticflora.org/> for vascular plants and Frisvoll and Elvebakk (1996) for bryophytes.

### 2.2 | Experimental design

The experiment followed a full-factorial design with two treatments (winter icing and summer warming), each at two levels (present/

TABLE 1 Overview of annual mean biotic and abiotic measurements.

	2016	2017	2018	2019	2020
Soil thaw onset day (range) and DOY [CI]	13th May (8th–22nd May), 134 [129;143]	21th May (3rd May–1st June), 141 [123;152]	10th May (6th–14th May), 130 [126;134]	30th Apr (18th Apr–20th May), 120 [108;140]	18th May (16th–22th May), 139 [137;143]
Air T°C June (range)	5.0 (3.6–6.7)	4.6 (3.2–6.5)	4.0 (2.3–6)	4.8 (3.3–6.8)	4.5 (2.8–6.8)
July (range)	9.0 (7.4–11.1)	6.9 (5.4–8.9)	7.2 (5.6–9.5)	8.4 (6.5–10.6)	9.8 (7.7–12.2)
Time of max. NDVI date and DOY [CI]	16th July, 197 [189;205]	NA	19th July, 199 [191;207]	12th July, 192 [184;200]	15th July, 196 [188;204]
Max. NDVI [CI]	0.74 [0.72;0.77]	NA	0.76 [0.73;0.78]	0.76 [0.73;0.78]	0.79 [0.76;0.81]
Flower count m <sup>2</sup> [CI]	263 [181;345]	229 [151;288]	314 [218;409]	295 [204;386]	491 [341;642]

Note: The daily average air temperatures were recorded at the Svalbard Airport (available on <https://eklima.met.no/>), while the soil thaw onset day, maximum Normalised Difference Vegetation Index (max. NDVI) metrics and the community inflorescence counts (Flower count) were recorded in control plots. Predicted means are presented with their 95% confidence intervals (CI) or range of values. The 'time of max. NDVI' is interpreted as the peak growing season. Abbreviation: DOY, day-of-year (i.e. Julian day, where 1 = 1 January).

absent), resulting in four treatment combinations: icing only (I), warming only (W), combined icing and warming (IW) and control (C). In summer 2015, we selected three blocks (150–780 m apart) of visually homogenous dry-to-mesic tundra. Within each block, we chose 12 plots of 60 × 60 cm with similar vegetation structure and composition of dominant species (listed above), and randomly assigned treatments to obtain three blocks of replicated treatment combinations (Figure 1). The 36 plots received their same respective treatment each year starting in January 2016 (Table S2). Resource constraints meant that not all traits and parameters were measured each year. A major field campaign in 2018, after the third consecutive winter with experimental icing, covered the entire growing season with weekly measurements.

### 2.2.1 | Winter icing

Icing was applied to I and IW plots over 2–3 days in January–February (Figure 1a). Snow was carefully removed from all plots to equalise disturbance, and natural ice occurrence was recorded. Natural ice was common in 2017, with partial ice coverage occurring in all plots (3 cm average thickness). Snow was immediately placed back on non-icing (C and W) plots. For I and IW treatments, a 13 cm high wooden frame (60 × 60 cm) was placed around the plot and gradually filled up with cold water until frozen solid (Figure 1a), following Milner et al. (2016). The resulting ice thickness was within the natural range after extreme rain-on-snow events (Peeters et al., 2019). In the 3 days after treatment application, sub-surface soil temperature increased on average by 2.9°C [2.3;3.5] compared to C plots (Figure S2). Frames did not influence snow accumulation pattern later in winter and were removed at the onset of snowmelt. Ice-covered plots (I and IW) melted almost simultaneously with snow-covered plots (C and W).

Permission to perform the experiment was accorded by Longyearbyen Lokaltstyre (RIS ID: 10484).

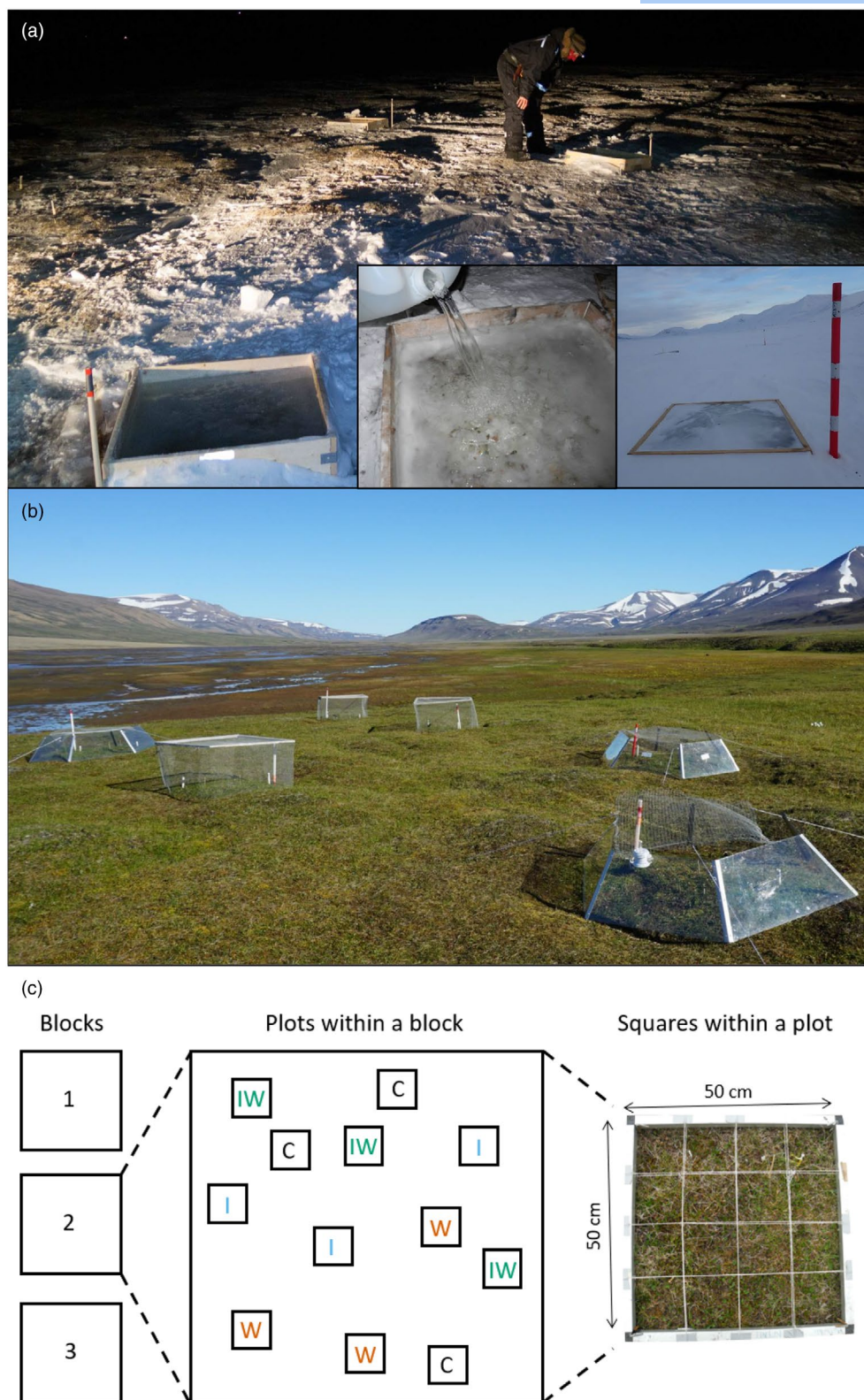
### 2.2.2 | Summer warming

Immediately after snowmelt (Table S2), hexagonal open top chambers (1.4 m base diameter and 33 cm height) made of plexiglass were deployed over the W and IW plots following the ITEX protocol (Henry & Molau, 1997). Open top chambers were deployed and removed at the start and end of the growing season.

### 2.2.3 | Herbivore exclusion

Vertebrate herbivores were excluded from all plots during the snow-free season using metal net cages over I and C plots and nets on top of the open top chambers in W and IW plots (mesh-size 1.9 × 1.9 cm, Figure 1b). Nets were deployed and removed on the same day as the open top chambers.





**FIGURE 1** The experimental set up in the Adventdalen valley, Svalbard. Overview of the field site (a) in the polar night, when applying the icing treatment by filling a wooden frame with water and (b) in summer with the open top chambers (plexiglass hexagons of 1.4 m base diameter and 33 cm height) simulating warming and net cages to protect against herbivory. (c) Experimental nested design with the vegetation frame fitting the plot size. C=control, I=icing, IW=icing  $\times$  warming, W=warming. Picture credits: Ø. Varpe and M. Le Moullec.

## 2.3 | Field measurements

Soil temperatures were recorded with iButton loggers (type DS1921G-F5,  $\pm 1.0^{\circ}\text{C}$  accuracy,  $0.5^{\circ}\text{C}$  resolution) every 120 min in summer and 240 min in winter. Sub-surface loggers were placed in the soil layer 5 cm below the top of the 1–3 cm thick bryophyte layer (Table S1a) to investigate how treatments influenced the upper soil melt onset. In spring 2018, we also investigated treatment effects on soil temperatures at greater depth (20 cm depth, not in *IW* plots), placing loggers in mid-September 2017 using a standard soil core sampler (3 cm diameter cylinder) and delicately replacing the extracted soil columns. We estimated the onset of soil thaw for each plot as the first day when soil sub-surface temperatures were  $\geq 0^{\circ}\text{C}$  for at least 10 consecutive days. We also monitored surface air temperature every 30 min (5 cm above the bryophyte layer) with HOBO loggers (type U23-003/UA-001;  $\pm 0.2^{\circ}\text{C}$  accuracy) in C and W (one pair per block) from 15 June to 1 September in the years 2016–2018 (Table S1a).

Above-ground plant productivity was monitored weekly during the growing season by measuring Normalised Difference Vegetation Index (NDVI, Pettorelli et al., 2005), at the central (30 cm diameter) part of each plot with a Skye SpectroSense2+ handheld device at 125 cm high (with integrated light sensor). We considered the peak growing season to be the average day control (C) plots reached maximum NDVI (Table 1). We also measured community-level vascular plant species abundance (hereafter referred to as 'abundance') shortly after peak growth (end of July) in 2016–2019 (Table 1; Table S2) using point intercept methodology (Bråthen & Hagberg, 2004). We used a 50 × 50 cm sampling frame elevated above the canopy (~20 cm high), with a double layer of strings creating 25 intersections. A pin of 3 mm diameter was then lowered onto the moss layer, at all intersections, recording all 'hits' of live vascular plant species and dead tissue classified as standing dead or litter.

We further investigated vegetative responses of the dominant vascular plant, *S. polaris*, by collecting leaves for trait measurements. Sampling design varied over years. In 2018, we randomly sampled entire shoots at five of the sampling frame's intersections and measured all leaves per shoot. Because of the high variability of leaf size and the time-consuming nature of this method, in 2019 and 2020 we targeted the upper tail of the size distribution, collecting the largest leaf from each of the 16 sub-squares of the 50 × 50 cm sampling frame (Figure 1c; Table S2). Leaves were kept moist until scanned for measuring their area, then oven-dried at  $60^{\circ}\text{C}$  for 4 days and weighed to the nearest 0.01 mg. Specific leaf area (SLA) was calculated as the ratio of leaf area to dry mass.

Flowering frequency was assessed in mid-July each year by counting inflorescences within the 50 × 50 cm area delineated by the recording frame, subdivided into 16 sub-squares, for the species *S. polaris* (males and females [catkins]), *B. vivipara*, *A. borealis*, *L. confusa* and *P. arctica* (Table S2).

Phenology of *S. polaris* was recorded for the most advanced vegetative and reproductive phenological stages ('phenophases') in each of the 16 sub-squares. Vegetative phenophases were: (1) leaves

starting to unfurl, (2) leaves fully expanded, (3) start of senescence and (4) leaves fully senesced. Reproductive phenophases were: (1) distinct inflorescence buds visible, (2) buds recognisable as female or male, (3) receptive stigmas (female) and open anthers releasing pollen (male), (4) stigmas and anthers withered and (5) seed dispersal. Monitoring rounds were most frequent in 2018 and absent in 2020 (Table S2). We investigated different phenology parameters. Timing: the day-of-year when a sub-square reached a given phenophase. Duration: the number of days between two phenophases at the sub-square level. Proportion: the proportion of sub-squares within a plot that reached a given phenophase at specific time points.

## 2.4 | Statistical analysis

We quantified the effects of winter icing, summer warming and their combination, across and within years, on a range of response variables. These included measurements of vegetation productivity at the community level (i.e. NDVI metrics, vascular plants' abundance) and species level (i.e. leaf size traits, *S. polaris* phenophase [timing, duration, proportion]); reproduction at the community level (i.e. inflorescence counts) and species level (*S. polaris* reproductive phenophase [timing, duration, proportion]); and potential environmental drivers (i.e. air/soil temperature, onset of soil thaw).

We used (generalised) linear mixed-effect models as the main analytical tool, to account for unbalanced data with spatial and temporal hierarchical structures defined by our study design and did so by using the 'lmer' and 'glmer' functions from the lme4 package (Bates et al., 2015) in software R-3.6.3 (R Core Team, 2020). We built four separate models for each response variable to: (1) estimate treatment effect sizes and predicted means per year (treatment [categorical] × year [factor] as main and interaction fixed effects), (2) estimate treatment effect sizes and predicted means over years (treatment [categorical] as fixed effect, year as a random intercept), (3) test for icing-warming interaction per year (icing [binary] × warming [binary] × year [factor] as main, two- and three-way interaction fixed effects), (4) test for icing-warming interaction over years (icing [binary] × warming [binary] as main and interaction fixed effects, year as a random intercept). Treatment estimates were identical in models 1 and 3, and models 2 and 4, but the variance of *IW* level (model 1 and 3) depends only on its sample size and residual variance, shared across the other levels (C, *I* and W), thereby avoiding the additional variance introduced by the higher-order interaction (models 3 and 4). For response variables spanning a period (i.e. phenophase proportion, soil temperature), we fitted (5) a model with treatment [categorical] × day-of-year [as factor for predictions per day, as numeric for predicted curves] as main and interaction fixed effects, separately for each year. In these models, we did not test for three-way interactions.

The random intercept structure (i.e. variation in means between replicated units) always included plot ( $n=36$ ) nested within block ( $n=3$ ), with day-of-year nested within year for repeated measurements. It also included sub-square nested within plot (16 sub-squares per plot,  $n=576$ ) for inflorescence count and phenology.

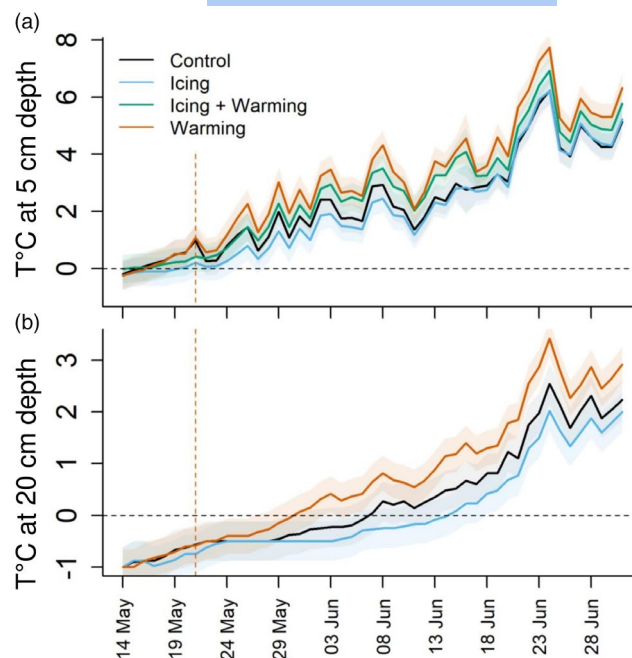
For NDVI data specifically, we first derived different metrics from repeated measurements by fitting a generalised additive model (GAM) to each plot-year combination. This approach was chosen because NDVI changes nonlinearly during the growing season. We used the R function 'gam' from the mgvc package (Wood et al., 2015), fitted with restricted maximum likelihood. Model fit was good (visual evaluation of model residuals and k-indices >1), except for summer 2017 when measurements stopped before the peak growing season (Table S2). Therefore, we predicted daily NDVI values for each plot, each year, except 2017. We then computed five NDVI metrics: (1) maximum NDVI value; (2) day-of-the-year of maximum NDVI; (3) early-season cumulative NDVI (hereafter 'cumulative start') by integrating daily predicted NDVI values from the first day of measurements to the day when C plots, on average, reached maximum NDVI; (4) late-season cumulative NDVI (hereafter 'cumulative end') by integrating daily values from the day when C plot, on average, reached the maximum NDVI to the last measurement day of the season; and (5) cumulative NDVI across the growing season (hereafter 'cumulative total') by integrating daily values from the first to the last day of measurements.

Response distributions were matched to data properties: air/soil temperature data and NDVI metrics were best summarised by a normal distribution; count data of plant abundance, phenophase timing and duration by a Poisson distribution; inflorescences number were square-root transformed and leaf size and weight were log-transformed before modelling. For analysis of phenophase proportion, each phenophase was converted to a binary data format to fit a logistic regression from a binomial distribution. This analytical method best reflects the way the data were collected: by recording whether a sampling unit (i.e. sub-square) was in a certain phenophase (yes/no), regardless of the calendar day. Percentage differences between treatments and controls were calculated from back-transformed predicted means. We interpreted that there was an effect when the 95% confidence interval (CI) of the estimates did not overlap with 0 (i.e. the predicted mean of C); a tendency when it marginally overlapped with 0 but showed a clear direction; and as no effect if it largely overlapped with 0.

### 3 | RESULTS

#### 3.1 | Summer soil temperature

Our icing (I) treatment consistently delayed soil thaw, while warming (W) increased soil temperature, and the combination of icing + warming (IW) produced intermediary effects. Following snowmelt and associated spring meltwater floods, I extended the period soils remained frozen, even though the experimental ice on top of these plots melted almost simultaneously with the snow on top of the control (C) plots (Figure 2a; Table S1a). The onset of sub-surface soil thaw was delayed by 6.0 [2.2;9.9] days in I plots and 4.4 [0.4;8.5] days in IW plots, while W plots remained unaffected (−0.9 [−5.0;3.1] days) relative to C plots (hereafter, results are reported as mean effect sizes [95% CI] in comparison to controls [C] over 2016–2020,



**FIGURE 2** Daily predicted means of soil temperatures during the soil-thawing period from mid-May to end of June 2018, at (a) 5 cm depth (i.e. sub-surface) and (b) 20 cm depth. Shaded areas represent 95% CIs and the orange vertical dashed lines show the day the warming treatments started.

unless specified differently). These icing-induced delays were longer at greater depth. At 20 cm depth (data only from 2018), thaw was delayed by 9.0 [1.3;16.7] days in I plots, staying for about 2 weeks at −0.5°C (i.e. 'the zero curtain period'; Outcalt et al., 1990; Figure 2b), resulting in lagged soil warming throughout June (Figure 2b).

Open top chambers increased soil sub-surface temperatures by 0.6 [0.3;1]°C in W plots and 0.4 [0.1;0.7]°C in IW plots (in 2016, 2018 and 2019) and surface air temperature by 0.8 [0.7;0.9]°C in W plots (June–August, 2016–2018; Table S1a). Soil moisture varied between years but not treatments (Table S1a).

#### 3.2 | Community-level vegetative responses: NDVI metrics and live vascular plant abundance

Icing delayed the onset of peak NDVI but enhanced late-season NDVI, while IW showed greatest seasonal increase. Maximum NDVI occurred on average on 15 July in C plots (Figure 3a; Table 1). It was reached 4 [2;7] days later in I plots, and consistently so over the years (Figure 3i). This delay was reduced in IW plots (2 [−1;4] days) and reversed in W plots (−3 [−5;0] days; Figure 3i, Table S3). Maximum NDVI was higher in I and IW plots in 2016, but this effect vanished in subsequent years as all treatments, including C plots, reached the saturation value of 0.8 (Figure 3b,j; Figure S3; Table S3; Myers-Smith et al., 2020). Cumulative NDVI metrics captured treatment differences in seasonal curve shape (Figure S3). In early-season, only IW increased 'cumulative start' NDVI; NDVI thereafter ('cumulative end') remained high, resulting in an overall increased



'cumulative total' NDVI (Figure 3k–m; Table S3). Icing plots caught up in late season with IW plots (NDVI 'cumulative end') exceeding C plots. Warming alone did not affect NDVI metrics (Figure 3i–m; Table S3). There were no significant interaction effects between *I* and *W* in NDVI metrics (Table S3).

Vascular plant abundance was unchanged under *I* (3% [−11;19]), but increased by 26% [9;46] under *IW*, with the largest effect in 2016 (69% [43;99]; Figure 3n; Table S4). These patterns were driven by the two most abundant vascular plants, the shrub *S. polaris* and the grass *A. borealis* (Table S4), whose abundances determine NDVI values (based on among-plot correlations; Table S5). Under *W*, vascular plant abundance tended to increase by 13% [−2;31] (Table S4). There were no interaction effects between *I* and *W* (Table S4).

### 3.3 | Species-level vegetative responses: leaf size traits and phenology of *S. polaris*

Leaf dry mass and leaf area of *S. polaris* tended to be reduced under *I*, but not under *IW* and *W*. Because dry mass decreased more than leaf area, SLA increased substantially under *I* (Figure 4; Table S6). SLA of *W* plots also tended to increase, but not *IW* plots, resulting in a negative interaction effect (Table S6).

The timing of early leaf phenophases was delayed in *I* plots, but advanced in *IW* and *W* plots. Leaves unfurled 4 [1;8] days later in *I* plots and by a similar number of days earlier in *IW* and *W* plots (−4 [−8;−1] days; Figure 5a,b; Table S7). The duration to unfold leaves (from unfurled to fully expanded) was, however, one-third shorter in *I* plots (−3 [−4;−1] days; Figure S4a; Table S8), offsetting much of the initial delay. In contrast, leaves from *IW* and *W* plots remained fully expanded 20% longer (6 [2;8] days; Table S8) before senescing. Onset of senescence happened almost simultaneously across treatments (Figure 5c,d; Figure S4a), but its rate was 30% faster in *IW* and *W* plots (−2 [−3;−1] days; Figure S4a; Table S8).

### 3.4 | Community-level reproductive response: inflorescence production

The total number of vascular plant inflorescences was reduced by one-third across all treatments, with variable year effects (Figure 3h;

Table S9). Icing had no effect in the first and last treatment years, while *W* effects tended to increase over time (Figure 3p; Table S9). There were more inflorescences under *IW* than expected from *I* and *W* effects being additive (i.e. interaction effect; Table S9). Although *S. polaris* produced the highest number of inflorescences (male and female catkins) and drove the observed inter-annual variability, inflorescences of the forb *B. vivipara* and graminoids were also reduced by *I* (and *IW*) in all years (respectively by −52% [−72;−27] and −67% [−95;−11]; Table S7). Under *W*, *S. polaris* inflorescences were strongly reduced (−70% [−81;−56]), while graminoid inflorescences increased in number (146% [46;225]). Note that in 2020, a summer with rapid soil thaw onset followed by a 'heat wave' in July, the number of inflorescences in C plots reached more than twice the amount of any other year (Figure 3h; Table 1; Table S9). For instance, even with the greatest relative reduction under *W* (Figure 3p), 2020 was still the year with the greatest inflorescence production under this treatment (Table S9).

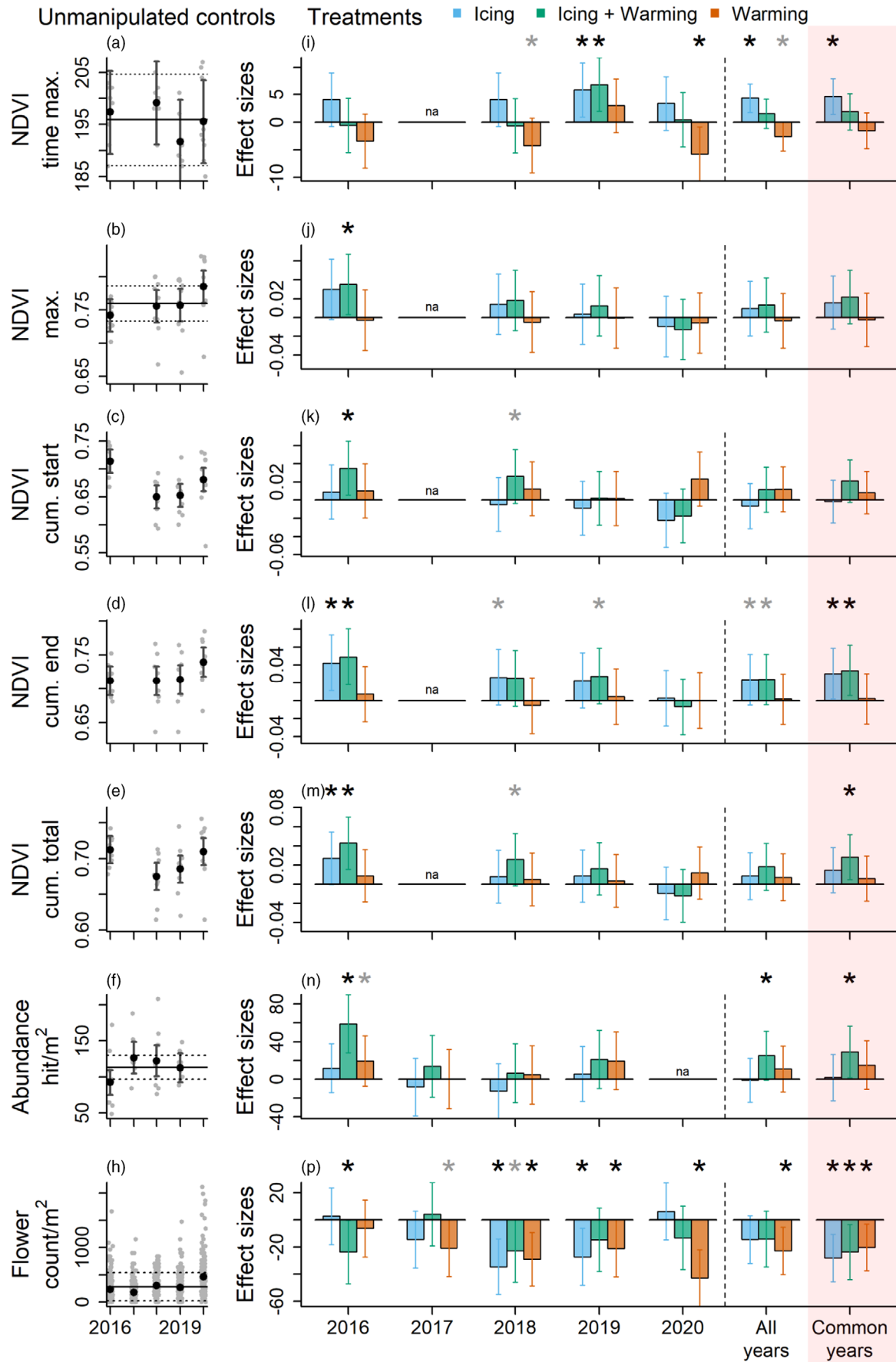
### 3.5 | Species-level reproductive responses: phenology of *S. polaris*

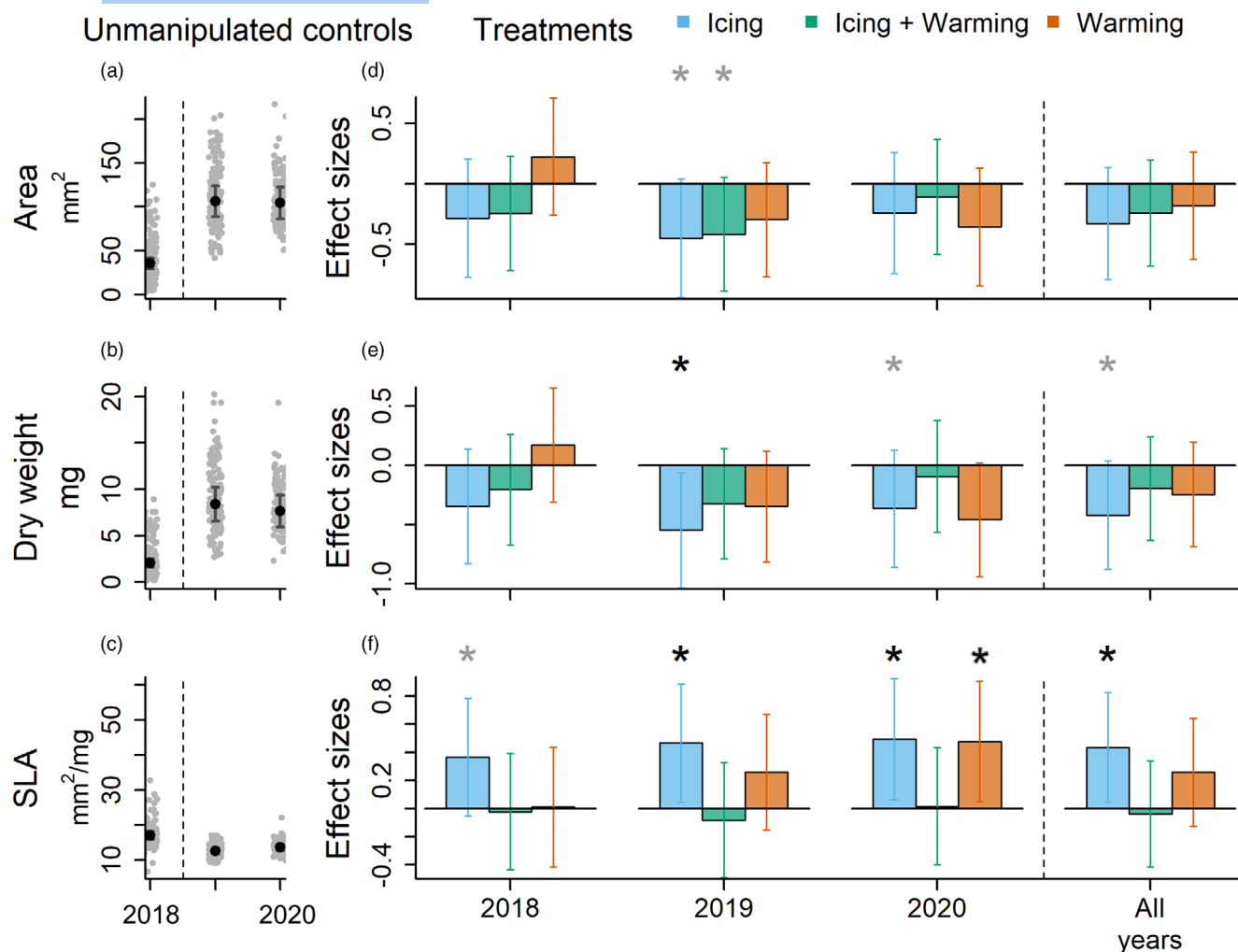
Reproductive phenophases of *S. polaris* were generally delayed under *I* but advanced under *IW* and *W*, with such effects occurring in all years and strongest in 2017 (Figure 5f–l; Figure S4b,c). Interaction effects were largest for seed dispersal: in 2018, when dispersal occurred slightly later in *I* plots (5 [−2;12] days), but markedly earlier in *IW* and, to some extent, *W* plots (−11 [−17;−6] and −5 [−12;2] days, respectively; Figure 5m; Table S7). This advanced seed dispersal in *IW* plots coincided with a 30% shorter duration for seeds to mature (−8 [−14;−1] days; Table S8), while the phase of stigmas being receptive to pollen was extended by 50% (2 [0;5] days; Table S8).

Male flower development followed similar patterns. The timing of anthers being visible and pollen release tended to be later in *I* plots (2 [−3;7] days in 2018; Figure 5j,k, but see Figure S4c for stronger delays in other years such as 2017), earlier in *IW* plots (−5 [−10;−0.4] days) and, to some extent, in *W* plots (−4 [−10;2] days; Figure 5j,k; Table S7). Because anther senescence was rather synchronous across treatments (Figure 5l), the duration of pollen release was twice as long in *IW* plots (5 [1;10] days) but 40% shorter in *I* plots (−2 [−3;0] days; Table S8).

**FIGURE 3** Effects of experimental treatments on Normalised Difference Vegetation Index (NDVI) metrics, vascular plant abundance and the community inflorescence counts (flower count). Left panel (a–h): Model predicted means of unmanipulated control plots (black dots) and their 95% confidence intervals (CIs), represented separately for each year. Horizontal full and dashed lines represent among-year model predictions and their 95% CIs, displayed only for variables comparable between years (i.e. with comparable sampling periods and/or design). Background dots show raw data at the plot ( $n=9$ ) or sub-square ( $n$  range = 109–214) level, jittered for display. Model predictions and their CIs were back-transformed on the response scale. Right panel (i–p): Effect sizes of treatments and their 95% CIs, displayed separately for each year, across all five years and across 'common years' having all parameters measured (i.e. 2016, 2018 and 2019). The reference level at 0 corresponds to the predicted means of controls (presented in the left panel). Asterisks indicate CIs not overlapping zero (black = 95% CI, grey = 90% CI). NDVI cum. end, NDVI cumulative end; NDVI cum. start, NDVI cumulative start; NDVI cum. total, NDVI cumulative total; NDVI max., Maximum NDVI values from −1 to 1; NDVI Time max., Time of maximum NDVI as day-of-year.





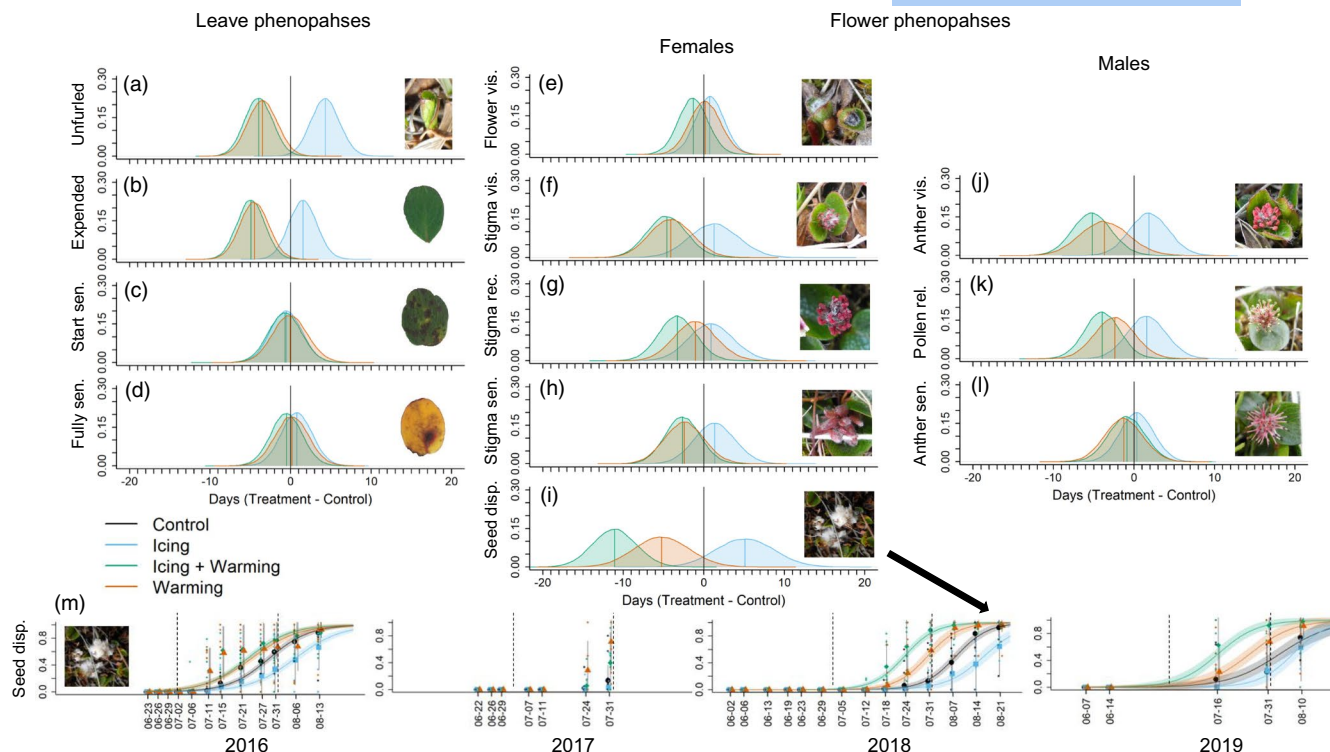


**FIGURE 4** Effects of experimental treatments on leaf size measurements of *Salix polaris*. Left panel (a, c, e): Predicted means of control plots (black dots) and their 95% confidence intervals (CIs), with raw data (grey dots in the background) jittered for display. In 2018, shoots were selected at random and all leaves per shoot measured. In 2019 and 2020, the largest leaf per sub-plot was measured ( $n=16$ ). Model predictions and their CIs were back-transformed on the response scale. Right panel (b, d, f): Effect sizes of treatments and their 95% CIs. The reference level at 0 corresponds to the predicted means of controls (presented in the left panels). Asterisks indicate CIs not overlapping zero (black = 95% CI, grey = 90% CI). SLA = Specific leaf area, the ratio of leaf area to dry weight.

## 4 | DISCUSSION

This field experiment simulating consecutive winters of extreme ice encapsulation of a high Arctic plant community revealed effects with an extent comparable to those of the much better-studied summer warming. Yet, we also found that the observed effects of winter icing could be more than counteracted by subsequent summer warming—supporting the need to account for net effects across seasons to predict climate change impacts. Icing initially delayed spring soil thaw, with a month-long lag in soil warming in the upper active layer, delaying early leaf phenophases of the dominant shrub, *S. polaris*. However, subsequent evidence showed a ‘catching up’ through accelerated leaf development resulting in smaller and thinner leaves and increased community-level primary productivity, reaching

similar or even greater levels later in the growing season. This compensatory growth and productivity seem to happen at the expense of reproduction with species-specific lag-effects. The community-level inflorescence production was reduced and *S. polaris* seed maturation delayed. Summer warming largely counteracted effects of icing, enhancing primary productivity across the entire growing season and offsetting icing-induced reductions in plant leaf size and weight. It also turned around phenophase delays, including the timing of seed dispersal by advancing it even more than under warming alone. Yet, it did not negate the considerable reduction in inflorescence production caused by icing, perhaps because warming was similarly suppressive, suggesting a complex interplay of seasonal effects. Still, given the absence of dramatic die-off or magnified treatment effects over the years, this first multi-year tundra icing experiment reveals



**FIGURE 5** Treatment effects on *Salix polaris* phenophases. (a–l) Differential distributions of the day-of-year each phenophase is reached in 2018 (treatment–controls). Normal distributions were drawn for display from simulations using predicted means and standard-errors obtained from generalised linear mixed-effect models (GLMMs, Poisson distribution; Table S8). The vertical black line at 0 corresponds to the predicted means of controls. Coloured vertical lines represents treatment predicted means. (m) Proportion of sub-squares reaching seed dispersal at each monitoring time-step from 2016 to 2019, with probability curves for each summer (except for 2017 due to insufficient monitoring in late-season). Probability curves for other phenophases and years are presented in Figure S8. The predicted mean at each time-step (symbols) and curves were computed with binomial GLMMs. Grey bars and shaded areas represent 95% CIs; small dots show the raw data at the plot level; the arrow indicates when seed dispersal was reached (i.e. proportion of 1; marking the correspondence between graph i and m for 2018); vertical dashed lines represent 1 July and 1 August. Vegetative phenophases: (a) leaves unfurled, (b) leaves fully expanded, (c) leaves started senescing, (d) leaves fully senesced; reproductive female phenophases: (e) inflorescence visible (female or male catkin), (f) stigmas visible, (g) stigmas receptive, (h) stigmas senesced, (i and m) seed dispersed; reproductive male phenophases: (j) anthers visible, (k) pollen released and (l) anthers senesced.

a rather high level of plant tolerance to icing, particularly when followed by warm summers offsetting some effects of icing.

#### 4.1 | Delayed growth but increased productivity: evidence of compensatory growth?

The apparent absence of major plant die-off due to icing, and even evidence of increased community-level primary productivity under combined icing + summer warming, contrasts with the documented 'Arctic browning' following extreme winter warm spells, a phenomenon in part related to damage in evergreen shrubs and reduced primary production (Bjerke et al., 2017; Phoenix & Bjerke, 2016). Evergreen shrubs, virtually lacking in our study plots, maintain long-lived leaves and dormant buds well above-ground, making them more vulnerable to icing than deciduous shrubs (Givnish, 2002). Dead shoots of evergreen shrubs will remain visible for several years, lowering spectral indices such as NDVI, even if growth may be enhanced in remaining surviving shoots (Bjerke et al., 2017;

Milner et al., 2016; Treharne et al., 2020). In contrast, in our experiment dominated by the deciduous shrub *S. polaris*, with large parts of shoots nested into the ground (Le Moullec et al., 2019), had no evident above-ground signs of damage, likely allowing compensatory productivity and growth later in the season, reflected by the enhanced NDVI. Thus, despite severe treatments, the plant community seemed able to withstand repeated icing, possibly even contributing to a further 'Arctic greening'. The latter may also be a landscape-scale response to increased icing, allowing for a further expansion of deciduous shrubs and graminoids at the expense of icing-sensitive evergreen shrubs responsible for much of the observed 'Arctic browning'.

Icing delayed thawing and warming of the upper soil active layer, despite an almost simultaneous melting of basal ice and snow across treatments, shortening the time window for processes enabling plant nutrient acquisition and growth (Darrouzet-Nardi et al., 2019; Frei & Henry, 2022). This can be explained by infiltration and re-freezing of meltwater in the underlying soil, increasing ice content in near-surface soils. Ice-rich or wet soils require more energy to warm,

especialmente sobre permafrost (Isaksen, Lutz, et al., 2022). El calentamiento estival acortó este lapso.

En la 'carrera' por maximizar la producción primaria dentro de la corta temporada de crecimiento ártica, *S. polaris* parece compensar los retrasos en la fenología de las hojas en las parcelas de heladas acelerando el desarrollo posterior de las hojas, produciendo hojas más pequeñas y más delgadas con un mayor SLA (i.e. reflejando una mayor tasa de crecimiento; Pérez-Harguindeguy et al., 2013). Un experimento de heladas en laboratorio sobre *S. polaris* respalda esto, encontrando hojas reducidas pero un mayor número de hojas y sin cambios en la capacidad fotosintética (Bjerke et al., 2018). Este arbusto caducifolio tiene la capacidad de activar yemas latentes para compensar el daño, como también se ve en respuesta a la herbivoría (Skarpe & Van der Wal, 2002). Sin embargo, la asignación de recursos al crecimiento primario puede venir a expensas del crecimiento secundario (e.g. crecimiento de anillos), lo que se ha encontrado reducido en *S. polaris* (Le Moullec et al., 2020). El helado puede alterar la fuerte correlación entre el crecimiento primario y secundario en esta especie (Le Moullec et al., 2019).

Como se esperaba, el calentamiento estival parcialmente contrarrestó los efectos. Además, nuestros resultados proporcionaron evidencia de tales efectos siendo más fuertes que simplemente aditivos. Por ejemplo, el calentamiento convirtió un retraso inducido por heladas en la fenología de *S. polaris* en un avance. En el bajo Ártico, donde los veranos son más cálidos y podrían ser comparables a nuestro tratamiento de calentamiento, un experimento de heladas coincidió con nuestros hallazgos sobre el avance temprano de la emergencia de las hojas de los arbustos explicada por el crecimiento compensatorio de los ramets dañados por heladas (Preece et al., 2012; Preece & Phoenix, 2014). Según esto, en nuestras parcelas, la productividad primaria aumentó temprano en la temporada de crecimiento bajo heladas + calentamiento y permaneció alta, con las hojas de *S. polaris* en la fase completamente expandida por el 20% más tiempo, su tamaño y masa se mantuvieron. Además, el dominante gramíneo *A. borealis* también aumentó en abundancia en algunos años, y juntos las abundancias de *S. polaris* y *A. borealis* moldearon la productividad a nivel de comunidad (i.e. NDVI; [Tabla S5](#)). Los mecanismos pueden incluir: (1) plantas dañadas (e.g. raíces o yemas) por heladas podrían lixiviar nutrientes en el suelo local que, con veranos más cálidos, podría estar disponible más temprano para las plantas sobrevivientes, particularmente en la tundra más seca (Bardgett et al., 2007); (2) activación más temprana de hormonas de crecimiento y redistribución interna de nutrientes y carbohidratos al crecimiento y reparación de potencial estrés inducido por heladas o daño (Chapin III, 1991); (3) mayor mineralización de nutrientes del suelo bajo heladas, llevando a una mayor disponibilidad para las plantas después del derretimiento (Rixen et al., 2022), aunque el hielo, a diferencia de la nieve, no aísla el suelo y puede restringir la respiración microbiana. Tales posibles respuestas necesitan más investigación.

A pesar de la mayor productividad primaria en la temporada tardía bajo heladas y heladas + calentamiento, la senescencia de las hojas ocurrió sincronizadamente entre los tratamientos, más probablemente impulsado por señales como el fotoperíodo. Sin embargo, el tiempo de la senescencia de las hojas en las parcelas de control varió entre años en más de 10 días ([Figura S4a](#)), sugiriendo interacciones con otras variables compartidas entre los tratamientos, como el tiempo de la nieve o el derretimiento (Bjorkman et al., 2015; Cooper et al., 2011; Kelsey et al., 2021; Wipf & Rixen, 2010). Bajo calentamiento, sin embargo, la senescencia fue ligeramente retrasada ([Figura S4a](#)), lo que concuerda con previas síntesis (Bjorkman et al., 2020; Collins et al., 2021).

## 4.2 | Reduced reproduction: Result of trade-offs?

El crecimiento compensatorio puede venir a un costo de reproducción, como observado después de ciclos de descongelación-congelación (Bokhorst et al., 2011), manipulación de la nieve (Cooper et al., 2011), adición de nutrientes (Petraglia et al., 2013), veranos más fríos (Johnstone & Henry, 1997) y heladas en arbustos perennifolios (Milner et al., 2016). En nuestro estudio, el helado redujo el número de inflorescencias (en algunos años a la mitad) posiblemente debido a un mayor sumidero vegetativo, daño directo de invierno a yemas reproductivas y un posterior cambio en la asignación de recursos de la reproducción al crecimiento, o un tiempo más corto entre el brote verde y la floración (i.e. el brote verde más retrasado que la floración), lo que puede reducir la inversión en floración (Collins et al., 2021; Gougherty & Gougherty, 2018). En el primer año experimental, *S. polaris* mostró ninguna reducción de inflorescencias, sugiriendo que las yemas de inflorescencia formadas el verano anterior sobrevivieron al invierno de heladas. En contraste, otras especies, como *B. vivipara* y *A. borealis*, exhibieron una fuerte reducción sin un lapso. En el mismo valle, eventos extremos de lluvia sobre nieve en el invierno de 2012 fueron sugeridos para causar un récord bajo número de inflorescencias en la comunidad (Semenchuk et al., 2013).

Mientras que el calentamiento estival suprimió la abundancia a nivel de comunidad de inflorescencias sin añadir al efecto negativo de heladas, claramente redujo el retraso inducido por heladas en las fases reproductivas de *S. polaris*, y a un nivel más allá del obtenido por calentamiento solo. Avanzar la floración y fructificación bajo calentamiento son consistentes con resultados de experimentos de calentamiento circumpolar (Arft et al., 1999; Bjorkman et al., 2020; Collins et al., 2021; Prevéy et al., 2019). Como Collins et al. (2021), encontramos mayores cambios en las fases reproductivas que vegetativas, considerablemente acortando el período entre la emergencia de las hojas y la dispersión de semillas. Interesantemente, la duración de las fases reproductivas no fue fija, lo que contradice los hallazgos generales de un experimento de manipulación de la nieve (Semenchuk et al., 2016). Por ejemplo, los períodos de inflorescencia y desarrollo de semillas fueron ambos acortados bajo calentamiento (independientemente de heladas), mientras que los de liberación de polen y receptividad de los estigmas al polen se duplicaron en duración bajo heladas + calentamiento. Tal flexibilidad fenológica puede ser adaptativa en entornos variables. Los estudios de fenología rara vez distinguen estos cortos, pero críticos, estadios para la reproducción del tiempo total de floración (Song & Saavedra, 2018).

## 4.3 | Evolutionary implications

Consecuencias evolutivas pueden esperarse de cambios en los trade-offs que influyen el esfuerzo reproductivo sexual y el éxito. Sin embargo, la mayoría de las especies de plantas árticas son perennifolios clonales que pueden retrasar la reproducción sexual hasta años favorables (Bazzaz et al., 1987; Jónsdóttir, 2011). Además, las plantas árticas muestran una gran variabilidad funcional intraspecífica (Thomas et al., 2020), lo que contribuye adicionalmente a la resistencia a nivel de comunidad a cambios ambientales (Jónsdóttir et al., 2023). En consecuencia, podemos



expect slow evolutionary responses to icing and summer warming. The lack of accentuated effects after five consecutive years of icing indicates that species of the mesic community studied here have evolved strategies to cope with such extreme events.

In the oceanic climate of Svalbard, winter rain-on-snow events have occurred for decades, although their frequency and extent have increased dramatically in recent years (Hansen et al., 2014; Peeters et al., 2019). Selection acting upon life-history traits might change when acute stress episodes from occasional icy winters become 'chronic', as projected for the Arctic (Bintanja & Andry, 2017). In regions that have more rarely encountered widespread icing events than Svalbard (e.g. the Canadian high Arctic), effects may be more severe, particularly where there would be no possible introgression of adapted genes from nearby habitat (Bjorkman et al., 2017). Such limits to adaptation could lead to tipping points and irreversible ecosystem change. Comparative studies among Arctic regions with different frequencies of rain-on-snow events could shed light on plants' adaptive capacities to icing.

#### 4.4 | Ecological implications

The effect sizes of winter icing versus summer warming were of similar magnitude and within the range of natural inter-annual variation of controls. Although the use of open top chambers only increased summer surface air and sub-surface soil temperatures by 0.8°C and 0.9°C, respectively, this increase was similar to the decadal trend found in this region (Etzelmüller et al., 2020; Isaksen, Nordli, et al., 2022) and to other high Arctic ITEX sites, where low irradiance restricts warming (Bokhorst et al., 2013; Hollister et al., 2023; Jónsdóttir et al., 2023). As for other decadal-scale warming experiments, there was no evidence of exacerbated treatment effects over the years, and increases in primary productivity were therefore limited (Barrett & Hollister, 2016; Jónsdóttir et al., 2023). Similar or larger effect sizes from treatments concerning winter rather than summer conditions have been found in other tundra studies, notably those manipulating snow accumulation (Cooper et al., 2011; Frei & Henry, 2022; Niittynen et al., 2020), underscoring that conditions during the 'dormant season' should not be overlooked.

In contrast to evergreen shrub communities, the mesic community studied here represents an important resource for Arctic herbivores (Chapin III et al., 1996). Widespread icing may thus affect plant-herbivore interactions in opposite ways: directly by limiting food accessibility during winter (Hansen et al., 2013; Langlois et al., 2017) but also indirectly by shifting phenology (e.g. phenological mismatches; Doiron et al., 2015; Post et al., 2009) and subsequently increasing summer forage availability.

Although our pioneer multi-year experiment revealed measurable effects, their magnitude remained moderate, highlighting the relatively high tolerance of the plants to icing and summer warming, which contributes to community resistance to these potential drivers of change (Finne et al., 2025; Hudson & Henry, 2010; Jónsdóttir

et al., 2023). However, our 60×60cm basal ice plots may not fully capture processes during large-scale icing events. For instance, anoxia could be more severe without edge effects, potentially causing asphyxiation of plants, arthropods and microorganisms (Coulson et al., 2000; Crawford et al., 1994). Likewise, effects on soil thermal properties might be stronger in natural conditions, potentially amplifying the responses observed here.

## 5 | CONCLUSIONS

High Arctic environments are naturally variable, and plants have developed elaborate strategies to persist under these extreme conditions, resulting in communities that are highly resistant to climate change, despite measurable impacts (Jónsdóttir et al., 2023). This study has shown that the impact of winter rain-on-snow and associated tundra icing can be of similar extent as summer warming, but without causing dramatic die-off. Following up on over 30 years of ITEX experimentation across the tundra biome, providing critical knowledge on tundra responses to summer warming (Henry et al., 2022; Hollister et al., 2023), we call for the implementation of similar coordinated, distributed experiments investigating effects of winter warming and rain-on-snow events in particular.

### AUTHOR CONTRIBUTIONS

Brage Bremset Hansen, Øystein Varpe, Ingibjörg Svala Jónsdóttir, René Van der Wal, Mathilde Le Moulec and Matteo Petit Bon conceived the ideas and designed methodology; Mathilde Le Moulec, with the help of all authors, collected the data; Mathilde Le Moulec, with the help of Anna-Lena Hendel analysed the data; Mathilde Le Moulec led the writing of the manuscript. All authors contributed critically to the drafts and gave final approval for publication.

### ACKNOWLEDGEMENTS

This project was funded by the Research Council Norway (FRIPRO 276080 and SFF-III 223257) and Svalbard Environmental Protection Fund (project 16/113). We thank UNIS logistics for logistical support. We are also deeply grateful for all the help in the field from Solvei B. Hovdal, Martha Grotheim, Hanne K. Haraldsen, Julia Greulich, Katrin Björnsdóttir, Kristine Valøen, Robin B. Zweigel, Ådne Nafstad, Marit Arneberg, Kristiane Midhaug, Marte Sørensen, Marianne Angård, Hamish Burnett, Lia Lechler, Charlotte Karlsen, Kjerstin Hilmarssen, Lukas Tietgen and Svea Zimmermann.

### CONFLICT OF INTEREST STATEMENT

The authors have no conflicts of interest to declare.

### PEER REVIEW

The peer review history for this article is available at <https://www.webofscience.com/api/gateway/wos/peer-review/10.1111/1365-2745.70234>.

## DATA AVAILABILITY STATEMENT

Data deposited in the open access Dryad Digital Repository at <https://doi.org/10.5061/dryad.9ghx3ffxx> (Le Moullec et al., 2025).

## ORCID

Mathilde Le Moullec  <https://orcid.org/0000-0002-3290-7091>  
 Anna-Lena Hendel  <https://orcid.org/0000-0003-0282-9931>  
 Matteo Petit Bon  <https://orcid.org/0000-0001-9829-8324>  
 Ingibjörg Svala Jónsdóttir  <https://orcid.org/0000-0003-3804-7077>  
 Øystein Varpe  <https://orcid.org/0000-0002-5895-6983>  
 René van der Wal  <https://orcid.org/0000-0002-9175-0266>  
 Larissa Teresa Beumer  <https://orcid.org/0000-0002-5255-1889>  
 Kate Layton-Matthews  <https://orcid.org/0000-0001-5275-1218>  
 Ketil Isaksen  <https://orcid.org/0000-0003-2356-5330>  
 Brage Bremset Hansen  <https://orcid.org/0000-0001-8763-4361>

## REFERENCES

- AMAP. (2021). *Arctic climate change update 2021: Key trends and impacts. Summary for policy-makers*. Arctic monitoring and assessment Programme (AMAP).
- Arft, A. M., Walker, M. D., Gurevitch, J., Alatalo, J. M., Bret-Harte, M. S., Dale, M., Diemer, M., Gugerli, F., Henry, G. H. R., Jones, M. H., Hollister, R. D., Jónsdóttir, I. S., Laine, K., Lévesque, E., Marion, G. M., Molau, U., Mølgaard, P., Nordenhäll, U., Raszhivin, V., ... Wookey, P. A. (1999). Responses of tundra plants to experimental warming: Meta-analysis of the international tundra experiment. *Ecological Monographs*, 69, 491–511. [https://doi.org/10.1890/0012-9615\(1999\)069\[0491:ROTPTE\]2.0.CO;2](https://doi.org/10.1890/0012-9615(1999)069[0491:ROTPTE]2.0.CO;2)
- Assmann, J. J., Myers-Smith, I. H., Phillimore, A. B., Bjorkman, A. D., Ennos, R. E., Prevé, J. S., Henry, G. H. R., Schmidt, N. M., & Hollister, R. D. (2019). Local snow melt and temperature—But not regional sea ice—Explain variation in spring phenology in coastal Arctic tundra. *Global Change Biology*, 25, 2258–2274. <https://doi.org/10.1111/gcb.14639>
- Bardgett, R. D., Van der Wal, R., Jónsdóttir, I. S., Quirk, H., & Dutton, S. (2007). Temporal variability in plant and soil nitrogen pools in a high-Arctic ecosystem. *Soil Biology & Biochemistry*, 39, 2129–2137. <https://doi.org/10.1016/j.soilbio.2007.03.016>
- Barrett, R. T., & Hollister, R. D. (2016). Arctic plants are capable of sustained responses to long-term warming. *Polar Research*, 35, 25405. <https://doi.org/10.3402/polar.v35.25405>
- Bates, D., Maechler, M., Bolker, B., & Walker, S. (2015). Fitting linear mixed-effects models using lme4. *Journal of Statistical Software*, 67(1), 1–48. <https://doi.org/10.18637/jss.v067.i01>
- Bazzaz, F. A., Chiariello, N. R., Coley, P. D., & Pitelka, L. F. (1987). Allocating resources to reproduction and defense. *BioScience*, 37(1), 58–67. <https://doi.org/10.2307/1310178>
- Berner, L. T., Massey, R., Jantz, P., Forbes, B. C., Macias-Fauria, M., Myers-Smith, I., Kumpula, T., Gauthier, G., Andreu-Hayles, L., Gaglioti, B. V., Burns, P., Zetterberg, P., D'Arrigo, R., & Goetz, S. J. (2020). Summer warming explains widespread but not uniform greening in the Arctic tundra biome. *Nature Communications*, 11(1), 4621. <https://doi.org/10.1038/s41467-020-18479-5>
- Billings, W. D., & Mooney, H. A. (1968). The ecology of Arctic and alpine plants. *Biological Reviews*, 43, 481–529. <https://doi.org/10.1111/j.1469-185X.1968.tb00968.x>
- Bintanja, R., & Andry, O. (2017). Towards a rain-dominated Arctic. *Nature Climate Change*, 7(4), 263–267. <https://doi.org/10.1038/nclimate3240>
- Bjerke, J. W. (2011). Winter climate change: Ice encapsulation at mild sub-freezing temperatures kills freeze-tolerant lichens. *Experimental and Experimental Botany*, 72(3), 404–408. <https://doi.org/10.1016/j.envexpbot.2010.05.014>
- Bjerke, J. W., Elverland, E., Jaakola, L., Lund, L., Zagajewski, B., Bochenek, Z., Kłos, A., & Tømmervik, H. (2018). High tolerance of a high-arctic willow and graminoid to simulated ice encasement. *Boreal Environment Research*, 23, 329–338.
- Bjerke, J. W., Treharne, R., Vikhamar-Schuler, D., Karlsen, S. R., Ravolainen, V., Bokhorst, S., Phoenix, G. K., Bochenek, Z., & Tømmervik, H. (2017). Understanding the drivers of extensive plant damage in boreal and arctic ecosystems: Insights from field surveys in the aftermath of damage. *Science of the Total Environment*, 599–600, 1965–1976. <https://doi.org/10.1016/j.scitotenv.2017.05.050>
- Bjorkman, A. D., Elmendorf, S. C., Beamish, A. L., Vellend, M., & Henry, G. H. R. (2015). Contrasting effects of warming and increased snowfall on Arctic tundra plant phenology over the past two decades. *Global Change Biology*, 21(12), 4651–4661. <https://doi.org/10.1111/gcb.13051>
- Bjorkman, A. D., García Criado, M., Myers-Smith, I. H., Ravolainen, V., Jónsdóttir, I. S., Westergaard, K. B., Lawler, J. P., Aronsson, M., Bennett, B., Gardfjell, H., Heiðmarsson, S., Stewart, L., & Normand, S. (2020). Status and trends in Arctic vegetation: Evidence from experimental warming and long-term monitoring. *Ambio*, 49(3), 678–692. <https://doi.org/10.1007/s13280-019-01161-6>
- Bjorkman, A. D., Myers-Smith, I. H., Elmendorf, S. C., Normand, S., Rüger, N., Beck, P. S. A., Blach-Overgaard, A., Blok, D., Cornelissen, J. H. C., Forbes, B. C., Georges, D., Goetz, S. J., Guay, K. C., Henry, G. H. R., HilleRisLambers, J., Hollister, R. D., Karger, D. N., Kattge, J., Manning, P., ... Weiher, E. (2018). Plant functional trait change across a warming tundra biome. *Nature*, 562(7725), 57–62. <https://doi.org/10.1038/s41586-018-0563-7>
- Bjorkman, A. D., Vellend, M., Frei, E. R., & Henry, G. H. R. (2017). Climate adaptation is not enough: Warming does not facilitate success of southern tundra plant populations in the high Arctic. *Global Change Biology*, 23, 1540–1551. <https://doi.org/10.1111/gcb.13417>
- Bokhorst, S., Bjerke, J. W., Davey, M. P., Taulavuori, K., Taulavuori, E., Laine, K., Callaghan, T. V., & Phoenix, G. K. (2010). Impacts of extreme winter warming events on plant physiology in a sub-Arctic heath community. *Physiologia Plantarum*, 140(2), 128–140. <https://doi.org/10.1111/j.1399-3054.2010.01386.x>
- Bokhorst, S., Bjerke, J. W., Street, L. E., Callaghan, T. V., & Phoenix, G. K. (2011). Impacts of multiple extreme winter warming events on sub-Arctic heathland: Phenology, reproduction, growth, and CO<sub>2</sub> flux responses. *Global Change Biology*, 17(9), 2817–2830. <https://doi.org/10.1111/j.1365-2486.2011.02424.x>
- Bokhorst, S., Jaakola, L., Karppinen, K., Edvinsen, G. K., Mæhre, H. K., & Bjerke, J. W. (2018). Contrasting survival and physiological responses of sub-Arctic plant types to extreme winter warming and nitrogen. *Planta*, 247, 635–648. <https://doi.org/10.1007/s00425-017-2813-6>
- Bokhorst, S., Huiskes, A., Aerts, R., Convey, P., Cooper, E. J., Dalen, L., Erschbamer, B., Gudmundsson, J., Hofgaard, A., Hollister, R. D., Johnstone, J., Jónsdóttir, I. S., Lebouvier, M., de Van Vijver, B., Wahren, C.-H., & Dorrepaal, E. (2013). Variable temperature effects of open top chambers at polar and alpine sites explained by irradiance and snow depth. *Global Change Biology*, 19(1), 64–74. <https://doi.org/10.1111/gcb.12028>
- Bokhorst, S., Pedersen, S. H., Brucker, L., Anisimov, O., Bjerke, J. W., Brown, R. D., Ehrich, D., Essery, R. L. H., Heilig, A., Ingvander, S., Johansson, C., Johansson, M., Jónsdóttir, I. S., Inga, N., Luojus, K., Macelloni, G., Mariash, H., McLennan, D., Rosqvist, G. N., ... Callaghan, T. V. (2016). Changing Arctic snow cover: A review of recent developments and assessment of future needs for observations, modelling, and impacts. *Ambio*, 45(5), 516–537. <https://doi.org/10.1007/s13280-016-0770-0>
- Bokhorst, S. F., Bjerke, J. W., Tømmervik, H., & Phoenix, G. K. (2009). Winter warming events damage sub-Arctic vegetation: Consistent

- evidence from an experimental manipulation and a natural event. *Journal of Ecology*, 97(6), 1408–1415. <https://doi.org/10.1111/j.1365-2745.2009.01554.x>
- Bråthen, K. A., & Hagberg, O. (2004). More efficient estimation of plant biomass. *Journal of Vegetation Science*, 15(5), 653–660. <https://doi.org/10.1111/j.1654-1103.2004.tb02307.x>
- CAVM Team. (2003). *Circumpolar Arctic vegetation map. (1:7,500,000 scale), conservation of Arctic Flora and Fauna (CAFF) map no. 1*. U.S. Fish and Wildlife Service.
- Chapin, F. S., III. (1991). Integrated responses of plants to stress. *BioScience*, 41(1), 29–36. <https://doi.org/10.2307/1311538>
- Chapin, F. S., III, Bret-Harte, M. S., Hobbie, S. E., & Zhong, H. (1996). Plant functional types as predictors of transient responses of arctic vegetation to global change. *Journal of Vegetation Science*, 7(3), 347–358. <https://doi.org/10.2307/3236278>
- Collins, C. G., Elmendorf, S. C., Hollister, R. D., Henry, G. H. R., Clark, K., Bjorkman, A. D., Myers-Smith, I. H., Prev  , J. S., Ashton, I. W., Assmann, J. J., Alatalo, J. M., Carbognani, M., Chisholm, C., Cooper, E. J., Forrester, C., J  nsd  ttir, I. S., Klanderud, K., Kopp, C. W., Livensperger, C., ... Suding, K. N. (2021). Experimental warming differentially affects vegetative and reproductive phenology of tundra plants. *Nature Communications*, 12(1), 3442. <https://doi.org/10.1038/s41467-021-23841-2>
- Cooper, E. J., Dullinger, S., & Semenchuk, P. (2011). Late snowmelt delays plant development and results in lower reproductive success in the high Arctic. *Plant Science*, 180(1), 157–167. <https://doi.org/10.1016/j.plantsci.2010.09.005>
- Coulson, S. J., Leinaas, H. P., Ims, R. A., & S  vik, G. (2000). Experimental manipulation of the winter surface ice layer: The effects on a high Arctic soil microarthropod community. *Ecography*, 23(3), 299–306. <https://doi.org/10.1111/j.1600-0587.2000.tb00285.x>
- Crawford, R. M. M., Chapman, H. M., & Hodge, H. (1994). Anoxia tolerance in high Arctic vegetation. *Arctic and Alpine Research*, 26(3), 308–312.
- Darrouzet-Nardi, A., Steltzer, H., Sullivan, P. F., Segal, A., Koltz, A. M., Livensperger, C., Schimel, J. P., & Weintraub, M. N. (2019). Limited effects of early snowmelt on plants, decomposers, and soil nutrients in Arctic tundra soils. *Ecology and Evolution*, 9, 1820–1844. <https://doi.org/10.1002/ece3.4870>
- Descals, A., Verger, A., Filella, I., Baldocchi, D., Janssens, I. A., Fu, Y. H., Piao, S., Peaucelle, M., Ciais, P., & Pe  uelas, J. (2020). Soil thawing regulates the spring growth onset in tundra and alpine biomes. *Science of the Total Environment*, 742, 140637. <https://doi.org/10.1016/j.scitotenv.2020.140637>
- Doiron, M., Gauthier, G., & L  vesque, E. (2015). Trophic mismatch and its effects on the growth of young in an Arctic herbivore. *Global Change Biology*, 21, 4364–4376. <https://doi.org/10.1111/gcb.13057>
- Etzelm  ller, B., Guglielmin, M., Hauck, C., Hilbich, C., Hoelzle, M., Isaksen, K., Noetzel, J., Oliva, M., & Ramos, M. (2020). Twenty years of European mountain permafrost dynamics—The PACE legacy. *Environmental Research Letters*, 15(10), 104070. <https://doi.org/10.1088/1748-9326/abae9d>
- Finne, E. A., Bjerke, J. W., Stordal, F., & Tallaksen, L. M. (2025). Lichens are more tolerant against winter warming stress than vascular and non-vascular plants: Insights from an alpine field experiment. *Journal of Ecology*, 113(3), 518–530. <https://doi.org/10.1111/1365-2745.14482>
- Frei, E. R., & Henry, G. H. R. (2022). Long-term effects of snowmelt timing and climate warming on phenology, growth, and reproductive effort of Arctic tundra plant species. *Arctic Science*, 8(3), 700–721. <https://doi.org/10.1139/as-2021-0028>
- Frisvoll, A. A., & Elvebakk, A. (1996). Part 2. Bryophytes. In A. Elvebakk & P. Prestt  r (Eds.), *Svalbard plants, fungi, algae and cyanobacteria* (p. 198). Norsk Polarinstitutt.
- Givnish, T. (2002). Adaptive significance of evergreen vs. deciduous leaves: Solving the triple paradox. *Silva Fennica*, 36(3), 535. <https://doi.org/10.14214/sf.535>
- Gougherty, A. V., & Gougherty, S. W. (2018). Sequence of flower and leaf emergence in deciduous trees is linked to ecological traits, phylogenetics, and climate. *New Phytologist*, 220, 121–131.
- Graham, R. M., Cohen, L., Petty, A. A., Boisvert, L. N., Rinke, A., Hudson, S. R., Nicolaus, M., & Granskog, M. A. (2017). Increasing frequency and duration of Arctic winter warming events. *Geophysical Research Letters*, 44(13), 6974–6983. <https://doi.org/10.1002/2017GL073395>
- Hansen, B. B., Grotan, V., Aanes, R., Saether, B. E., Stien, A., Fuglei, E., Ims, R. A., Yoccoz, N. G., & Pedersen, A. O. (2013). Climate events synchronize the dynamics of a resident vertebrate community in the high Arctic. *Science*, 339(6117), 313–315. <https://doi.org/10.1126/science.1226766>
- Hansen, B. B., Isaksen, K., Benestad, R. E., Kohler, J., Pedersen,   .   ., Loe, L. E., Coulson, S. J., Larsen, J. O., & Varpe,   . (2014). Warmer and wetter winters: Characteristics and implications of an extreme weather event in the high Arctic. *Environmental Research Letters*, 9, 114021. <https://doi.org/10.1088/1748-9326/9/11/114021>
- Hanssen-Bauer, I., F  rland, E. J., Hisdal, H., Mayer, S., Sand  , A. B., & Sorteberg, A. (2019). *Climate in Svalbard 2100—A knowledge base for climate adaptation* (NCCS Report No. 1/2019). [https://klimaservicesenter.no/kss/rapporter/rapporter-og-publikasjoner\\_2](https://klimaservicesenter.no/kss/rapporter/rapporter-og-publikasjoner_2)
- Harris, R. M. B., Beaumont, L. J., Vance, T. R., Tozer, C. R., Remenyi, T. A., Perkins-Kirkpatrick, S. E., Mitchell, P. J., Nicotra, A. B., McGregor, S., Andrew, N. R., Letnic, M., Kearney, M. R., Wernberg, T., Hutley, L. B., Chambers, L. E., Fletcher, M.-S., Keatley, M. R., Woodward, C. A., Williamson, G., ... Bowman, D. M. J. S. (2018). Biological responses to the press and pulse of climate trends and extreme events. *Nature Climate Change*, 8(7), 579–587. <https://doi.org/10.1038/s41558-018-0187-9>
- Henry, G. H. R., Hollister, R. D., Klanderud, K., Bj  rk, R. G., Bjorkman, A. D., Elphinstone, C., J  nsd  ttir, I. S., Molau, U., Petraglia, A., Oberbauer, S. F., Rixen, C., & Wookey, P. A. (2022). The International Tundra Experiment (ITEX): 30 years of research on tundra ecosystems. *Arctic Science*, 8(3), 550–571. <https://doi.org/10.1139/as-2022-0041>
- Henry, G. H. R., & Molau, U. (1997). Tundra plants and climate change: The International Tundra Experiment (ITEX). *Global Change Biology*, 3(S1), 1–9. <https://doi.org/10.1111/j.1365-2486.1997.gcb132.x>
- Hollister, R. D., Elphinstone, C., Henry, G. H. R., Bjorkman, A. D., Klanderud, K., Bj  rk, R. G., Bjorkman, M. P., Bokhorst, S., Carbognani, M., Cooper, E. J., Dorrepaal, E., Elmendorf, S. C., Fetcher, N., Gallois, E. C., Gu  mundsson, J., Healey, N. C., J  nsd  ttir, I. S., Klarenberg, I. J., Oberbauer, S. F., ... Wookey, P. A. (2023). A review of open top chamber (OTC) performance across the ITEX network. *Arctic Science*, 9(2), 331–344. <https://doi.org/10.1139/as-2022-0030>
- Hudson, J. M. G., & Henry, G. H. R. (2010). High Arctic plant community resists 15 years of experimental warming. *Journal of Ecology*, 98, 1035–1041. <https://doi.org/10.1111/j.1365-2745.2010.01690.x>
- IPCC. (2021). Regional fact sheet—Polar regions. In V. Masson-Delmotte, P. Zhai, A. Pirani, S. L. Connors, C. P  an, S. Berger, N. Caud, Y. Chen, L. Goldfarb, M. I. Gomis, M. Huang, K. Leitzell, E. Lonnoy, J. B. R. Matthews, T. K. Maycock, T. Waterfield, O. Yelek  i, R. Yu, & B. Zhou (Eds.), *Climate change 2021: The physical science basis. Contribution of Working Group I to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change*. Cambridge University Press. <https://www.ipcc.ch/report/ar6/wg1/resources/factsheets>
- Isaksen, K., Lutz, J., S  rensen, A. M., God  y,   ., Ferrighi, L., Eastwood, S., & Aaboe, S. (2022). Advances in operational permafrost monitoring on Svalbard and in Norway. *Environmental Research Letters*, 17, 095012. <https://doi.org/10.1088/1748-9326/ac8e1c>
- Isaksen, K., Nordli,   ., Ivanov, B., K  ltzow, M. A.   ., Aaboe, S., Gjelt  n, H. M., Mezghani, A., Eastwood, S., F  rland, E., Benestad, R. E., Hanssen-Bauer, I., Br  kkan, R., Sviashchennikov, P., Demin, V., Revina, A., & Karandasheva, T. (2022). Exceptional warming over the Barents area. *Scientific Reports*, 12, 9371. <https://doi.org/10.1038/s41598-022-13568-5>

- Johnstone, J. F., & Henry, G. H. R. (1997). Retrospective analysis of growth and reproduction in *Cassiope tetragona* and relations to climate in the Canadian high Arctic. *Arctic and Alpine Research*, 29, 459–469.
- Jónsdóttir, I. S. (2011). Diversity of plant life histories in the Arctic. *Preslia*, 83(3), 281–300.
- Jónsdóttir, I. S., Halbritter, A. H., Christiansen, C. T., Althuisen, I. H. J., Haugum, S. V., Henn, J. J., Björnsdóttir, K., Maitner, B. S., Malhi, Y., Michaletz, S. T., Roos, R. E., Klanderud, K., Lee, H., Enquist, B. J., & Vandvik, V. (2023). Intraspecific trait variability is a key feature underlying high Arctic plant community resistance to climate warming. *Ecological Monographs*, 93(1), e1555. <https://doi.org/10.1002/ecm.1555>
- Kausrud, K. L., Mysterud, A., Steen, H., Vik, J. O., Østbye, E., Cazelles, B., Framstad, E., Eikeset, A. M., Mysterud, I., Solhøy, T., & Stenseth, N. C. (2008). Linking climate change to lemming cycles. *Nature*, 456(7218), 93–97. <https://doi.org/10.1038/nature07442>
- Kelsey, K. C., Pedersen, S. H., Leffler, A. J., Sexton, J. O., Feng, M., & Welker, J. M. (2021). Winter snow and spring temperature have differential effects on vegetation phenology and productivity across Arctic plant communities. *Global Change Biology*, 27, 1572–1586. <https://doi.org/10.1111/gcb.15505>
- Langlois, A., Johnson, C.-A., Montpetit, B., Royer, A., Blukacz-Richards, E. A., Neave, E., Dolant, C., Roy, A., Arhonditsis, G., Kim, D.-K., Kaluskar, S., & Brucker, L. (2017). Detection of rain-on-snow (ROS) events and ice layer formation using passive microwave radiometry: A context for Peary caribou habitat in the Canadian Arctic. *Remote Sensing of Environment*, 189, 84–95. <https://doi.org/10.1016/j.rse.2016.11.006>
- Le Moulec, M., Buchwal, A., van der Wal, R., Sandal, L., & Hansen, B. B. (2019). Annual ring growth of a widespread high-arctic shrub reflects past fluctuations in community-level plant biomass. *Journal of Ecology*, 107(1), 436–451. <https://doi.org/10.1111/1365-2745.13036>
- Le Moulec, M., Hendel, A.-H., Petit Bon, M., Jónsdóttir, I. S., Varpe, Ø., van der Wal, R., Beumer, L. T., Layton-Matthews, K., Isaksen, K., & Hansen, B. B. (2025). Data from: Towards rainy high Arctic winters: How experimental icing and summer warming affect tundra plant phenology, productivity and reproduction. *Dryad Digital Repository*, <https://doi.org/10.5061/dryad.9ghx3ffxx>
- Le Moulec, M., Sandal, L., Grøtan, V., Buchwal, A., & Hansen, B. B. (2020). Climate synchronises shrub growth across a high-arctic archipelago: Contrasting implications of summer and winter warming. *Oikos*, 129(7), 1012–1027. <https://doi.org/10.1111/oik.07059>
- Lemaire, M., Bokhorst, S., Witheford, A., Macias-Fauria, M., & Salguero-Gomez, R. (2025). Increases in Arctic extreme climatic events are linked to negative fitness effects on the local biota. *Global Change Biology*, 31, e70157. <https://doi.org/10.1111/gcb.70157>
- Maes, S. L., Dietrich, J., Midolo, G., Schwiager, S., Kumm, M., & Dorrepaal, E. (2024). Environmental drivers of increased ecosystem respiration in a warming tundra. *Nature*, 629, 105–113. <https://doi.org/10.1038/s41586-024-07274-7>
- Milner, J. M., Varpe, Ø., Van der Wal, R., & Hansen, B. B. (2016). Experimental icing affects growth, mortality, and flowering in a high arctic dwarf shrub. *Ecology and Evolution*, 6(7), 2139–2148. <https://doi.org/10.1002/ece3.2023>
- Myers-Smith, I. H., Kerby, J. T., Phoenix, G. K., Bjerke, J. W., Epstein, H. E., Assmann, J. J., John, C., Andreu-Hayles, L., Angers-Blondin, S., Beck, P. S. A., Berner, L. T., Bhatt, U. S., Bjorkman, A. D., Blok, D., Bryn, A., Christiansen, C. T., Cornelissen, J. H. C., Cunliffe, A. M., Elmendorf, S. C., ... Wipf, S. (2020). Complexity revealed in the greening of the Arctic. *Nature Climate Change*, 10(2), 106–117. <https://doi.org/10.1038/s41558-019-0688-1>
- Niittynen, P., Heikkinen, R. K., Aalto, J., Guisan, A., Kempainen, J., & Luoto, M. (2020). Fine-scale tundra vegetation patterns are strongly related to winter thermal conditions. *Nature Climate Change*, 10, 1143–1148. <https://doi.org/10.1038/s41558-020-00916-4>
- Outcalt, S. I., Nelson, F. E., & Hinkel, K. M. (1990). The zero-curtain effect: Heat and mass transfer across an isothermal region in freezing soil. *Water Resources Research*, 26(7), 1509–1516. <https://doi.org/10.1029/WR026i007p01509>
- Pan, C. G., Kirchner, B. P., Kimball, S. J., Kim, Y., & Du, J. (2018). Rain-on-snow events in Alaska, their frequency and distribution from satellite observations. *Environmental Research Letters*, 13(7), 075004.
- Panchen, Z. A., Frei, E. R., & Henry, G. H. R. (2022). Increased Arctic climate extremes constrain expected higher plant reproductive success in a warmer climate. *Arctic Science*, 8(3), 680–699. <https://doi.org/10.1139/as-2020-0045>
- Peeters, B., Pedersen, A. O., Loe, L. E., Isaksen, K., Veiberg, V., Stien, A., Kohler, J., Gallet, J.-C., Aanes, R., & Hansen, B. B. (2019). Spatiotemporal patterns of rain-on-snow and basal ice in high Arctic Svalbard: Detection of a climate-cryosphere regime shift. *Environmental Research Letters*, 14(1), 015002. <https://doi.org/10.1088/1748-9326/aefb3>
- Pérez-Harguindeguy, N., Díaz, S., Garnier, E., Lavorel, S., Poorter, H., Jaureguiberry, P., Bret-Harte, M. S., Cornwell, W. K., Craine, J. M., Gurvich, D. E., Urcelay, C., Veneklaas, E. J., Reich, P. B., Poorter, L., Wright, I. J., Ray, P., Enrico, L., Pausas, J. G., de Vos, A. C., ... Cornelissen, J. H. C. (2013). New handbook for standardised measurement of plant functional traits worldwide. *Australian Journal of Botany*, 61(3), 167–234. <https://doi.org/10.1071/BT12225>
- Petraglia, A., Carbognani, M., & Tomaselli, M. (2013). Effects of nutrient amendments on modular growth, flowering effort and reproduction of snowbed plants. *Plant Ecology and Diversity*, 6(3), 475–486. <https://doi.org/10.1080/17550874.2013.795628>
- Pettorelli, N., Vik, J. O., Mysterud, A., Gaillard, J.-M., Tucker, C. J., & Stenseth, N. C. (2005). Using the satellite-derived NDVI to assess ecological responses to environmental change. *Trends in Ecology & Evolution*, 20(9), 503–510. <https://doi.org/10.1016/j.tree.2005.05.011>
- Phoenix, G. K., & Bjerke, J. W. (2016). Arctic browning: Extreme events and trends reversing arctic greening. *Global Change Biology*, 22(9), 2960–2962. <https://doi.org/10.1111/gcb.13261>
- Phoenix, G. K., Bjerke, J. W., Björk, R. G., Blok, D., Bryn, A., Callaghan, T. V., Christiansen, C. T., Cunliffe, A. M., Davidson, S. J., Epstein, H. E., Lorant, M. M., Martin, A. C., Myers-Smith, I. H., Olofsson, J., Parker, T. C., Parmentier, F.-J. W., Stordal, F., Treharne, R., Tømmervik, H., & Voigt, C. (2025). Browning events in Arctic ecosystems: Diverse causes with common consequences. *PLOS Climate*, 4(1), e0000570. <https://doi.org/10.1371/journal.pclm.0000570>
- Post, E., Forchhammer, M. C., Bret-Harte, M. S., Callaghan, T. V., Christensen, T. R., Elberling, B., Fox, A. D., Gilg, O., Hik, D. S., Høye, T. T., Ims, R. A., Jeppesen, E., Klein, D. R., Madsen, J., McGuire, A. D., Rysgaard, S., Schindler, D. E., Stirling, I., Tamstorf, M. P., ... Aastrup, P. (2009). Ecological dynamics across the Arctic associated with recent climate change. *Science*, 325(5946), 1355–1358. <https://doi.org/10.1126/science.1173113>
- Preece, C., Callaghan, T. V., & Phoenix, G. K. (2012). Impacts of winter icing events on the growth, phenology and physiology of sub-arctic dwarf shrubs. *Physiologia Plantarum*, 146(4), 460–472. <https://doi.org/10.1111/j.1399-3054.2012.01640.x>
- Preece, C., & Phoenix, G. K. (2014). Impact of early and late winter icing events on sub-arctic dwarf shrubs. *Plant Biology*, 16(1), 125–132. <https://doi.org/10.1111/plb.12015>
- Prevé, J. S., Rixen, C., Rüger, N., Høye, T. T., Bjorkman, A. D., Myers-Smith, I. H., Elmendorf, S. C., Ashton, I. W., Cannone, N., Chisholm, C. L., Clark, K., Cooper, E. J., Elberling, B., Fosaa, A. M., Henry, G. H. R., Hollister, R. D., Jónsdóttir, I. S., Klanderud, K., Kopp, C. W., ... Wipf, S. (2019). Warming shortens flowering seasons of tundra plant communities. *Nature Ecology & Evolution*, 3(1), 45–52. <https://doi.org/10.1038/s41559-018-0745-6>



- Putkonen, J., & Roe, G. (2003). Rain-on-snow events impact soil temperatures and affect ungulate survival. *Geophysical Research Letters*, 30(4). <https://doi.org/10.1029/2002GL016326>
- R Core Team. (2020). *R: A language and environment for statistical computing*. R Foundation for Statistical Computing.
- Rasmus, S., Kivinen, S., & Irannezhad, M. (2018). Basal ice formation in snow cover in northern Finland between 1948 and 2016. *Environmental Research Letters*, 13(11), 114009. <https://doi.org/10.1088/1748-9326/aae541>
- Rixen, C., Høye, T. T., Macek, P., Aerts, R., Alatalo, J. M., Anderson, J. T., Arnold, P. A., Barrio, I. C., Bjerke, J. W., Björkman, M. P., Blok, D., Blume-Werry, G., Boike, J., Bokhorst, S., Carbognani, M., Christiansen, C. T., Convey, P., Cooper, E. J., Cornelissen, J. H. C., ... Zong, S. (2022). Winters are changing: Snow effects on Arctic and alpine tundra ecosystems. *Arctic Science*, 8(3), 572–608. <https://doi.org/10.1139/as-2020-0058>
- Semenchuk, P. R., Elberling, B., & Cooper, E. J. (2013). Snow cover and extreme winter warming events control flower abundance of some, but not all species in high arctic Svalbard. *Ecology and Evolution*, 3(8), 2586–2599. <https://doi.org/10.1002/ece3.648>
- Semenchuk, P. R., Gillespie, M. A. K., Rumpf, S. B., Baggesen, N., Elberling, B., & Cooper, E. J. (2016). High Arctic plant phenology is determined by snowmelt patterns but duration of phenological periods is fixed: An example of periodicity. *Environmental Research Letters*, 11(12), 125006. <https://doi.org/10.1088/1748-9326/11/12/125006>
- Serreze, M. C., Gustafson, J., & Barrett, A. P. (2021). Arctic rain on snow events: Bridging observations to understand environmental and livelihood impacts. *Environmental Research Letters*, 16, 105009. <https://doi.org/10.1088/1748-9326/ac269b>
- Skarpe, C., & Van der Wal, R. (2002). Effects of simulated browsing and length of growing season on leaf characteristics and flowering in a deciduous arctic shrub, *Salix polaris*. *Arctic, Antarctic, and Alpine Research*, 34(3), 282–286.
- Sokolov, A. A., Sokolova, N. A., Ims, R. A., Brucker, L., & Ehrlich, D. (2016). Emergent rainy winter warm spells may promote boreal predator expansion into the Arctic. *Arctic*, 69(2), 121–129. <https://doi.org/10.14430/arctic4559>
- Song, C., & Saavedra, S. (2018). Structural stability as a consistent predictor of phenological events. *Proceedings of the Royal Society B: Biological Sciences*, 285, 20180767.
- Thomas, H. J. D., Björkman, A. D., Myers-Smith, I. H., Elmendorf, S. C., Kattge, J., Diaz, S., Vellend, M., Blok, D., Cornelissen, J. H. C., Forbes, B. C., Henry, G. H. R., Hollister, R. D., Normand, S., Prevéy, J. S., Rixen, C., Schaepman-Strub, G., Wilmking, M., Wipf, S., Cornwell, W. K., ... Vries, F. T. (2020). Global plant trait relationships extend to the climatic extremes of the tundra biome. *Nature Communications*, 11, 1351. <https://doi.org/10.1038/s41467-020-15014-4>
- Treharne, R., Bjerke, J. W., Tømmervik, H., & Phoenix, G. K. (2020). Development of new metrics to assess and quantify climatic drivers of extreme event driven Arctic browning. *Remote Sensing of Environment*, 243, 111749. <https://doi.org/10.1016/j.rse.2020.111749>
- Trisos, C. H., Merow, C., & Pigot, A. L. (2020). The projected timing of abrupt ecological disruption from climate change. *Nature*, 580(7804), 496–501. <https://doi.org/10.1038/s41586-020-2189-9>
- Van Beest, F. M., Barry, T., Christensen, T., Heiðmarsson, S., McLennan, D., & Schmidt, N. M. (2022). Extreme event impacts on terrestrial and freshwater biota in the arctic: A synthesis of knowledge and opportunities. *Frontiers in Environmental Science*, 10, 983637. <https://doi.org/10.3389/fenvs.2022.983637>
- Van der Wal, R., & Stien, A. (2014). High-arctic plants like it hot: A long-term investigation of between-year variability in plant biomass. *Ecology*, 95(12), 3414–3427. <https://doi.org/10.1890/14-0533.1>
- Vickers, H., Høgda, K. A., Solbø, S., Karlsen, S., Tømmervik, H., Aanes, R., & Hansen, B. B. (2016). Changes in greening in the high arctic: Insights from a 30 year AVHRR max NDVI dataset for Svalbard. *Environmental Research Letters*, 11, 105004. <https://doi.org/10.1088/1748-9326/11/10/105004>
- Wipf, S., & Rixen, C. (2010). A review of snow manipulation experiments in Arctic and alpine tundra ecosystems. *Polar Research*, 29(1), 95–109. <https://doi.org/10.3402/polar.v29i1.6054>
- Wood, S. N., Goude, Y., & Shaw, S. (2015). Generalized additive models for large data sets. *Journal of the Royal Statistical Society. Series C, Applied Statistics*, 64(1), 139–155. <https://doi.org/10.1111/rssc.12068>
- You, Q., Cai, Z., Pepin, N., Chen, D., Ahrens, B., Jiang, Z., Wu, F., Kang, S., Zhang, R., Wu, T., Wang, P., Li, M., Zuo, Z., Gao, Y., Zhai, P., & Zhang, Y. (2021). Warming amplification over the Arctic Pole and Third Pole: Trends, mechanisms and consequences. *Earth-Science Reviews*, 217, 103625. <https://doi.org/10.1016/j.earscirev.2021.103625>
- Zhang, W., Döscher, R., Koenig, T., Miller, P. A., Jansson, C., Samuelsson, P., Wu, M., & Smith, B. (2020). The interplay of recent vegetation and sea ice dynamics—Results from a regional earth system model over the Arctic. *Geophysical Research Letters*, 47(6), e2019GL085982. <https://doi.org/10.1029/2019GL085982>

## SUPPORTING INFORMATION

Additional supporting information can be found online in the Supporting Information section at the end of this article.

**Table S1:** Summary statistics of summer soil temperature and volumetric water content.

**Table S2:** Sampling dates and number of repeated measurements.

**Table S3:** Summary statistics of NDVI metrics.

**Table S4:** Summary statistics of the abundance of vascular plants.

**Table S5:** Correlation coefficients between maximum NDVI and abundance.

**Table S6:** Summary of leaf trait statistics for *Salix polaris*.

**Table S7:** Summary statistics of the day-of-year a certain phenophase is reached ('timing').

**Table S8:** Summary statistics of time spent in a certain phenophase ('duration').

**Table S9:** Summary statistics of flower counts.

**Figure S1:** Mesic habitat species composition.

**Figure S2:** Immediate icing treatment effect on soil temperature.

**Figure S3:** Estimated NDVI curves across treatments and years.

**Figure S4:** *Salix polaris* phenology.

**How to cite this article:** Le Moulec, M., Hendel, A.-L., Petit Bon, M., Jónsdóttir, I. S., Varpe, Ø., van der Wal, R., Beumer, L. T., Layton-Matthews, K., Isaksen, K., & Hansen, B. B. (2026). Towards rainy high Arctic winters: How experimental icing and summer warming affect tundra plant phenology, productivity and reproduction. *Journal of Ecology*, 114, e70234. <https://doi.org/10.1111/1365-2745.70234>